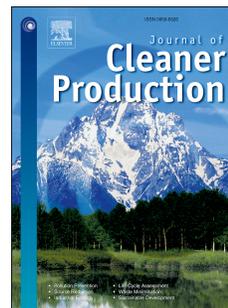


Journal Pre-proof

Investigating on the environmental sustainability of organic animal products? The case of organic eggs

Michele Costantini, Daniela Lovarelli, Luigi Orsi, Andrea Ganzaroli, Valentina Ferrante, Pierluigi Febo, Marcella Guarino, Jacopo Bacenetti



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CRediT author statement

Michele Costantini, Daniela Lovarelli: Conceptualization, Methodology, Software elaboration and Writing- Original draft preparation **Andrea Ganzaroli, Luigi Orsi, Pierluigi Febo:** Writing - Review & Editing, **Jacopo Bacenetti:** Data curation, Writing- Original draft preparation, Writing - Review & Editing, Supervision. **Valentina Ferrante, Marcella Guarino:** Supervision, Writing- Reviewing and Editing

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1 **Investigating on the environmental sustainability of organic animal products? The case of**
2 **organic eggs**

3

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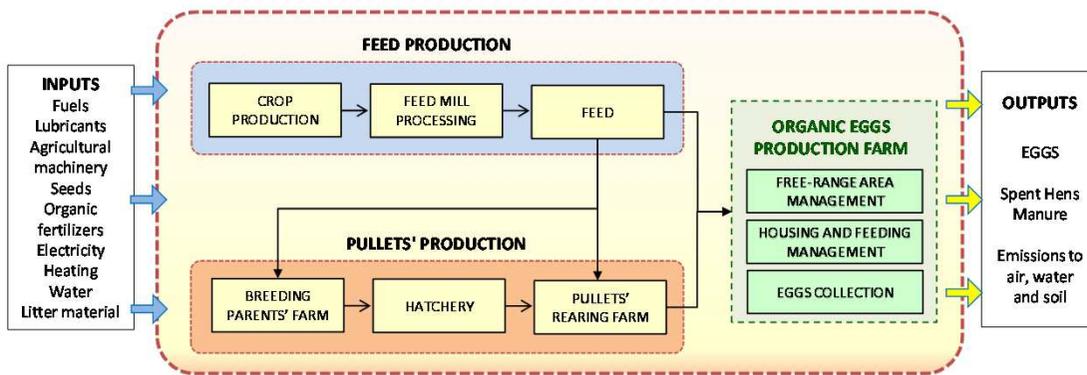
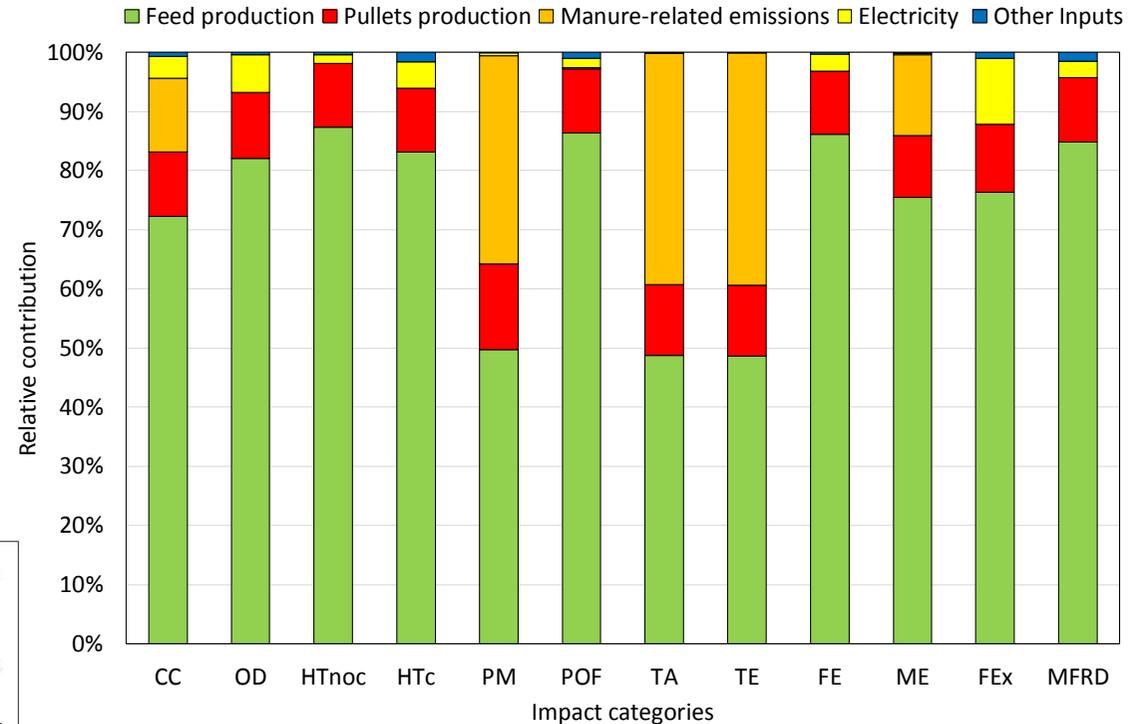
AIM



→ quantify the environmental impact related to the production of eggs according to the organic method in Northern Italy

→ identify some potential mitigation solutions that could be implemented for impact reduction

RESULTS



FUNCTIONAL UNIT

1 kg of fresh shelled eggs produced with organic method

Climate change (CC), Ozone depletion (OD), Particulate matter formation (PM), Human toxicity-no cancer effect (HTnoc), Human toxicity- cancer effect (HTc), Photochemical ozone formation (POF), Terrestrial acidification (TA), Terrestrial eutrophication (TE), Freshwater eutrophication (FE), Marine eutrophication (ME), Freshwater ecotoxicity (FEx), Mineral and fossil resource depletion (MFRD)

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14

15 **Abstract**

16

17 The organic farming of laying hens is experiencing a growing trend in Italy, following an
18 increase in consumer demand for organic eggs. The present study aimed to investigate the
19 environmental performance of organic egg production for the first time in the Italian context. To
20 this end, the Life Cycle Assessment (LCA) of organic egg production in a farm rearing laying hens
21 located in Northern Italy was performed. The analysis was carried out in a cradle to farm gate
22 perspective, with 1 kg of eggs selected as functional unit. Primary data relating to animal
23 performances and resources consumed was collected on site, and subsequently integrated with
24 secondary data, including estimates of manure-related emissions. In order to model in a
25 representative way the organic feed consumed, data relating to typical cropping systems of the
26 country has been used for the various ingredients, keeping the organic production method
27 specifications into account. Inventory data was then converted on an annual basis and
28 characterized using the ILCD method, and twelve impact categories were assessed. Moreover, the

29 influence on impact results of different allocation choices and efficiency in terms of hen-day egg
30 production were explored with a sensitivity analysis.

31 The main environmental burden for organic egg production showed to be feed production and
32 supply, with a share ranging from 49% to 87% over all the evaluated impact categories. Other
33 hotspots are pullets rearing, responsible for a share between 10 and 14% over all categories, and
34 manure-related emissions, which weighed significantly for PM (35%), TA (39%) and TE (39%). A value
35 for CC of 1.56 kg CO₂ eq/kg shelled eggs was obtained, thanks to good production performances
36 together with some benefits given by organic feed use, particularly the avoidance of mineral
37 fertilizer consumption and of land use change related emission. At the same time, the results show
38 clearly that environmental improvements should be sought primarily in the same feed area. This
39 must be done both on-farm, which was highlighted also by the sensitivity analysis on hen-day egg
40 production, and at the supply chain level, acting on the impact related to crop production and
41 pullets rearing phases. Starting from the results, some environmental weaknesses and strengths of
42 organic farming have been discussed. Future studies must further investigate the impact of this
43 rearing system in a wider perspective and explore possible scenarios of mitigation practices.

44

45

46 **Keywords**

47 Life Cycle Assessment, Organic production, Animal Product, Poultry, Environmental impact

48 List of acronyms

- 49 CC – Climate Change
- 50 CP – Crude Protein
- 51 FCR – Feed Conversion Ratio
- 52 FE – Freshwater Eutrophication
- 53 FEx – Freshwater Ecotoxicity
- 54 FU – Functional Unit
- 55 HTc – Human toxicity, cancer effects
- 56 HTnoc – Human toxicity, non-cancer effects
- 57 LCA – Life Cycle Assessment
- 58 LUC – Land Use Change
- 59 LW – Live Weight
- 60 ME – Marine Eutrophication
- 61 MFRD – Mineral, Fossil and renewable Resource Depletion
- 62 OD – Ozone Depletion
- 63 PM – Particulate Matter formation
- 64 POF – Photochemical Oxidant Formation
- 65 TA – Terrestrial Acidification
- 66 TE – Terrestrial Eutrophication

67

68

69 1. Introduction

70

71 The global demand for products of animal origin has been steadily increasing over the past
72 50 years driven by rapid population and income growth and it is expected to grow further in the
73 near future (FAO, 2018). At the same time, animal food products are key drivers of adverse
74 environmental effects such as climate change, water and air pollution, land degradation and
75 biodiversity loss, to name a few (Poore & Nemecek, 2018; Crenna et al., 2019). The poultry industry is
76 strongly linked to these phenomena since today it appears the most developed among the supply

77 chains of animal products and is experiencing continuous growth that shows no signs of stopping.
78 Although the poultry supply chain has been identified by several authors as the most
79 environmentally efficient among those of animal products (Roma et al., 2015), raising concerns
80 about the environmental sustainability of livestock productions have put even this sector under
81 investigation.

82 As regards in particular the chicken egg production, in 2017 it accounted globally for 80.0
83 Mt, the maximum production ever reached. The European Union (EU-28), with 7.1 Mt produced in
84 the same year, contributes significantly to the market for this product (FAOSTAT, 2020). Italy ranks
85 among the most important European egg producers: in 2018, it hosted a population of about 39
86 million laying hens and the national production reached about 12.6 billion eggs, which made the
87 country self-sufficient for eggs demand (ISMEA, 2019).

88 With regard to the rearing systems, in the EU countries, until the early 2000s the majority of
89 egg farms practiced rearing of laying hens in cages, while recently alternative housing methods
90 are spreading, mainly due to a growing public concern for animal welfare. A survey conducted by
91 the European Union, in fact, indicates that an absolute majority of citizens (94%) are of the view it is
92 important to protect the welfare of farmed animals (Eurobarometer, 2016). Moreover, most
93 respondents (59%) indicate a higher willingness-to-pay for improved animal welfare. Referring
94 specifically to laying hens, some recent studies (Leenstra et al., 2014; Zakowska-Biemans & Tekien,
95 2017; Rahmani et al., 2019) have shown that European consumers may consider free-range farming
96 systems positively influencing animal welfare, which explains the increase in consumer demand for
97 eggs deriving from cage-free facilities. Furthermore, the European Union ban on conventional
98 cage rearing (Council Directive, 1999/74/EC), which came into force on January 1, 2012, has
99 forced many farmers to shift towards different housing techniques. Still in accordance with EU
100 Council Directive 1999/74/EC, the currently accepted rearing systems of laying hens are *furnished*
101 cages, more commonly known as enriched, and three forms of *alternative systems* (cage-free),
102 namely barn, free-range and organic.

103 In 2018, Italian egg production was divided as follows: 45% came from barn systems, 42% from
104 enriched cages, 9% from organic and 3% from free-range systems. However, market trends in

105 recent years show that sales of eggs from caged hens are continuously decreasing at the expense
106 of those from alternative systems (ISMEA, 2019).

107 Generally, organic farming claims to lighten the environmental burden of agricultural
108 production while generating ecological benefits. However, organic agriculture is accused of being
109 less productive than traditional agriculture, and this efficiency gap is often identified negatively
110 from an environmental point of view (Clark & Tilman, 2017). As concerns egg production, the
111 organic system generally appears to be less efficient of the other rearing systems, particularly of
112 cages. According to Rööös *et al.* (2018), organic hens usually show a higher feed conversion ratio
113 (FCR) between +2% and +20% compared to hens housed in cages or aviaries, and commonly
114 present even an increased mortality due to injuries and diseases. Considering the important role of
115 feed on the overall impact (Clune *et al.*, 2017), many authors agree that organic eggs perform
116 worse than non-organic in environmental terms, while on the other side there is who claims that the
117 production gap is compensated by low input of resources to the system (Leinonen & Kyriazakys,
118 2016). However, both sides show a degree of variability in literature and there is not one that
119 performs better than the other with consistency, nor for all impact categories in the same way
120 (Leinonen *et al.*, 2012; Pelletier, 2017). For all these reasons, the issue is still hotly debated.

121 Up to now, authors have conducted studies on this topic mainly in order to track a
122 representative egg environmental performance for a regional or a country production, namely for
123 Sweden (Cederberg *et al.*, 2009), Australia (Wiedemann & McGahan, 2010), Netherlands (Dekker
124 *et al.*, 2011), United Kingdom (Leinonen *et al.*, 2012), Iowa (USA) (Pelletier *et al.*, 2013), Alborz (Iran)
125 (Ghasempour & Ahmadi, 2016) and Canada (Pelletier, 2017). Some of these studies additionally
126 aimed to compare different rearing systems and verify their influence on environmental
127 performance. In particular, the organic system was included in the analysis by Dekker *et al.* (2011),
128 Leinonen *et al.* (2012) and Pelletier (2017). Abín *et al.* (2018) instead, investigated the impacts of a
129 single rearing system (i.e. enriched cages) in Spain. Van Hal *et al.* (2019) performed an LCA study
130 on the impact of egg production by exploring an innovative allocation method that rewards low-
131 opportunity-cost-feedstuffs as they avoid feed-food competition. Compared to other livestock
132 sectors, however, studies on the environmental impact of egg production are limited.

133 All the previously quoted studies adopted the environmental Life Cycle Assessment (LCA)
134 approach. LCA, whose normative reference is represented by ISO 14040 (ISO, 2006) and 14044 (ISO,
135 2018), is a method originally developed for industrial processes but recently has also been
136 commonly employed in the agri-food sector (Lovarelli & Bacenetti, 2017a; Bernardi et al., 2018;
137 Tedesco et al., 2019). LCA allows to quantify the potential environmental impact of a product (or
138 process) during its whole life cycle, or a part of it, and to identify hotspots within the examined
139 systems, thus to generate targeted mitigation strategies (Notarnicola et al., 2017). In the Italian
140 context, although several livestock systems have already been analysed with the environmental
141 LCA approach, there is a lack of studies related to egg production. This is the first study focused on
142 the environmental performance of organic egg production in Italy and therefore aims to: quantify
143 its impact in order to provide a national reference; highlight the main contributors to this impact,
144 also referred to as *environmental hotspots*; starting from these, discuss possible mitigation strategies,
145 highlighting any environmental critical issues of the system and rooms for improvement.

146

147 **2. Materials and methods**

148 **2.1. Goal and Scope definition**

149 Based on the gaps identified in the introduction, this LCA study was performed with the
150 following aims:

151 – quantify the environmental impact related to the production of eggs according with the
152 organic method in Northern Italy;

153 identify the processes mainly responsible for this impact and, consequently, discuss some potential
154 mitigation solutions that could be implemented for impact reduction;

155 – strike up a debate on the influence of the rearing system on the environmental impact of
156 egg production, with a keen eye on main parameters that determine differences between
157 them.

158 The analysis is based on data collected from a farm specialized in organic egg production located
159 in the North-Eastern Italian region of Friuli-Venezia Giulia, which, given its characteristics, can be
160 considered representative of this egg production system in the Po valley.

161 The outcomes of this study could be useful for the stakeholders involved in the poultry industry
162 (e.g. feed producers, farmers and processors) to have a better understanding of the environmental
163 performance of the supply chain they work in. At the same time, the results could support actions
164 and/or policies to raise environmental awareness and sustainability by farmers, as well as by
165 individuals, organizations, associations and policymakers.

166

167 **2.2. System description**

168 The analysed farm rears laying hens in an organic farming system, following the indications of
169 Council Regulations (EC) n. 834/2007 and 889/2008 on organic production.

170 The production cycle begins with the purchase of a 3000 pullets flock, raised with organic
171 method in another specialized farm that is located in the neighboring region. This flock corresponds
172 to the maximum number of animals that can be raised simultaneously in a poultry house for laying
173 hens, according to the aforementioned European regulations on organic production. Pullets are
174 imported at 16 weeks of age and a live weight (LW) of 1.55 kg, ready to start the laying cycle. In
175 fact, the analysed farm manages only the egg-laying phase, which lasts up to 72 weeks. During this
176 period, animals spend part of the time inside the barn, and they also have free daily access to a
177 free-range area, normally allowed by the farmer during daylight hours. The barn has an area of 500
178 m², which is partly covered with sand as litter material, and is also equipped with a raised tier with a
179 slatted floor, on which automatic feeders, perches and special nests are placed. Eggs are
180 collected from nests with automatic belts. The free-range is 1.25 ha and is cultivated with alfalfa,
181 which is renewed every 3 years. In detail, the mechanized operations carried out are plowing and
182 harrowing, seed bed preparation, and subsequently sowing and rolling. No other intervention is
183 carried out during the 3 years. Since the farm does not own any land other than that destined for
184 the free-range area, the entire feed supply is purchased externally by a specialized feed mill
185 authorized in organic feed production. During the laying cycle, two different feeds are provided:
186 the first with a crude protein (CP) content of 18.5% for a period of 24 weeks, and the second with
187 17% CP until the end of the cycle, for a total of 48 weeks. These two commercial feeds are
188 produced locally, as required by the organic method legislation, and are mainly composed by
189 maize and wheat grains, wheat bran and soybean meal.

190 At the end of the productive cycle, at a LW of 1.95 kg, spent hens are sent to the
191 slaughterhouse. Then, following the all-in/all-out principle, a depopulation period of 21 days is
192 carried out before a new batch of pullets is brought in. In this time frame, the barn is completely
193 cleansed with steam and the whole amount of manure (40 tons) produced during the cycle is
194 collected. The manure is entirely exported to a nearby cereal farm producing maize grain, which
195 forms part of the maize used as feed ingredient.

196

197 **2.3. Functional unit and system boundary**

198 The Functional Unit (FU) is the unit to which refer all the inventory data and the environmental
199 results. In this study, 1 kg of fresh shelled eggs produced with organic method was selected as FU, in
200 accordance with Livestock Environmental Assessment and Performance Partnership (LEAP)
201 guidelines (LEAP, 2016).

202 The study was carried out in a *cradle-to-farm gate* perspective, thus including in the system
203 boundary the eggs life cycle until the end of the agricultural phase (**Figure 1**). Therefore, all inputs
204 related to the rearing cycle, such as the production, purchase and on-farm transportation of feed
205 and pullets, as well as the supply of other inputs such as electricity, water and litter material, were
206 considered. On the contrary, packaging, from-farm transport, possible processing, market
207 distribution and consumption phases, and all related waste disposal were excluded.

208

209 **Figure 1 around here**

210

211 As crop intended for feed production is concerned, this assessment considered raw material
212 extraction (e.g., fossil fuels and minerals); manufacture (seeds, fertilizers and agricultural machines);
213 use (diesel fuel consumption and engines exhaust gas emissions, tire abrasion emissions and fertilizer
214 related emissions); maintenance and final disposal of agricultural machines.

215 Emissions due to feces deposition during animal housing have been included in the
216 assessment, while the impact related to manure export and its field spreading was not, since
217 considered as falling outside the system boundaries.

218 Due to the lack of information, similarly to previous LCA studies on poultry productions (Roma
219 et al., 2015), impacts related to drug use were excluded, and the same was done for production
220 and maintenance of farm's facilities and devices.

221 As concerns agricultural land directly and indirectly involved in the analysis, no changes in the
222 soil organic carbon were taken into account because considered to be in a steady state. In fact,
223 according to the legislation on organic agricultural products, all feed ingredients are produced
224 locally, thus on agricultural land of the Po Valley destined for arable crops where no land use
225 change (LUC) has occurred recently. This assumption agrees with previous studies on the soil
226 carbon stock topic (Huang et al., 2012; Han et al., 2016) and with previous LCA analyses focused
227 on agricultural production in Northern Italy (Noya et al., 2018; Bacenetti, 2019).

228

229 **2.4. Inventory data collection**

230 Primary data refer to a farm located in North-Eastern Italy, whose production system is
231 described in detail in section 2.2, and were gathered through a survey at the farm and a
232 questionnaire completed by the farmer.

233 **Table 1** reports the main data about the animals' productive parameters, as well as quantity
234 and type of inputs consumed and outputs generated, relating to a production cycle that took
235 place between 2018 and 2019.

236

237 **Table 1 – around here**

238

239 The inventory data for the three-year renewal of alfalfa cultivation in the free-range area were
240 obtained directly from the farmer, and subsequently integrated with information from *Bacenetti et*
241 *al.* (2018), who explored the environmental impact of alfalfa cultivation in the Northern Italian
242 context.

243 Feed composition (**Table 2**) was modeled starting from commercial ingredients lists and
244 main nutritional values, provided by the farmer. Ingredients below 1% w/w (i.e. feed additives) were
245 excluded from the study. In calculating the farm feed consumption, an additional 2.5% was
246 considered compared to the average animals' intake, due to wastes declared by the farmer that

247 occur during animal feeding. This share of spilled feed falls to the ground and mixes with litter
248 material and animal excreta, contributing to manure formation.

249 As regards the inventory data relating to crop production of feeds, detailed information on
250 field operations and related inputs and outputs were recovered by *Bacenetti et al. (2016)*,
251 *Bacenetti et al. (2017a)*, *Dal Ferro et al., (2017)* as regards maize, wheat and soybean; *Bacenetti et*
252 *al. (2017b)* as for flax and *Forleo et al. (2018)* for sunflower. These studies reflect peculiarities of the
253 Italian agricultural context for the different feed ingredients production. However, to ensure the
254 consistency with the system under study, i.e. to comply with organic production specifications,
255 mineral fertilizers inputs and relative distribution operations were not included in the inventory. The
256 same was done with pesticides, which were replaced with mechanical weeding interventions, for
257 which data were taken from *Lovarelli & Bacenetti (2017b)* and *Lovarelli et al. (2017)*. Finally, to take
258 into account the significant drop in yields normally observed in organic farming (*de Ponti et al.,*
259 *2012*), a yield ratio of 70% was assumed for all the cultures with respect to the conventional farming
260 references, similarly to the average value observed by *Dal Ferro et al. (2017)* in Northern Italy.

261

262 **Table 2 – around here**

263

264 Background data relating to feed processing, production of pullets, electricity, water, litter and
265 on-farm transports were obtained from the established life cycle inventories data sources Ecoinvent
266 V3.5 (*Weidema et al., 2013*, *Moreno Ruiz et al., 2018*) and Agrifootprint (*Blonk Consultants, 2014*). All
267 electricity consumptions, both direct on the farm and those related to upstream processes, were
268 modeled using the medium voltage Italian energy mix at grid. Impact due to the on-farm transport
269 of pullets, litter material and feed inputs was computed through the ton-per-kilometer method,
270 knowing the distance from the suppliers.

271 Emissions related to animal housing and manure management were estimated by following
272 the IPCC guidelines (*IPCC, 2006; 2019*); in particular, Tier 2 was used to quantify methane (CH_4)
273 emissions and Tier 1 for nitrous oxide (N_2O) and ammonia (NH_3). Methane emissions from enteric
274 fermentations were assumed to be zero, in consistence with previous LCA studies on poultry

275 systems, both when intended for the production of eggs (Leinonen et al., 2012; Pelletier, 2017) and
276 broiler meat (Cesari et al., 2017; Da Silva et al., 2019).

277 In order to estimate CH₄ emissions from manure, volatile solids excretion was predicted using
278 the equation n. 5 provided by LEAP (2016). Due to the lack of specific data, the diet digestibility
279 was considered to be fixed at 80% and manure ash content at 10% (LEAP, 2016). Subsequently, a
280 methane producing capacity factor (B₀) of 0.39 m³ CH₄/kg of volatile solids and a methane
281 conversion factor of 1.5% were used for computing manure CH₄ emissions (IPCC, 2006; 2019).

282 The excreted nitrogen, required to estimate nitrogen compounds emissions, was calculated by
283 deducting nitrogen retention from nitrogen intake. The latter was derived from the dry matter
284 intake of the animals and the diet CP content, both provided by the farmer. The retention instead
285 was predicted using the equation n. 10.33D provided by IPCC (2006; 2019), thus considering egg
286 mass production, average nitrogen content in eggs (1.85% on mass), daily weight gain and
287 average nitrogen content in LW (2.8% on mass).

288 Subsequently, nitrous oxide and ammonia emission were calculated using the Tier 1 conversion
289 factors of IPCC (2006; 2019). For both nitrogen compounds, specific conversion factors were used
290 for the interior of the barn and the free-range area. In this regard, considering the average time
291 spent by the hens in the different areas of the farm, it has been assumed that the animals deposit
292 70% of the feces inside and 30% outside. For the share deposited externally, nitrate (NO₃⁻) leaching
293 was calculated according with *Brentrup et al.* (2000) considering as if it were spread manure.

294 As in the majority of LCA studies focused on livestock systems (Roma et al., 2015), CO₂ emitted
295 by the animals was not accounted because considered balanced by that fixed by means of
296 photosynthesis by the plants involved in the system, i.e. crops (IPCC, 2006).

297 Further information on crop production elaboration (Tables S1, S2, S3, S4, S5) and emissions
298 estimation (Table S6) can be found in the supplementary materials.

299

300 **2.5. Allocation**

301 When multifunctional processes are studied, allocation should be introduced in order to assign
302 impact shares to the different co-products. As the function of the system under analysis in this study
303 is to supply eggs, the environmental impacts were not allocated but attributed entirely to eggs, by

304 far the most relevant product of the system both in economic, mass and nutritional value terms.
305 However, to check if this methodological choice had a significant effect on the results, a sensitivity
306 analysis was carried out which explored the influence of other possible allocation methods, as
307 suggested in LEAP guidelines (2016). The setting of this analysis is discussed in section 2.7.

308

309 **2.6. Impact assessment**

310 In the life cycle impact assessment phase, emissions and resource data identified during the
311 inventory are translated into indicators that reflect environment pressures as well as resource
312 scarcity. The impact assessment was performed using the ILCD (International Reference Life Cycle
313 Data System) 2011 Midpoint + method (V1.03), which is the impact assessment method endorsed
314 by the European Commission, and processed with the SimaPro® 8.0.5 software. The evaluated
315 impact categories are the following: climate change (CC), ozone depletion (OD), particulate
316 matter formation (PM), photochemical oxidant formation (POF), terrestrial acidification (TA),
317 terrestrial eutrophication (TE), freshwater eutrophication (FE), marine eutrophication (ME),
318 freshwater ecotoxicity (FEx), mineral and fossil resources depletion (MFRD).

319

320 **2.7 Sensitivity analysis**

321 A sensitivity analysis was carried out with the variation of some methodological choices and key
322 parameters to investigate their effect on the environmental impact. The following aspects were
323 taken in consideration:

324 – Allocation method. Results variation was evaluated when considering the meat of spent
325 hens as a relevant co-product. In this regard, LEAP guidelines (2016) suggest that it would
326 be preferable to handle this co-production using biophysical allocation. This was made by
327 portioning metabolizable energy requirements between growth and egg production. As a
328 reference for these requirements, 5.5 kcal/g of LW gain (starting from the hatchling weight)
329 and 2.07 kcal/g of egg were considered (NRC, 1994).
330 In addition, economic allocation was evaluated, taking reference prices from the
331 wholesale poultry exchange of Forlì, one of the most important for the sector in Northern
332 Italy. The prices used were € 14.68/100 units for organic eggs and € 0.27/kg of LW for

333 medium-sized spent hens (averages for January 2020, ISMEA, 2020). It should be noted that
334 the value of spent hens' meat is much lower than that of broiler meat. Moreover, the
335 rearing system from which spent hens come does not influence the meat price, as it is
336 however considered low quality poultry meat.

337 – Hen-day egg production. Indeed, it is a key production parameter, directly connected to
338 the FCR, and which may possibly undergo variations due to environmental (temperature,
339 humidity, etc.) and health (dysmetabolic diseases, leg injuries, etc.) conditions, and possibly
340 due to the presence of other stressors, such as sporadic attacks by predators in the free-
341 range area. The environmental performance was therefore explored for hen-day egg
342 production of 85% (LOW scenario) and 93% (HIGH scenario). As a consequence to daily
343 egg-laying changes, variations in animals' nitrogen retention and excretion rates were
344 taken into account, thus estimates of manure-related emissions were adapted.

345 When cross-referencing both aspects, changes due to different total productions on the total value
346 of eggs produced and energy requirements were also taken into consideration.

347

348 **3. Results**

349 **3.1. Environmental results for organic egg production**

350 The absolute results for 1 kg of organic eggs are reported in **Table 3**, while **Figure 2** shows the
351 contribution analysis.

352

353 **Table 3 – Around here**

354

355 In more detail, for each evaluated impact category, **Figure 2** shows the relative contribution
356 of the different inputs. Each column of the graph can be read as a pie chart, facilitating the
357 identification of environmental hotspots. The input and output flows considered to make the
358 analysis run are grouped for the results presentation into five classes: feed production (referring to
359 consumption that takes place during the egg-laying cycle), pullets production, manure-related
360 emissions, electricity (referring to on-farm consumption for lighting, automatic feeder and egg
361 collection belts) and other inputs (referring to the sum of impacts due to on-farm transport of

362 pullets, feed and litter; litter material production; water supply; free-range area management
363 operations).

364

365 **Figure 2 around here**

366

367 The impact of each class is discussed below:

- 368 - Feed production is the main environmental hotspot, being the major driver of the impact for
369 all the evaluated categories with a share ranging from 49% for TA and TE to 87% for HTnoc.
370 This impact is related to energy requirements and machinery use for agricultural production
371 and feed mill processing, as well as to emissions occurring during crop cultivation.
372 Breaking the feed impact down between the two different feeds used, it emerges that the
373 feed with 17% CP impacts for approximately two thirds (from 64% to 68% over all the
374 evaluated impact categories) of the total, while the feed with 18.4% CP for the remainder.
375 This is due to the fact that they have a similar composition but the first is used for a period of
376 twice the duration of the other within the laying cycle therefore about twice the mass is
377 consumed.
- 378 - Manure-related emissions affect significantly CC (12%), PM (35%), TA (39%), TE (39%) and ME
379 (14%). Among these, the manure-related impact is entirely due to NH₃ volatilization for PM,
380 TA and TE. NH₃ is also responsible for 83% of the manure-related share for ME that, even if to
381 a lesser extent (17%), is also influenced by NO₃⁻ leaching in the free-range area. Regarding
382 the CC, the contribution of manure-related emissions is mainly due to N₂O, which represents
383 77% of the manure-related greenhouse gases emissions, while CH₄ contributed the
384 remaining 23%.
- 385 - Pullets production is responsible for a share ranging from 10% to 14% for all the evaluated
386 impact categories. This impact, in turn, is due especially to the feed consumption necessary
387 for feeding the growing pullets in the specialized farm (Figure S1). Other contributing factors
388 are emissions related to manure and housing management, electricity and heat
389 consumptions and the supply of one-day-old chickens to the pullets farm.

- 390 - Electricity consumption on-farm has a limited contribution on the total impact for all the
391 evaluated impact categories, resulting in less than 4% for all categories except for OD and
392 FEx, for which it makes up respectively 6% and 11%.
- 393 - Other inputs were grouped due to their small overall impact. Indeed, their share is below 2%
394 for all the evaluated impact categories.

395 Given the growing global concern on the climate change issue and its strong connection with the
396 livestock sector (McClelland et al., 2018), specific attention is hereby paid to this aspect. Feed
397 production affects the total CC impact (1.56 kg CO₂ eq/kg eggs) for 72%, representing by far the
398 major contributor, followed by manure-related emissions (12%), pullets production (11%), electricity
399 (4%) and other inputs (1%). **Figure 3** shows the results of a contribution analysis conducted on
400 greenhouse gases.

401

402 **Figure 3 – around here**

403

404 The overall CC impact, expressed in kg CO₂ eq., is caused mainly by CO₂ emissions (58%),
405 followed by N₂O (34%) and, lastly, CH₄ (8%) (Table S8). The effect of other substances is negligible
406 (about 0.1%), thus not reported in the graph. Feed production is the main emitter for all the three
407 gases (Table S9): huge amounts of CO₂ are emitted during mechanized field operations and feed
408 processing, while N₂O and CH₄ are emitted from organic fertilizers storage, handling and spreading
409 on fields. Pullets production is the second contributor for CO₂; manure-related emissions, on the
410 other hand, represent a significant share of N₂O and CH₄.

411

412 **3.2. Sensitivity analysis results**

413 Results of the sensitivity analysis are reported in **Table 4**.

414 As regards the allocation choice, the results show that if the biophysical allocation is adopted,
415 the environmental impacts associated with egg production significantly decrease. This result,
416 despite being the method recommended primarily by ISO standards and LEAP guidelines (2016)
417 when allocating this type of product, raises doubts regarding the distribution of impacts among the
418 co-products of the system, considering that the primary function of the examined system is to

419 supply eggs, while spent hens' meat is a low-value and low-quantity co-product. Indeed, in
420 previous LCA studies on egg production no author has applied this allocation method.

421 The economic allocation, instead, shows to be almost irrelevant, resulting in a reduction of the
422 impact always <1% compared to the baseline scenario with no allocation, and regardless of the
423 hen-day egg production considered. This is due to the abovementioned large gap existing
424 between the two co-products both in terms of mass produced and price, even assuming different
425 hens' productivity. However, it should be underlined that this result could undergo fluctuations
426 depending on the prices variability.

427

428 **Table 4 – around here**

429

430 As regards the influence of the hen-day egg production, the variation is uniform among
431 categories within the two alternative scenarios when considering same allocation methods, with
432 the categories influenced by manure-related emissions that undergo slightly more marked
433 variations than the others. More specifically, the impact categories highly influenced by NH₃
434 emissions (PM, TE and ME) are those most affected both in terms of impact reduction in the HIGH
435 scenario and of impact increase in the LOW scenario. In general, these results highlight the
436 influence of system efficiency on environmental performance while keeping unaltered the system
437 inputs. When considering the absolute values, the impact ratio between the HIGH and LOW
438 scenarios vary overall in a range of 87.1-92.1%.

439

440 **3.3. Uncertainty analysis results**

441 An uncertainty analysis was carried out with the Monte Carlo technique (5,000 iterations and a
442 confidence interval of 95%) to test the robustness of the achieved results. The results, reported in
443 **Figure 4**, show that the uncertainty due to selection of the data source, model imprecision and
444 variability of background data does not significantly affect the results for CC, PM, POF, TA, TE, FE
445 (coefficient of variation < 3%) and ME (coefficient of variation < 8%) while it has great influence on
446 OD, the toxicity related impact categories and MFRD.

447

448 **Figure 4 – around here**

449

450

451 **4. Discussion**

452 The results clearly revealed feed production as the main hotspot for any impact category
453 for the system under analysis. The primary role of feed in determining impacts and its relative
454 contribution appears in line with previous LCA studies on egg production (Dekker et al., 2011,
455 Leinonen et al., 2012, Pelletier, 2017, Abín et al., 2018), and without particular differences
456 attributable to geographic location or adopted rearing system. Also, pullets production and
457 emissions related to manure contributed significantly. As concerns the substances contribution
458 analysis of CC, the results obtained in this study are in accordance with *Dekker et al. (2011)* and
459 *Wiedemann and McGahan (2010)*, while other authors (e.g. Abín et al., 2018) identified N₂O
460 instead of CO₂ as the main greenhouse gas emitted in terms of CO₂ eq from the egg production
461 system.

462 The absolute value for CC observed in this study for the selected FU (1.56 kg of CO₂ eq/kg
463 eggs) falls within the lower quartile of the values range identified in a recent literature review (1.3-
464 6.0 kg of CO₂ eq/kg eggs, Clune et al., 2017). With the awareness that some of the variability
465 observed between different LCA studies is due to methodological choices, the result still shows the
466 analysed farm to have a good environmental performance in terms of carbon footprint. The farm's
467 productivity and the avoidance of some inputs that normally weigh heavily from an environmental
468 point of view in conventional systems, such as mineral fertilizers and imported feed ingredients,
469 have contributed to achieving this result. An important role in determining the impact was certainly
470 played by the replacement of South American soybeans with Italian-produced one, thus excluding
471 impacts related to long distances transportation and LUC at the expense of forest areas. The result
472 also confirms that eggs, represent together with milk the most sustainable product of animal origin
473 with regard to CC (De Vries & De Boer, 2010), beyond the production system adopted.

474 For the other impact categories it is difficult to make comparisons because different
475 assumptions were made on the model used for emissions estimation and/or on the characterization
476 factors used (De Vries & De Boer, 2010), as well as because of the use of a different functional unit

477 or system boundary. At the same time, it is important to consider a full set of impact categories to
478 avoid impact shifting among them as well as a misrepresentation of trade-offs between impacts.
479 This calls for comprehensive studies that cover a large spectrum of environmental impacts to avoid
480 the previously mentioned burden-shifting in order to assess the environmental trade-offs between
481 conventional and organic egg production systems.

482 The results show that mitigation interventions on feed production would be the most incisive
483 option to reduce the overall impact, leading to an environmental improvement for all impact
484 categories. What can be done at the farm level in this regard is to seek continuous improvement in
485 FCR in order to lower feed consumption. The goal must be to optimize this consumption without
486 sacrificing productivity. However, the environmental mitigation on feed production must also take
487 place beyond the farm boundaries: it is essential that sustainable production starts from crops,
488 adopting strategies to obtain an increase in yields and/or a reduction in resource consumption,
489 thus amortizing the impact per kg of consumed feed. This mitigation strategy is limited by the fact
490 that the farmer cannot intervene directly in this field, since the analysed farm depends entirely on
491 the external purchase of feed. What would therefore be needed is a joint action between
492 stakeholders of the supply chain, which should be undertaken at a collective level from farmers
493 since the scarce availability of land for self-production of feed is a common feature to most of the
494 poultry holdings in Italy (Eurostat, 2020). The same should be done for the mitigation of the impact
495 due to pullets rearing.

496 Concerning the impact categories strongly affected by manure-related emissions, namely
497 PM, TA and TE, a considerable mitigation could be obtained by deploying abatement
498 technologies in farm facilities, such as ventilation systems that allow rapid drying of the excreta
499 after deposition or air scrubbers, which have been proven to be a good solution in particular for
500 ammonia (Van der Heyden et al., 2015). However, this latter implementation is unlikely to be
501 applied in organic farms' barns since they normally use natural ventilation systems. A careful
502 fulfilment of protein requirements has also been identified as a potential mitigation strategy (Vesela
503 et al., 2016), so that it can simultaneously avoid protein overdoses in the ration, reducing inputs of
504 protein ingredients, and limiting the nitrogen excretion of animals. A weakness of organic farming in

505 this sense is that restrictions exist on the use of some ingredients, such as synthetic amino acids,
506 making it more difficult to accurately meet nutritional requirements of the animals.

507 Another mitigation opportunity could be the partial replacement of traditional protein
508 sources with alternative ones (insects and worms). Recently, many studies have explored this topic
509 in the poultry sector, highlighting promising results both in production and environmental
510 sustainability terms (Khan, 2018). Further research efforts are needed to better investigate the
511 impacts that alternative protein meals can have if implemented regularly and in large quantities in
512 laying hens' diets.

513 Finally, even if manure is collected only once at the end of the cycle, its energetic
514 valorisation through anaerobic digestion and biogas production would be a viable option for
515 further mitigation. By providing renewable energy, biogas production from manure could
516 contribute to the "eco-functional intensification" of organic farming (Siegmer et al., 2015) and also
517 foster the circular economy concept in agriculture (Toop et al., 2017). Several LCA studies (Uusitalo
518 et al., 2014; Bacenetti & Fiala, 2015; Ingrao et al., 2019) have shown how a biogas plant fed with
519 animal manure can reach a negative value (i.e., environmental benefits) for the Climate Change
520 impact category when are considered the credits for the avoided traditional storage of manure.
521 Anaerobic digestion of poultry manure would also be positive from an agronomic point of view
522 since the anaerobic digestion increases the proportion of ammoniacal nitrogen on total nitrogen in
523 digestate (Röös et al., 2018) and, consequently, its effectiveness in the substitution of mineral
524 fertilizers. Moreover, poultry manure, if digested together with other biomasses (to avoid/reduce the
525 risk of high ammonia concentration into the digesters) has an excellent energy potential both
526 because it normally has a higher volatile solids content than other livestock effluents and it presents
527 also a higher CH₄ yield per kg of volatile solids (Garcia et al., 2019).

528 Due to the lack of information, this study did not include the effect of the system on land
529 use, water use, biodiversity and carbon sequestration, therefore the discussion may present
530 restrictions. Future analyses should deepen the relationships between organic laying hens farming
531 and these topics in a wider environmental perspective. Also, a challenge for future studies is to
532 encompass the environmental dimension of sustainability, creating a connection with animal
533 welfare aspects: several studies have investigated and compared the influence that different

534 rearing systems, feeding management and housing conditions can have on animal welfare (EFSA,
535 2005; Nicol et al., 2017). Nevertheless, these aspects are not taken into consideration when carrying
536 out an LCA analysis. Recently, a preliminary approach to include animal welfare in a social-LCA
537 model for chicken production was proposed (Tallentire et al., 2019). Next step could be to combine
538 this approach together with the LCA in a broader life cycle framework.

539

540 **5. Conclusions**

541 The environmental impact of organic egg production was evaluated by the LCA approach.
542 The results show that feed production is the major impact contributor for all the evaluated
543 categories. Consequently, a greater effort in reducing the environmental burden of egg
544 production should be made first of all in the optimization of feed consumption on farm and in the
545 crop production phase.

546 The outcomes of the sensitivity analysis point out the importance of allocation choice in
547 defining the environmental load of eggs when considering co-products of the system, as it
548 happens for other products of animal origin, and also highlight that the production efficiency of the
549 rearing system plays an important role in determining positive environmental performance.

550 Each rearing system (conventional or organic) has peculiar positive and negative
551 characteristics from an environmental point of view, and some variability can be observed across
552 and within the different systems. Despite some criticisms that organic farming receives in literature,
553 the farm analysed in this study has a good productivity, which was decisive for determining the
554 environmental impact. The impact largely depends on farmers' management choices, rather than
555 on the adopted rearing system and, a priori, there is not a best farming system for all
556 circumstances. In any case, opportunities for further mitigation exist at the farm level in terms of
557 feeding strategies, manure management and measures related to the housing system. Efforts at
558 the supply chain level are instead needed to mitigate the impact of feed production and supply.

559 As this represents, to the authors' knowledge, the first LCA analysis on eggs in Italy, a starting
560 point for future studies is to expand the sample to other farms and explore with comparative LCA
561 different rearing systems, in order to directly compare the performance of organic and
562 conventional production in the same context. Even some alternative scenarios with the mitigation

563 solutions proposed in this study could be explored, evaluating their feasibility and social (animal
564 welfare in particular), environmental and economic performance.

565

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Table 1 – Productive parameters of the analysed farm

Purchased pullets	n. · cycle ⁻¹	3000
Laying cycle duration	weeks	72
Production cycles	n. · year ⁻¹	0.64
Live weight, at pullets purchase	kg	1.55
Live weight, end of the cycle	kg	1.95
Feed intake, as fed	g · day ⁻¹ · head ⁻¹	130
Egg weight	g	58
Hen-day egg production	%	90
Feed Conversion Ratio (FCR)	kg feed intake (as dry matter) · kg eggs produced ⁻¹	2.19
Mortality	% · cycle ⁻¹	4
Broken eggs	%	1.5
Indoor stocking density	n. hens/m ²	6
Electricity consumption	kWh · year ⁻¹	7200
Water consumption	l · day ⁻¹	700
Sand, as litter material	kg · cycle ⁻¹	3200

Table 2 – Feed ingredients composition, expressed in dag · kg⁻¹, as fed. Both feeds have 88 % dry matter on fresh mass.

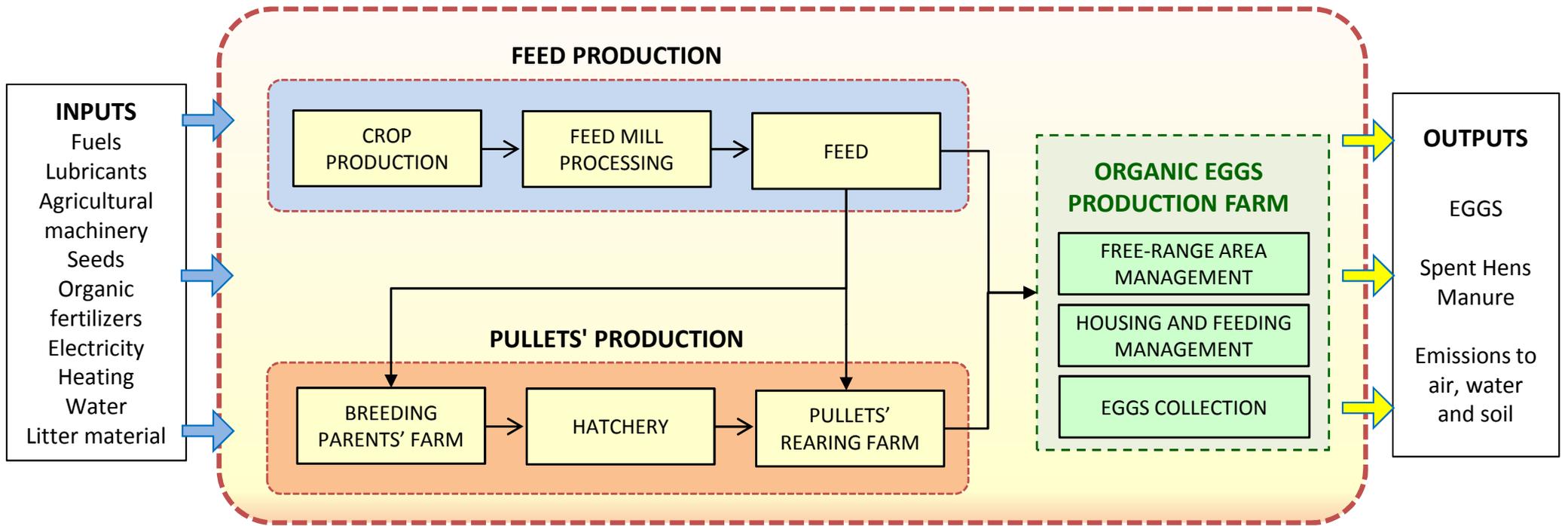
Feed component	Feed 1 st phase	Feed 2 nd phase
	(18.4 % CP on dry matter basis)	(17.0 % CP on dry matter basis)
Maize grain	50	56
Soybean meal, extracted	15	12
Wheat bran	10	8
Wheat grain	10	8
Sunflower meal	8	8
Calcium carbonate	4	5
Maize germ	2	2
Linseed meal	1	1

Table 3 - Absolute results for the selected FU (1 kg of organic shelled eggs)

Impact category	Unit of measure	Score
CC	kg CO ₂ eq	1.56
OD	mg CFC-11 eq	0.10 ⁴
HTnoc	CTUh	5.13 · 10 ⁻⁷
HTc	CTUh	4.02 · 10 ⁻⁸
PM	g PM2.5 eq	4.12
POF	g NMVOC eq	7.63
TA	molc H ⁺ eq	0.17
TE	molc N eq	0.75
FE	g P eq	0.55
ME	g N eq	17.8
FEx	CTUe	3.17
MFRD	g Sb eq	0.03

Table 4 - Results of the sensitivity analysis: impact variations with respect to the baseline scenario (BS) (i.e. no allocation and 90% hen-day egg production).

Impact category	Allocation method	Hen-day egg production		
		93% (HIGH)	90% (BS)	85% (LOW)
CC	No allocation (BS)	-3.3 %	—	6.1 %
	Biophysical	-18.9 %	-16.0 %	-10.7 %
	Economic	-4.1 %	-0.8 %	5.3 %
OD	No allocation (BS)	-3.2 %	—	5.9 %
	Biophysical	-18.8 %	-16.0 %	-10.9 %
	Economic	-4.0 %	-0.8 %	5.0 %
HTnoc	No allocation (BS)	-3.2 %	—	5.9 %
	Biophysical	-18.8 %	-16.0 %	-10.9 %
	Economic	-4.0 %	-0.8 %	5.0 %
HTc	No allocation (BS)	-3.2 %	—	5.9 %
	Biophysical	-18.8 %	-16.0 %	-10.9 %
	Economic	-4.0 %	-0.8 %	5.0 %
PM	No allocation (BS)	-3.6 %	—	6.6 %
	Biophysical	-19.2 %	-16.0 %	-10.2 %
	Economic	-4.4 %	-0.8 %	5.8 %
POF	No allocation (BS)	-3.2 %	—	5.9 %
	Biophysical	-18.8 %	-16.0 %	-10.9 %
	Economic	-4.0 %	-0.8 %	5.0 %
TA	No allocation (BS)	-3.7 %	—	6.7 %
	Biophysical	-19.3 %	-16.0 %	-10.1 %
	Economic	-4.5 %	-0.8 %	5.9 %
TE	No allocation (BS)	-3.7 %	—	6.7 %
	Biophysical	-19.3 %	-16.0 %	-10.1 %
	Economic	-4.5 %	-0.8 %	5.9 %
FE	No allocation (BS)	-3.2 %	—	5.9 %
	Biophysical	-18.8 %	-16.0 %	-10.9 %
	Economic	-4.0 %	-0.8 %	5.0 %
ME	No allocation (BS)	-3.3 %	—	6.0 %
	Biophysical	-18.9 %	-16.0 %	-10.8 %
	Economic	-4.1 %	-0.8 %	5.2 %
FEx	No allocation (BS)	-3.2 %	—	5.9 %
	Biophysical	-18.8 %	-16.0 %	-10.9 %
	Economic	-4.0 %	-0.8 %	5.0 %
MFRD	No allocation (BS)	-3.2 %	—	5.9 %
	Biophysical	-18.8 %	-16.0 %	-10.9 %
	Economic	-4.0 %	-0.8 %	5.0 %



Feed production Pullets production Manure-related emissions Electricity Other Inputs

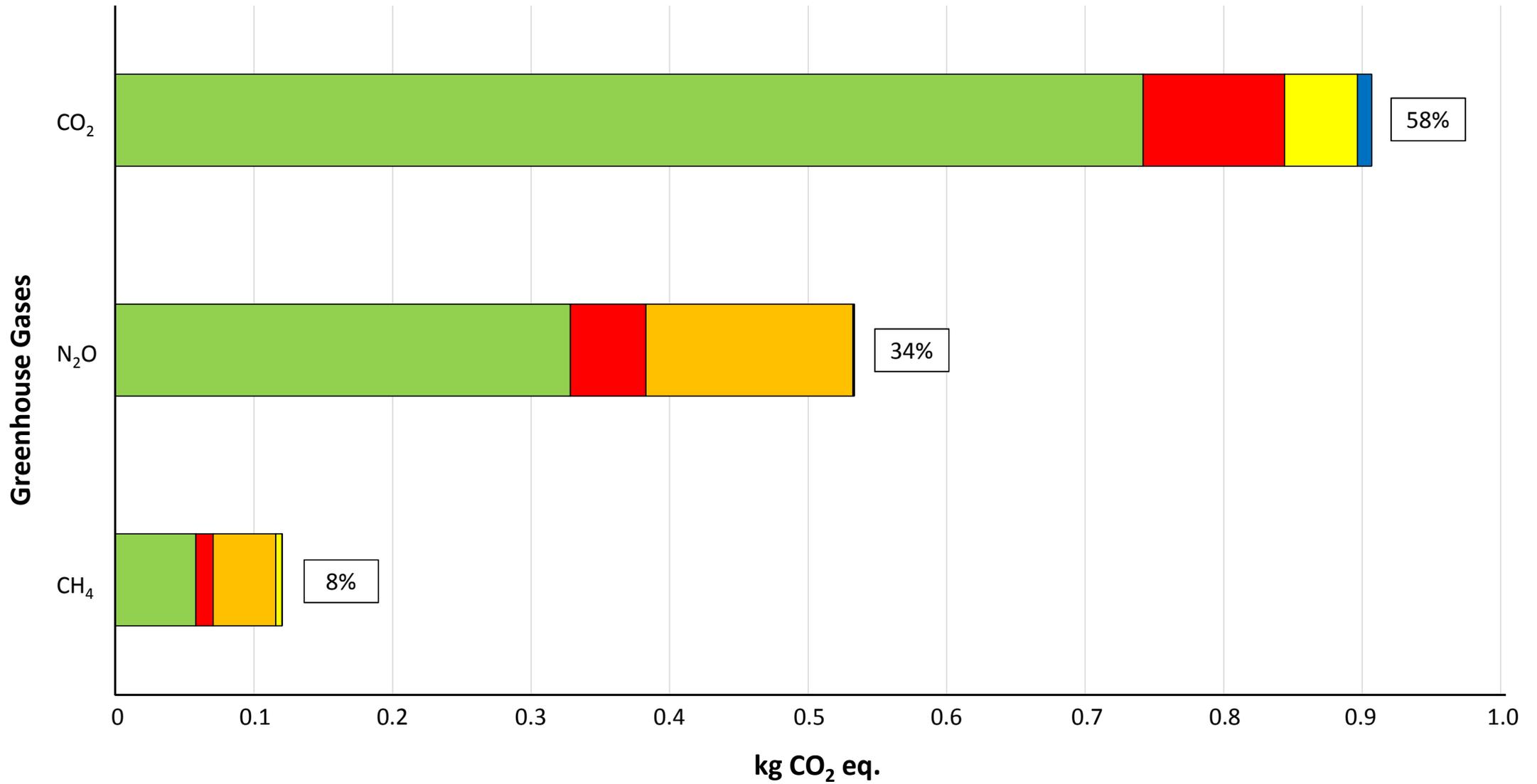


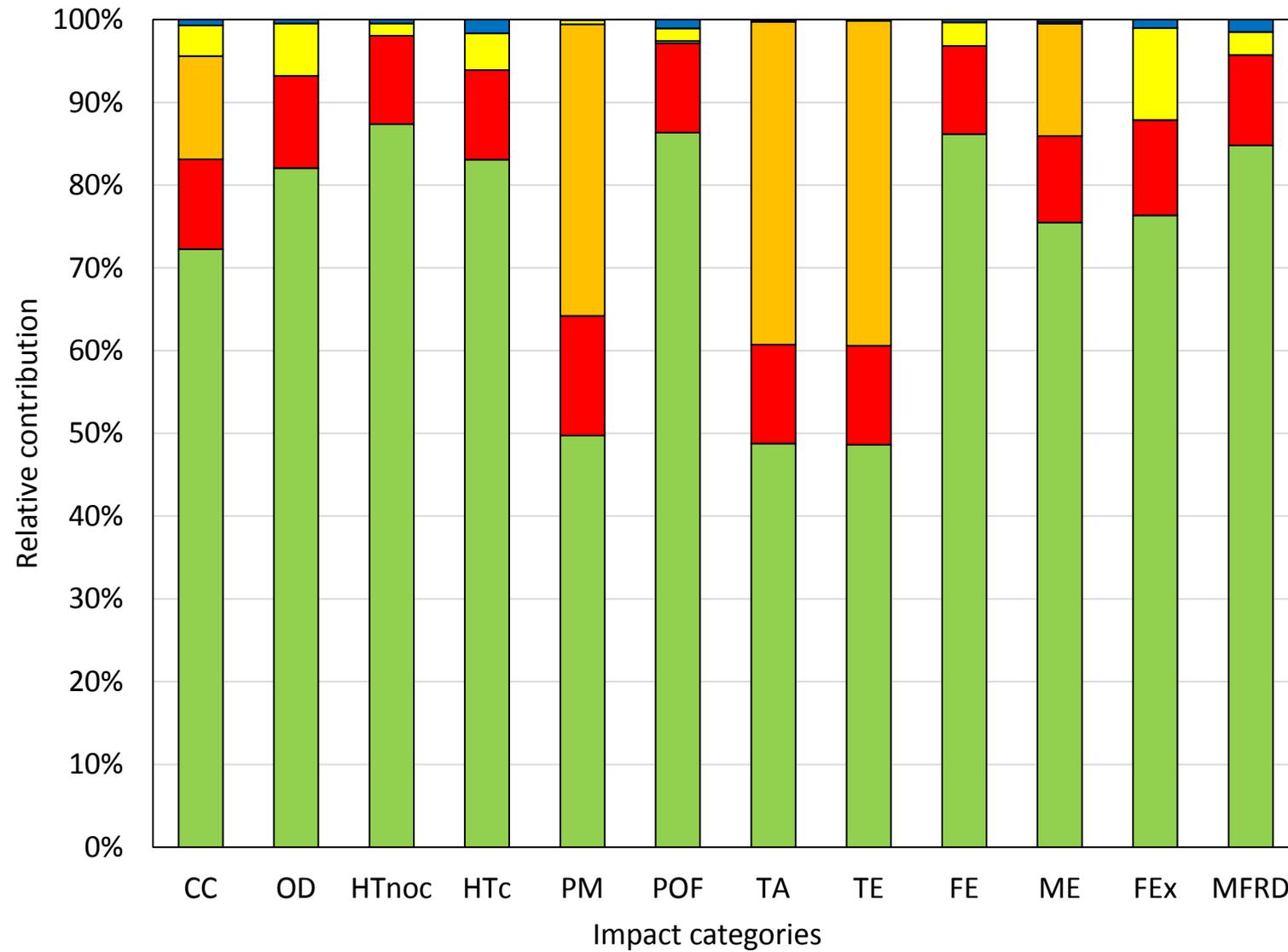
Figure 1 – System boundaries.

Figure 2 – Contribution analysis of organic egg production.

Figure 3 – Contribution analysis of greenhouse gases emission for 1 kg of organic eggs.

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■ Feed production ■ Pullets production ■ Manure-related emissions ■ Electricity ■ Other Inputs



SimaPro 8.5.2.0
Progetto

Valutazion Data: ##### Periodo: 23:11
_Pilmo

Calculation: Analizza
Results: Valutazione dell'impatto
Product: 1 kg Organic eggs production {ITA} (of project _Pilmo)
Metodo: ILCD 2011 Midpoint+ V1.10 / EC-JRC Global, equal weighting
Indicatore: Caratterizzazione
Skip categories: Mai
Esclude processi di infrastrutture: No
Esclude le emissioni di lungo termine: No
Sorted on item: Categoria d'impatto
Sort order: Ascendente

Etichetta	Unità	Totale	Organic eg	Transport,	Transport,	Transport,	BIO Laying	Mangime :	Mangime :	Tap water,	Erba medic	Sand {GLO	NO3- leac	NH3 emiss	CH4 Manu	N2O manu	Electricity,	medium voltage {IT}
Riscaldamento globale	kg CO2 eq	1.561819	0	0.00192	2.44E-05	0.00534	0.16933	0.75418	0.374685	0.001577	0.001551	0.000429	0	0	0.045274	0.149384	0.058127	
Assotigliamento strato ozono	kg CFC-11	1.02E-07	0	3.93E-12	4.99E-14	1.09E-11	1.13E-08	5.52E-08	2.83E-08	1.18E-10	2.76E-10	6.06E-11	0	0	0	0	6.44E-09	
Tossicità umana - Effetti non canc.	CTUh	5.13E-07	0	3.21E-11	4.07E-13	8.91E-11	5.49E-08	2.98E-07	1.5E-07	7.29E-10	1.37E-09	1.17E-10	0	0	0	0	7.6E-09	
Tossicità umana - Effetti canc.	CTUh	4.02E-08	0	1.44E-12	1.83E-14	4.01E-12	4.35E-09	2.23E-08	1.12E-08	4.44E-10	1.81E-10	2.8E-11	0	0	0	0	1.8E-09	
Formazione polveri sottili	kg PM2.5 eq	0.004107	0	1.85E-07	2.17E-09	5.16E-07	0.000593	0.001314	0.000729	7.43E-07	1.22E-06	3.45E-07	0	0.001448	0	0	2.02E-05	
Formazione smog	kg NMVOC	0.007625	0	1.5E-05	1.65E-07	4.17E-05	0.000826	0.004358	0.002225	3.98E-06	1.68E-05	2.88E-06	0	0	1.83E-05	0	0.000117	
Acidificazione	molc H+ eq	0.16786	0	1.18E-05	1.31E-07	3.28E-05	0.020034	0.052465	0.029415	8.26E-06	1.55E-05	3.34E-06	0	0.06554	0	0	0.000335	
Eutrofizzazione terrestre	molc N eq	0.745549	0	6.08E-05	6.64E-07	0.000169	0.089006	0.232169	0.130468	1.45E-05	6.05E-05	1.05E-05	0	0.292975	0	0	0.000616	
Eutrofizzazione acque dolci	kg P eq	0.000548	0	6.19E-09	7.86E-11	1.72E-08	5.82E-05	0.000318	0.000154	9.9E-07	5.91E-07	9.35E-08	0	0	0	0	1.58E-05	
Eutrofizzazione marina	kg N eq	0.01776	0	5.55E-06	6.06E-08	1.54E-05	0.001857	0.009049	0.004357	1.42E-06	2.07E-05	9.53E-07	0.000414	0.001997	0	0	4.22E-05	
Ecotossicità acque dolci	CTUe	3.173034	0	0.000818	1.04E-05	0.002275	0.366018	1.600739	0.821336	0.014505	0.011343	0.003025	0	0	0	0	0.352964	
Consumo di ris. minerali e fossili	kg Sb eq	2.7E-05	0	4.51E-10	5.73E-12	1.26E-09	2.95E-06	1.5E-05	7.86E-06	1E-07	2.6E-07	3.59E-08	0	0	0	0	7.53E-07	

Etichetta	Unità	Totale	
Riscaldamento globale	kg CO2 eq	1.562	1
Assotigliamento strato ozono	kg CFC-11	0.102	1000000
Tossicità umana - Effetti non canc.	CTUh	5.13E-07	1
Tossicità umana - Effetti canc.	CTUh	4.02E-08	1
Formazione polveri sottili	kg PM2.5 eq	4.107	1000
Formazione smog	kg NMVOC	7.625	1000
Acidificazione	molc H+ eq	0.168	1
Eutrofizzazione terrestre	molc N eq	0.746	1
Eutrofizzazione acque dolci	kg P eq	0.548	1000
Eutrofizzazione marina	kg N eq	17.760	1000
Ecotossicità acque dolci	CTUe	3.173	1
Consumo di ris. minerali e fossili	g Sb eq	26.979	1000000

market for | APOS, U

Journal Pre-proof

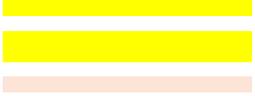
Esportato da SimaPro 8.5.2.0
 Esportato il: 24/01/2020 a23:11:42
 Analizzando 1 kg 'Organic eggs production (ITA)';
 Metodo: ILCD 2011 Midpoint+ V1.10 / EC-JRC Global, equal weighting / Caratterizzazione
 Unità usata %

Etichetta	Transport, truck >20t, EU	Transport, truck >20t, EL	Transport, truck >20t, EURO	BIO Laying hens <17 wex	Mangime 2 Bio pg 17%	Mangime 1 Bio pg 18.5%	Tap water, at user [Euro]	Erba medica per parchet	Sand [GLO]	market for NO3- leaching - uova bio	NH3 emissions - uova bio	CH4 Manure managemen	N2O manure managemen	Electricity, medium volta
Riscaldamento globale	0.1229	0.0016	0.3419	10.8418	48.2885	23.9903	0.1009	0.0993	0.0274	0	0	2.8988	9.5647	3.7217
Assotigliamento strato ozono	0.0039	4.9E-05	0.0107	11.152	54.248	27.8053	0.1163	0.2712	0.0596	0	0	0	0	6.3328
Tossicità umana - Effetti non canc.	0.0062	7.93E-05	0.0174	10.6937	58.1021	29.2657	0.1421	0.2677	0.0227	0	0	0	0	1.4823
Tossicità umana - Effetti canc.	0.0036	4.56E-05	0.01	10.8146	55.3692	27.7153	1.1037	0.45	0.0695	0	0	0	0	4.464
Formazione polveri sottili	0.0045	5.3E-05	0.0126	14.4358	32.001	17.7518	0.013	0.0297	0.0084	0	35.2459	0	0	0.4922
Formazione smog	0.1965	0.0022	0.5464	10.836	57.1515	29.1798	0.0522	0.221	0.0378	0	0	0.2399	0	1.5368
Acidificazione	0.007	7.82E-05	0.0196	11.935	31.255	17.5236	0.0049	0.0093	0.002	0	39.044	0	0	0.1995
Eutrofizzazione terrestre	0.0081	8.91E-05	0.0227	11.9383	31.1406	17.4995	0.0019	0.0081	0.0014	0	39.2965	0	0	0.0827
Eutrofizzazione acque dolci	0.0011	1.43E-05	0.0031	10.6226	58.0245	28.1592	0.1807	0.1079	0.017	0	0	0	0	2.8838
Eutrofizzazione marina	0.0312	0.0003	0.0869	10.4537	50.9541	24.5333	0.008	0.1163	0.0054	2.3313	11.242	0	0	0.2374
Ecotossicità acque dolci	0.0258	0.0003	0.0717	11.5353	50.4482	25.8849	0.4571	0.3575	0.0953	0	0	0	0	11.1239
Consumo di ris. minerali e fossili	0.0017	2.12E-05	0.0047	10.9403	55.6614	29.1329	0.3707	0.9648	0.1332	0	0	0	0	2.7903

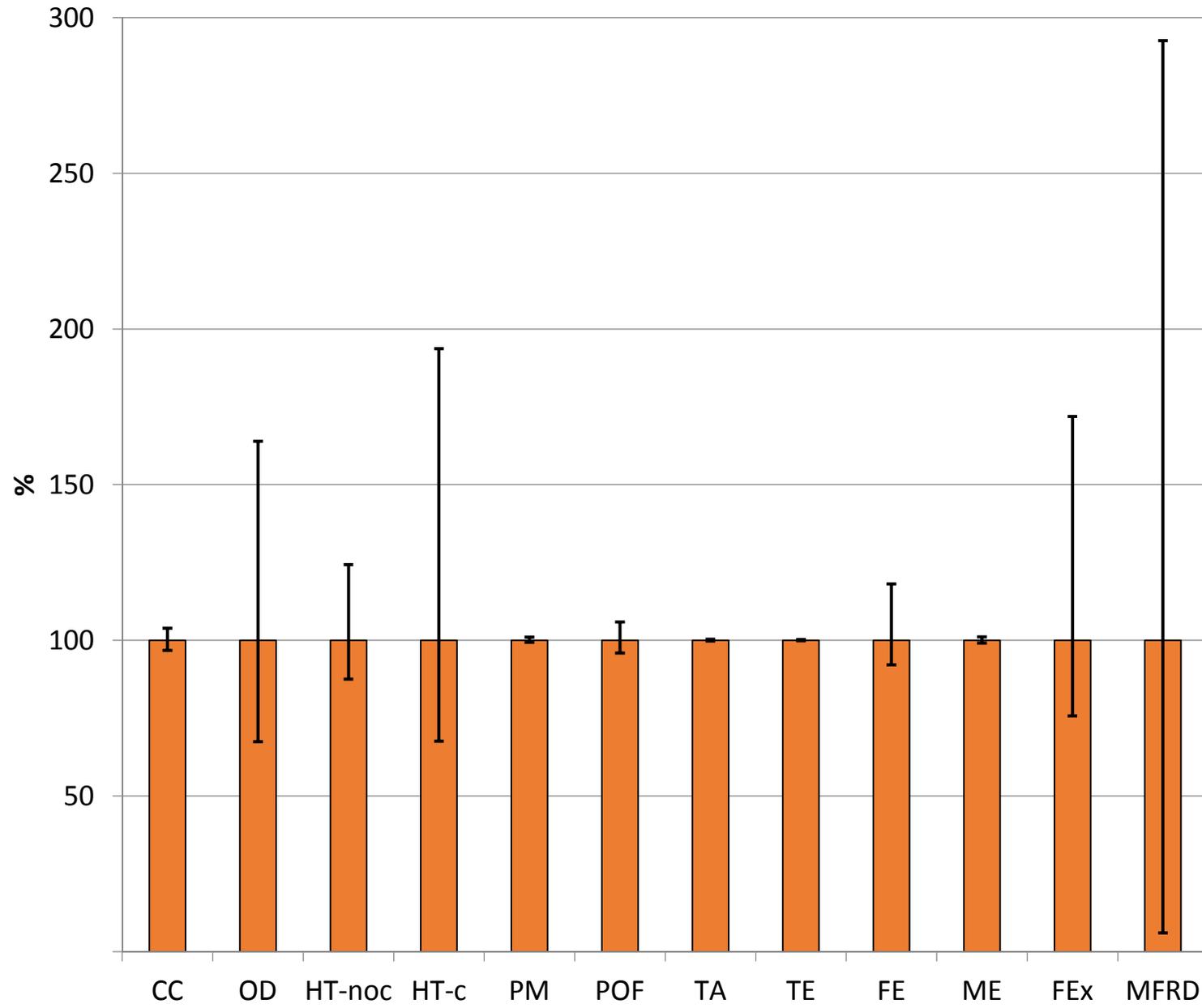
Etichetta	Trasporti	Altri Fattori Produttivi	Mangime 18,5% pg	Mangime 17% pg	Pollastre	Emissioni gestione pollina
Riscaldamento globale	0.4664	3.9494	23.9903	48.2885	10.8418	12.4635
Assotigliamento strato ozono	0.0147	6.7799	27.8053	54.2481	11.1521	0.
Tossicità umana - Effetti non canc.	0.0237	1.9148	29.2657	58.1021	10.6937	0.
Tossicità umana - Effetti canc.	0.0136	6.0873	27.7153	55.3692	10.8146	0.
Formazione polveri sottili	0.0171	0.5484	17.7518	32.0011	14.4358	35.2459
Formazione smog	0.745	1.8478	29.1798	57.1515	10.836	0.2399
Acidificazione	0.0267	0.2156	17.5236	31.255	11.935	39.0441
Eutrofizzazione terrestre	0.0309	0.0941	17.4995	31.1406	11.9383	39.2965
Eutrofizzazione acque dolci	0.0043	3.1895	28.1592	58.0245	10.6226	0.
Eutrofizzazione marina	0.1185	0.3671	24.5333	50.9541	10.4537	13.5733
Ecotossicità acque dolci	0.0978	12.0338	25.8849	50.4482	11.5353	0.
Consumo di ris. minerali e fossili	0.0063	4.259	29.1329	55.6614	10.9403	0.

	Contributo dei due mangimi sul totale del feed						
	Feed production	Pullets production	Manure-related emissions	Electricity	Other Inputs	Mangime 18,5% pg	Mangime 17% pg
CC	72.2788	10.8418	12.4635	3.7217	0.6941	33	67
OD	82.0534	11.1521	0.	6.3328	0.4618	34	66
HTnoc	87.3678	10.6937	0.	1.4823	0.4562	33	67
HTc	83.0845	10.8146	0.	4.464	1.6368	33	67
PM	49.7529	14.4358	35.2459	0.4922	0.0733	36	64
POF	86.3312	10.836	0.2399	1.5368	1.0561	34	66
TA	48.7786	11.935	39.0441	0.1995	0.0428	36	64
TE	48.6401	11.9383	39.2965	0.0827	0.0424	36	64
FE	86.1836	10.6226	0.	2.8838	0.3100	33	67
ME	75.4875	10.4537	13.5733	0.2374	0.2481	32	68
FEx	76.3331	11.5353	0.	11.1239	1.0078	34	66
MFRD	84.7943	10.9403	0.	2.7903	1.4751	34	66

Figure 17 | market for APOS, U



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SimaPro 8.5.2.0
Progetto

Valutazioni Data: ##### Periodo: 19:49
_Pilmo

Titolo:
Indicatore:

Analisi di incertezza di 1 kg 'Organic eggs production {ITA}', metodo: ILCD 2011 Midpoint+ V1.10 / EC-JRC Global, equal weighting, intervallo
Caratterizzazione

Categoria d'impatto	Unità	Mean	Median	Standard deviation	CV	2.5%	97.5%	SEM
CC	kg CO2 eq	1.562617	1.561087	0.028441	1.820093	1.510303	1.622086	0.000402
OD	kg CFC-11 eq	1.01E-07	9.71E-08	2.44E-08	24.18038	6.55E-08	1.59E-07	3.46E-10
HT-noc	CTUh	5.14E-07	5.05E-07	1.12E-07	21.82982	4.42E-07	6.28E-07	1.59E-09
HT-c	CTUh	4.03E-08	3.73E-08	1.74E-08	43.15835	2.52E-08	7.22E-08	2.46E-10
PM	kg PM2.5 eq	0.004108	0.004106	1.82E-05	0.443124	0.004079	0.004148	2.57E-07
POF	kg NMVOC	0.007632	0.007612	0.000201	2.628044	0.007303	0.008062	2.84E-06
TA	molc H+ eq	0.167881	0.167851	0.000272	0.162208	0.167469	0.168468	3.85E-06
TE	molc N eq	0.745615	0.745558	0.00081	0.108678	0.744202	0.747346	1.15E-05
FE	kg P eq	0.000549	0.000541	4.1E-05	7.477261	0.000498	0.000638	5.8E-07
ME	kg N eq	0.017763	0.017757	9.36E-05	0.52688	0.017598	0.017956	1.32E-06
FEx	CTUe	3.171724	2.994263	0.807895	25.47178	2.266677	5.147021	0.011425
MFRD	kg Sb eq	2.73E-05	2.49E-05	1.92E-05	70.32954	-3.7E-06	7.3E-05	2.71E-07

Intervallo di Confidenza:

95

ilto di confidenza: 95 %

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Unità usata %

Etichetta	Organic eggs production	Basso	Alto	
CC	100	3.253	3.9074	
OD	100	32.558	63.9114	
HT-noc	100	12.4889	24.3194	
HT-c	100	32.402	93.6918	
PM	100	0.6629	1.0304	
POF	100	4.0641	5.9017	
TA	100	0.2276	0.3674	
TE	100	0.182	0.2398	
FE	100	7.8803	18.0898	
ME	100	0.8985	1.1183	
FEx	100	24.2993	71.896	
MFRD	100	94.	192.6774	

, intervallo di confidenza: 95 %

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- 1 • Organic egg production was evaluated using the Life Cycle Assessment approach
- 2 • The environmental impact for 1 kg of organic eggs was evaluated
- 3 • A “from cradle to farm gate” approach was considered
- 4 • Feed supply is the main hotspot (from 49% to 87%) for all the impact categories
- 5 • For Climate Change (1.56 kgCO₂ eq/kg) the impact is lower than for conventional eggs

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Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

On behalf of all the Authors

Josef Beneš

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