

An Environmental Life Cycle Assessment of Living Wall Systems

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Abstract: The Life-Cycle Assessment (LCA) is a standard approach for evaluating the environmental impacts of products and processes. This paper presents the LCA of Living Wall Systems (LWS), a new technology for greening the building envelope and improve sustainability. Impacts of manufacture, operation, and use of the systems selected, were evaluated through an LCA. LWS are closely related to several environmental benefits, including improved air quality, increased biodiversity, mitigation of heat island effects, and reduced energy consumption due to savings in indoor cooling and heating. Two prototypes have been selected, taking into account the modularity and the use of organic substrate as selection criteria. The systems evaluated were a plastic-based modular system and a felt-based modular system. The inventory data was gathered through the manufacturers. The LCA approach has been used to assess the impact of these solutions by focusing on the construction phase and its contribution to both the energy balance and the entire life cycle of a building. The study found that out of the two systems through the manufacturing, construction, and maintenance stage of the LCA, the felt-based LWS has an impact on almost 100% of the impact categories analyzed, while plastic-based LWS has the lowest influence on the total environmental impact.

Keywords: Living Wall Systems, Life Cycle Assessment, Sustainability, Green Walls

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33 **Contents**
34

35 **Introduction**

36 Today, the European construction sector represents 40% of primary energy consumed from
37 non-renewable resources, out of a total of 87% globally. In turn, the human ecological footprint
38 has increased to 80% between 1960 and 2000 [1]. One of the most important challenges in
39 construction is the use of raw materials, and the implications in terms of energy balance,
40 consumption and the sustainability of the building during its useful life [2]. The reduction in
41 energy consumption and its associated emissions is a main issue in architecture and
42 engineering

43
44 The duality of the life cycle concept and the construction sector can be summed up in concepts
45 such as that of “low energy building” or “NZEB” (near zero energy buildings), which aims to
46 achieve the reduction of the impact on the environment during the building life cycle, the
47 minimisation of the energy and resources consumption, as well as land use [3]. An energy
48 efficient building uses active and passive technologies to counteract transmission heat loss that
49 affect energy consumption. The highest energy input in a building is found in the materials,
50 known as embodied energy. Dixit et al. [4] define the embodied energy like the energy
51 sequestered in buildings and building materials during the entire life cycle. The construction
52 sector has one of the most important environmental impacts on cities, and to face its
53 consequences and reduce energy consumptions is necessary to promote solutions with an
54 efficient performance during its entire lifecycle.

55 New technologies and building construction processes are being developed in order to improve
56 the sustainability and efficiency of building envelopes. Research has been carried out to develop
57 new adaptable and intelligent facades that highlight their thermal behaviour and adaptability to
58 different climatic contexts [5], within these, the vegetable façades are particularly noteworthy.

59 Greening the building envelope provides benefits related to improved efficiency, a contribution
60 to the immediate context through temperature regulation and reduced wind speed, as well as
61 increased biodiversity in dense urban environments [6]. Living wall systems (LWS) as part of
62 vertical green solutions can improve the quality of urban living and reduce the global
63 environmental impact caused by climate change [7]. The use of plants on buildings creating

64 green facades have aesthetical and environmental benefits [8]; improve the air quality by
65 reducing the air pollution [9,10] reduce fine dust levels in the air [11]; increase biodiversity [6];
66 reduce the heat island effect in cities [12,13], and reduce the energy consumption for indoor
67 cooling and heating [14,15]. Some of the aspects that influence the performance of a LWS are
68 the density of the foliage, the humidity of the substrate and the air chamber between some
69 layers, as well as the properties of the materials used [16].

70 The following studies investigated the ability of green facades and living wall systems to reduce
71 energy consumptions by intercepting solar radiation. A study carried out by C.Y Jim. et al., [21]
72 [17] studied the thermodynamic transmission process of the vertical vegetation ecosystem,
73 monitoring solar radiation and climatic conditions, and simulating heat flow and temperature
74 variations. Their results show that seasonal heat flows in the green wall will vary with fluctuating
75 meteorological driving forces, protecting the vegetation efficiency of the green wall that absorbs
76 radiant energy and prevents it from reaching the building surface. Coma J. et al. [18], studied
77 the behaviour of vegetal facades in a Continental Mediterranean climate during the summer.
78 The results show the capacity of vegetation to reduce the surface temperature of the exterior
79 façade by up to 14°C, and the effect of shade on the reduction of the internal temperature by up
80 to 1°C. Manso M. et al. [19], studied a modular system of vegetal façade called Geogreen,
81 through the analysis of local climatic conditions in three different periods. The experiment was
82 carried out based on two measurements, one on a reference wall and one on a wall covered
83 with vegetation modules. Results proved the capacity of vegetation to reduce maximum
84 temperatures and increase minimum temperatures. Specifically, the studied system has
85 demonstrated the ability to mitigate heat transfer up to a maximum of 75% input heat, and 60%
86 quality heat, improving thermal insulation. Nadia S. et al., [20] studied the influence of green
87 walls on the thermal behaviour of buildings in semi-arid regions during the summer period.
88 Outcomes showed that vegetation coverage optimises indoor temperature and reduces heat
89 exchange through the wall structure, characterized by reduced temperature and increased
90 relative humidity. Perez G. et al., [21] through research determined that the surface temperature
91 of a building wall in a shaded area was on average 5.5°C higher than in areas partially covered
92 by vegetation. This difference was greatest during the summer, reaching an average
93 temperature of 15.2°C on the southwest side in September. Olivieri F. et al., [22] carried out an

94 evaluation of the thermal behaviour of a modular plant façade on drainage cells, and the results
95 indicated that the performance of this pre-vegetated façade was better than a solar protection
96 system, since it reduced overheating by 33% in the cooling system compared to other ventilated
97 façade solutions. Mazzali U. et al., [23] tested three LWS to investigate the potential effects of
98 energy behaviour on building envelopes under different climatic conditions in Mediterranean
99 contexts. Their results showed similar behaviour in similar climatic conditions. During sunny
100 days the differences in air temperature of the vegetal wall were from a minimum of 12°C to a
101 maximum of 20°C, and during cloudy days the differences are reduced to 1°C-2°C. From these
102 studies, the capabilities of LWS as a technology to improve the performance and thermal
103 insulation of buildings are evident. Therefore, it can be said that these systems have the
104 capacity to limit the heat fluxes is the same in all the vertical greening systems. The differences
105 on the performance might be by the presence of factors like the foliage index, the moisture
106 content, vegetation type and materials involved.

107

108 Life Cycle Assessment (LCA) is one of many tools for assessing environmental issues. It is
109 defined by ISO 14040 as: *"A technique for assessing the environmental aspects and potential*
110 *impacts associated with a product, by compiling an inventory of relevant inputs and outputs of a*
111 *product system; evaluating the potential environmental impacts; and interpreting the results of*
112 *the inventory analysis and impact assessment phases"* [24]. The LCA approach has been used
113 in the construction sector since the 1990s [25], and its popularity is due to the compilation of all
114 material-related data and its environmental impact. It is a tool to promote sustainable design
115 and construction. Jeswani et al. [26], identified LCA like a systematic and robust tool for
116 quantifying potential environmental burdens and impacts of a process or product selection, and
117 also for improving design and optimization. When a building LCA is carried out, only the building
118 itself is studied, and the outcome is an assessment of the entire building process. In case the
119 LCA concerns a part of the building, such as a component or building material, the results might
120 be called *"building material and component combination"* (BMCC) [27].

121 According to the ISO standards, 14040/44 (ISO, 2006) [28] "a Life Cycle Assessment is carried
122 out in four distinct and interdependent phases:

123 - Goal and scope include functional unit selection and system boundary definition;

- 124 - Life cycle inventory involves the definition of energy and material flows between the
125 systems and the environment and through the different subsystems and operations of
126 the evaluated systems;
- 127 - Impact assessment, during which the inventory data are converted into environmental
128 indicators, discussion and interpretation of the results, the results from the inventory
129 analysis and impact assessment are summarized, sensitivity and uncertainty analysis
130 are carried out and recommendations are given”.

131

132 Many researchers have made LCA studies calculating the environmental impacts of some
133 construction materials to determine guidelines for the improvement of the building's
134 performance. Asif et al. [29], carried out a study of CO₂ emissions from eight different building
135 materials, including wood, concrete, aluminium, slate, glass, ceramics, and plasterboard. From
136 the study, it was concluded that the material with the highest emissions and energy
137 incorporated was concrete with 61%; Broun et al. [30], studied three types of partition walls from
138 a life-cycle approach: clay bricks, hollow concrete blocks, and a traditional wooden structure.
139 The results showed that the most relevant material is brick both in terms of energy consumption
140 and environmental impacts related to the life cycle. Kosareo et al. [31], conducted a LCA of
141 intensive and extensive green roofs through a comparison with conventional solutions. The
142 results obtained demonstrated the energy benefits provided by vegetation due to the lower
143 thermal conductivity of the substrate. Altan et al. [32], conducted the LCA of five different types
144 of green wall systems in the UK, researching the environmental impacts and benefits
145 associated with all phases of the life cycle. The results evidenced the lower impact of
146 continuous unsupported solutions due to the lower maintenance and reuse of their components.

147

148 Faced with this series of studies and proven benefits, in recent years numerous LWS solutions
149 have been launched on the market, among which the modular ones stand out. However, most
150 of the studies developed have to do with the performance during the use phase, without taking
151 into account the emissions and energy incorporated from manufacturing to disassembly. Living
152 Wall Systems can be assessed through LCA to study environmental impacts related to the
153 entire lifecycle. These results could be a useful support tool for researchers and manufacturers

154 in sustainable design [33]. Particularly, the building sector, LCA helps to evaluate the important
155 aspects related to embodied energy, embodied carbon and consumption energy of the
156 materials and greenhouse gases emissions [34].

157

158 **Objective**

159 The aim of this study is to evaluate the energy and environmental life cycle of two living wall
160 systems using different materials, types of assembly, and components. The purpose is to
161 quantify the impacts and benefits associated with the manufacture, construction, and
162 maintenance of a plastic and a geotextile based LWS.

163 A comparison of the results will also be carried out to obtain guidelines that will lead to
164 improving the environmental sustainability of the systems during their useful life. With the final
165 purpose of promoting design with less environmental impact and more environmentally
166 sustainable constructions. This study will help architects, ecologists, and engineers to find new
167 nature-based solutions to address the consequences of climate change from the construction
168 sector.

169 **Materials and methods**

170 **Functional unit**

171
172 According to ISO 14040 [35], the functional unit is the measurement value for quantifying the
173 results in an LCA. In this study, emissions, energy consumption, and materials are based on
174 1m² of LWS. The results of this analysis are calculated as the total environmental impact over
175 the lifetime, excluding the decommissioning phase. With this data, we can choose between
176 options and select the one that is compatible with the environment. The results show the total
177 environmental impact throughout the useful life of each system. Also, these results allow the
178 identification of improvements compatible with the concept of sustainability and environmental
179 awareness.

180 **System boundaries**

181 The system boundary comprises the manufacture of the system components, construction and
182 maintenance (fig.2).

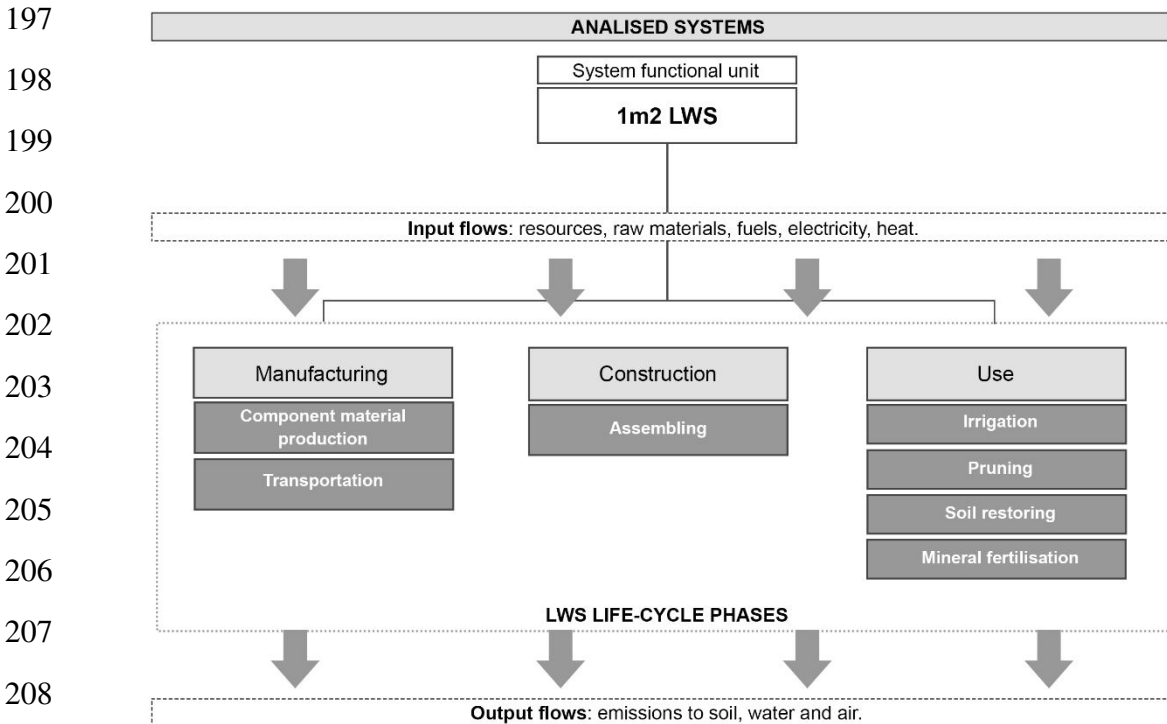
183 Manufacturing and construction cover the resources and process for producing the materials for
184 the system components. The construction phase comprises all electricity consumption per
185 square meter of LWS. The maintenance phase comprises the water consumption of the two
186 LWSs, based on both system requirements and fertilization. Finally, all activities related to the
187 use and disposal phases are excluded.

188 The study of the aspects that potentially affect the environment has been based on 10 years of
189 useful life. The data that has been supplied by the manufacturers. It is assumed that the useful
190 life of both LWS is 10 years, as well as that of all materials.

191 The replacement frequencies of plants for the LWS made in plastic are 10% replacement per
192 year, and 20% replacement per year for the system made with felt layers.

193 The LWS need a nutrient solution if has a non-organic substrate, which is considered only for
194 the system made with felt layers. The water consumption for the plastic-based LWS is assumed
195 to be 8 l/day and for the felt-based LWS on 2 l/day. Irrigation systems are not considered.

196



209 **Fig.2: Boundaries of the analyzed systems**

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214 ***Manufacturing stage***

215 The production phase focuses on analyzing the materials used to manufacture each of the
216 systems. This helps to understand the energy content of the materials and the carbon
217 emissions of the materials itself. The data was collected from the Ecoinvent Database v.3.5
218 [36].

219 During manufacturing, two methods were considered for the construction of the systems. In the
220 case of the LWS made with felt layers it is built by hand, which may require 1 to 2 people to
221 assembly. Thus, it is not necessary to use heavy machinery to assemble these systems. In the
222 case of LWS made of plastic, specialized machinery is required for their assembly, and it has an
223 electrical energy consumption of 0.044kWh for the production of panels and 0.8Wh for the
224 production of anchoring systems.

225

226 ***Construction stage***

227 In this phase, the assembly of each system and its materials, the mode of transport and the
228 distance traveled are analyzed, as well as the CO₂ emissions resulting from the transport of
229 these materials. These factors have been important in obtaining the total environmental impact
230 of each material during its life cycle. Each phase is calculated using SimaPro 8.5.

231

232 ***Maintenance stage***

233 The maintenance phase studies the life cycle burden of the two systems attributed to water
234 consumption, considering the number of times the systems need to be irrigated. This phase
235 helps to obtain data on the system with the greatest impact due to resource consumption. Water
236 consumption is an important factor that should be considered as it provides important insights
237 into the water input needed to keep systems operating throughout the useful life. In this case,
238 the plastic-based LWS has the highest water consumption (8 liters/m² per day), while the one in
239 the felt-based LWS is lower (2 liters/m² per day).

240

242 **Description of the studied LWS**

243 Living Wall Systems (LWS), are often built from modular panels, in which the substrate can be
 244 organic, from natural compounds such as hummus, or hydroponics, with an artificial culture
 245 media such as foam, felt, perlite or mineral wool, i.e., that uses nutrient solutions for fertilizing
 246 the plants [32]. Figure 1 shows the difference between a LWS made with felt layers (a), and
 247 LWS made with planter boxes (b).



262 **Fig.1:** (a) Living wall system based on planter boxes; (b) Living wall system made with felt layers

263

264 The characteristics of the two types of living wall systems used in this study are:

- 265 - *The felt modular system, a type of modular system that uses plants, which can be pre-*
- 266 *grown and inserted into gaps.. The system was produced by a Spanish company,*
- 267 *whose objective is to design and manufacture sustainable solutions to create horizontal*
- 268 *and vertical green spaces in urban environments. Its design was developed in the field*
- 269 *of air purification, to allow the growth of roots in contact with the air, favoring*

270 *biofiltration. Thus, the main objective is to decontaminate the air through the*
271 *rhizosphere of plants.*

272 - *The modular system in boxes is a vertical system formed by plastic modules. These*
273 *panels provide the rigidity and impermeability of the entire system. Vegetation can be*
274 *inserted before or after installation. This system requires an irrigation system and can*
275 *be automated.*

276 *This project has been carried out by a multidisciplinary group of Italian researchers in*
277 *collaboration with small companies with experience in prefabricated modular*
278 *construction, waste recycling, and textiles. The modules were designed, prototyped,*
279 *and implemented through an environmental approach based on the use of recycled*
280 *materials, high environmental performance, thermal, acoustic, and agronomic.*

281

282 Through an inventory analysis, the two LWSs have been analysed. The data about the
283 materials used in each system were collected from manufacturers and suppliers. A complete
284 LCA includes five different stages: manufacturing, construction, use, maintenance, and end of
285 life. In this study, only three phases have been considered: manufacturing, construction, and
286 maintenance.

287 The use phase has been excluded. It is assumed that the capacities of these systems in terms
288 of thermal insulation and temperature reduction are the same in all systems in which plants and
289 substrates are present, with some differences that are not relevant. This statement is supported
290 by Nyuk Hien Wong et al. [37], who studied 8 different vertical vegetation systems to evaluate
291 their thermal impacts on system performance. Their results demonstrated the same thermal
292 benefits in all system. These benefits minimize the demand for cooling and heating, and energy
293 costs in buildings.

294

295 Inventory data collection

296 The data and details of each system were gathered by the use of Ecoinvent Database v 3.5
297 [36], and also provided by the manufacturers. For this LCA, all the components of the two living
298 wall systems selected were examined. The differences between the two systems came from the
299 materials used and the way they are assembled. In the case of the LWS made of plastic, the
300 system has only one three-dimensional structure for the plants and another that serves as an air

301 chamber. The second LWS is made with felts, which involve several layers to root, waterproof,
302 and support.

303 The data used for this inventory was collected from material data sheets and information
304 obtained directly from manufacturers.

305 All elaboration phases play important roles in LCA studies but, the inventory analysis is
306 considered the most important [33]. The final product has been studied to calculate the impacts
307 related to its materials and processes. In this work, an inventory analysis was carried out by
308 obtaining information on the production, construction and maintenance of the systems.

309

310 LWS is used as an external surface of buildings that provides a thermal insulation benefit that
311 impacts on interior well-being. Modular LWSs are often made using a frame and a series of
312 layers that act as a climatic barrier to insulate the interior and exterior of the building. The
313 difference between the proportions of materials that impact the environmental load of the two
314 systems comes from the layers involved (Fig.3).

315 In the case of the plastic modular system, the layers consist of a box made of that material
316 which can be HDPE (High-Density Polyethylene), polypropylene and other recycled plastics,
317 filled with potting soil. In the case of the modular felt system, it has several layers to root,
318 waterproof, and support the substrate and plants.

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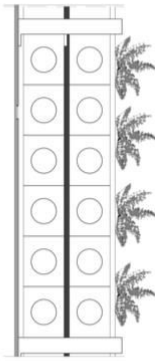
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Living Wall 1: plastic planter box



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Material	Thickness
Aluminum anchorage	0.2cm
Polyester fiber layer	0.2cm
Polypropylene box	5cm
Polyester fiber layer	0.2cm
Polypropylene box (with the growing medium)	5cm
Polyester fiber layer	0.2cm
Total thickness	10,8cm

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Living Wall 2: felt layer system



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Material	Thickness
Alveolar polycarbonate in Lexan resin	0.6cm
Growing medium	3.6cm
Polypropylene monofilament geomat-grid	3.6cm
Non-woven viscose fabrics layer	0.1cm
Aluminum alloy	0.2cm
Non-woven geotextile layer with a superficial polypropylene fiber layer	0.4cm
Total thickness	3.6cm

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Fig.3: Main components and thickness of the living wall systems studied

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All the transportation distances used are to and from Madrid. For the LWS the majority of the materials are local; plants and substrate come from an area 40km away from Madrid. The materials used in the LWS studied are an aluminum alloy, polypropylene monofilament, polypropylene fiber, growing medium, vegetal species biomass, felts and polyester. As for fertilizers, the following have been considered in the analysis 0.73 kg Nitrogen (N), 0.73 diphosphorus pentoxide (P₂O₅), and 0.73 potassium oxide K₂O.

355 The materials analyzed in each LWS are shown in tables 1 and 2. The raw materials,
 356 manufacturing energy use, and emissions associated with each of these materials were
 357 obtained from processes in the Ecoinvent Database v 3.5 [36].

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Components	Material	Mass (kg)	Distances (Km)		Service life (years)
External finishing layer	Polyester	0.25	50		10
Bearing structure	Polypropylene boxes	1.34	80		10
Hydrophilic layer	Polyester	0.25	50		10
	Coconut fibre, turf and				
Growing medium	humus	4	40		10
Closing layer	Polypropylene fibre and non-woven	0.25	50	80	10
External finishing layer	Polyester	0.53			10
Hooking system	geotextile				
	Aluminium	0.6	10		10
Bearing structure	Aluminium alloy			10	10
Vegetation layer	<i>Hedera spp</i> stems	3.9			
Hydrophilic layer	Non-woven viscose fabrics	1.50	40	50	10
	biomass	1.15			
Growing medium	Polypropylene monofilament geomat	2	50		10

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Table 1: Analysis of the components of the Living Wall System made with plastic planter boxes

containment layer	grid			
Growing medium	50% of raw soil; 30% of SAP; 15% of coco-coir; 5% of peat moss.	2.1	40	10
Closing layer	Alveolar polycarbonate in Lexan resin	2	50	10
Vegetation layer	Lonicera n. stems biomass	1.66	40	10

371

372 **Table 2:** Analysis of the components of the Living Wall System made with felt layers mass

373

374 **Life Cycle Impact Assessment**

375 The following impact categories were evaluated using the ILCD (International Reference Life
376 Cycle Data System) midpoint method [38], the LCIA method endorsed by the European
377 Commission:

- 378 - Climate Change (CC, expressed as kg CO₂ eq.);
- 379 - Ozone Depletion (OD, expressed as kg CFC-11 eq.);
- 380 - Particulate Matter Formation (PM, expressed as kg PM_{2.5} eq.);
- 381 - Human Toxicity-No Cancer Effect (HTnoc, expressed as CTUh);
- 382 - Human Toxicity-Cancer Effect (HTC, expressed as CTUh);
- 383 - Photochemical Ozone Formation (POF, expressed as kg NMVOC eq.);
- 384 - Terrestrial Acidification (TA, expressed as molc H⁺ eq.);
- 385 - Terrestrial Eutrophication (TE, expressed as molc N eq.);
- 386 - Freshwater Eutrophication (FE, expressed as kg P eq.);
- 387 - Marine Eutrophication (ME, expressed as kg N eq.);
- 388 - Freshwater Ecotoxicity (FEx, expressed as CTUe);
- 389 - Land Use (LU, expressed as kg C deficit);
- 390 - Water resource depletion (WU, expressed as m³ water eq.);
- 391 - Mineral and Fossil Resource Depletion (MFRD, expressed as kg Sb eq.)

392

393 **Results and discussion**

394 **Environmental impact of the LWS**

395 The results show that in every impact category evaluated, the plastic-based LWS is the one with
 396 the lowest environmental impact. The results show the highest impact of the systems in the
 397 manufacturing phase (Tables 4 and 5), and the use phase is the second with the highest
 398 impact.

399 Table 3 shows the environmental impacts for the LWS made with plastic. The results compare
 400 each phase studied concerning the impact categories, and agree with the previous works [37],
 401 where the LWS based on plastic boxes has no major environmental impact. The phase that
 402 affects in a non-proportional way in the impact categories is the manufacturing phase.

403 In the manufacturing phase all impact categories influence in almost the same way, excluding
 404 water resource depletion, which represents only 0.80% while the rest of the categories influence
 405 99% during the manufacturing process. The construction phase has a low influence during the
 406 study, with an average of 0.2% in all categories. The primary impact category for the use phase
 407 is water resource depletion, which represents 99.17% of the total, while the other categories
 408 have not an impact. The phase with the highest impact is the manufacturing phase, which is
 409 focused on analyzing the materials used for making the system. This explains the
 410 environmental impact contribution of the used materials.

411

412

Impact category	Unit of measure	Manufacturing	Construction	Maintenance
Climate change	kg CO ₂ eq	99.73%	0.26%	0.00%
Ozone depletion	kg CFC-11 eq	99.83%	0.16%	0.00%
Human toxicity, non-cancer effects	CTUh	99.99%	0.01%	0.00%
Human toxicity, cancer effects	CTUh	99.99%	0.01%	0.00%
Particulate matter	kg PM2.5 eq	99.87%	0.13%	0.00%
Photochemical ozone formation	kg NMVOC eq	99.86%	0.13%	0.00%
Acidification	molc H+ eq	99.76%	0.24%	0.00%
Terrestrial eutrophication	molc N eq	99.81%	0.19%	0.00%
Freshwater eutrophication	kg P eq	99.99%	0.00%	0.00%
Marine eutrophication	kg N eq	99.83%	0.17%	0.00%
Freshwater ecotoxicity	CTUe	99.99%	0.06%	0.00%
Land use	kg C deficit	97.13%	0.09%	0.00%
Water resource depletion	m ³ water eq	0.80%	0.03%	99.17%
Mineral, fossil & ren resource depletion	kg Sb eq	99.99%	0.01%	0.00%

420

421

Table 3: Environmental impacts for 1m² of the plastic-based LWS

422 Table 4 shows the environmental impacts for the system based on felts for the three phases
 423 considered. It is important to denote that the results, in this case, do not include any data
 424 related to the use of electrical energy for the construction of the system since it is done
 425 manually. The results are particularly higher to the system made in plastic. The impact
 426 generated by the system is concentrated in the manufacturing and use phase, in which it varies
 427 considerably according to the impact category.

Impact category	Unit of measure	Manufacturing	Construction	Maintenance
Climate change	kg CO ₂ eq	20.74%	0.00%	79.26%
Ozone depletion	kg CFC-11 eq	26.73%	0.00%	73.26%
Human toxicity, non-cancer effects	CTUh	44.60%	0.00%	55.40%
Human toxicity, cancer effects	CTUh	48.06%	0.00%	51.94%
Particulate matter	kg PM2.5 eq	35.14%	0.00%	64.86%
Photochemical ozone formation	kg NMVOC eq	35.27%	0.00%	64.72%
Acidification	molc H+ eq	23.46%	0.00%	76.53%
Terrestrial eutrophication	molc N eq	13.99%	0.00%	86.00%
Freshwater eutrophication	kg P eq	35.90%	0.00%	64.09%
Marine eutrophication	kg N eq	16.68%	0.00%	83.32%
Freshwater ecotoxicity	CTUe	60.04%	0.00%	39.95%
Land use	kg C deficit	5.20%	0.00%	94.79%
Water resource depletion	m ³ water eq	4.31%	0.00%	95.69%
Mineral, fossil & ren resource depletion	kg Sb eq	92.52%	0.00%	7.48%

435 **Table 4:** Environmental impacts for a 1m² of the felt-based LWS

436
 437 During the production phase, related to the use of materials, the greatest impact is given by
 438 mineral, fossil and renewable resource depletion with 92.52%, followed by freshwater
 439 ecotoxicity 60.04% and human toxicity cancer effects 48.06%. On the contrary, during the use
 440 phase, the categories with greater impact were water resource depletion 95,69%, land use
 441 94,79% and Ionizing radiation 90,33%. The rest of categories have an impact proportional to the
 442 previously mentioned. These results reveal the environmental impact that this system has
 443 related to the materials used and during the useful life considered as 10 years.

444 For both systems, the LCA shows that the highest environmental impacts are associated with
 445 the manufacturing and use phase, that accounts for more than 80% of the total environmental
 446 impact in almost all the categories analysed. It is particularly elevated for water resource
 447 depletion, land use, and mineral, fossil and renewable resource depletion. For these categories,
 448 the manufacturing phase accounts for 90-95% of the total environmental impacts.

449
 450 The main difference between the two LWS is mainly due to the materials involved in the
 451 anchorage and supporting systems. Figure 4 and 5 show the influence of the materials for the
 452 anchorage and supporting systems on the evaluated impact categories. Because of this, the

453 LWS plastic-based has the lowest environmental impact. In the case of the living wall system
454 made with felt layers, the fertilization has an impact of 99.17% on water resource depletion, due
455 to the necessity of doing annual chemical fertilizing.

456

457 For the impact categories related to toxicity and depletion of water resources, the plastic-based
458 early warning system has a double impact than the felt-based early warning system (Fig.4). The
459 results showed the environmental impact of two materials, mainly polypropylene and aluminium
460 layers. In this case, a solution could be to avoid the use of aluminum or to use recycled
461 aluminum, since the environmental impact can be reduced.

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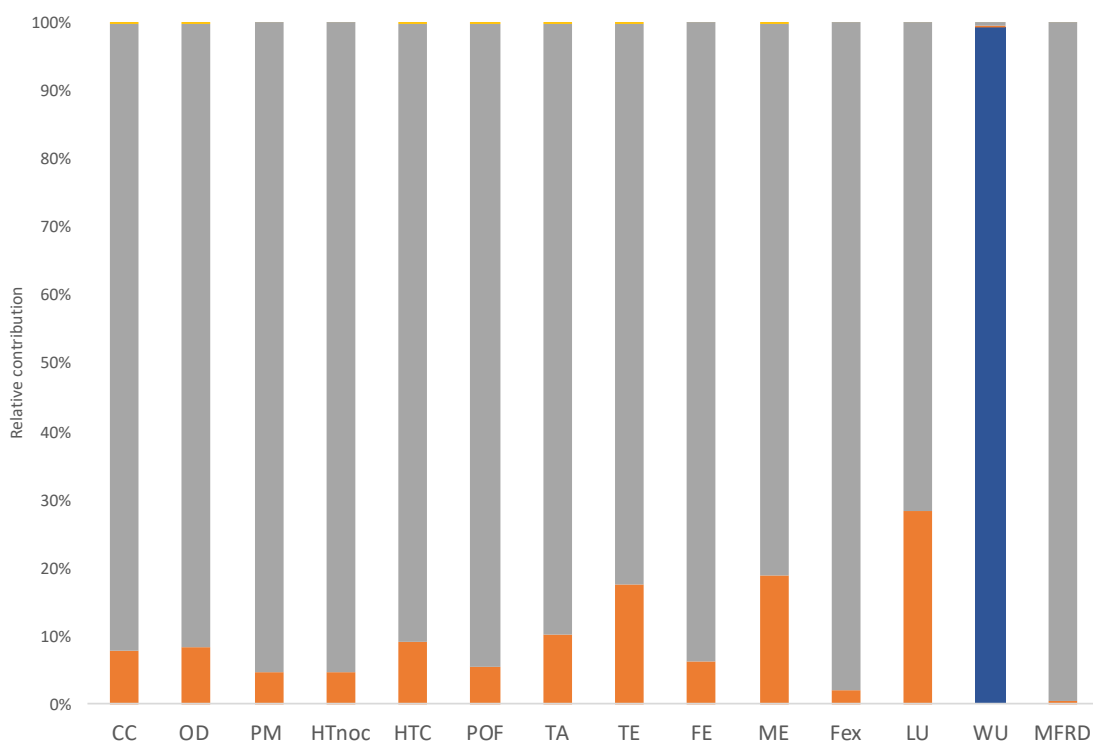
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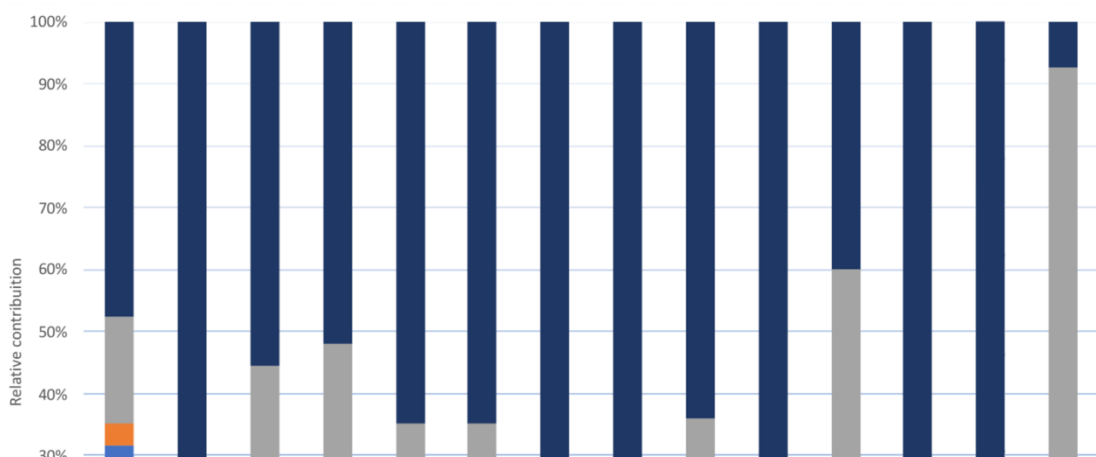
Fig.4: Environmental hotspots for the plastic-based LWS

The peat mixture used in the substrate has an impact on the category of water resource depletion, this is because peat is the result of the accumulation of dead organic matter from leaves, stems, and roots partially decomposed from different mosses and plants that have been concentrated in a water-saturated environment in the absence of oxygen.

The plastic-based LWS is a lightweight one due to the reduced number of materials, which means less energy consumption and less environmental impact. Thus, it could be used as a building element in buildings, in order to reduce energy consumption and energy incorporation.

Unlike this, the living wall system made with felt layers have the highest environmental impact in almost all the categories. This is due to the environmental impact coming from the use of aluminium for supporting the system and the use of fertilizers during the use phase of the system. Ottelé et al., [8] have investigated the environmental impact of four materials commonly used for the vertical support of living walls systems. Results show that aluminium can be up to 10 times more polluting than other materials such as plastics, wood and coated steel.

Both materials mentioned lead to increment the environmental burden profile. Furthermore, from figure 5, it can be seen that the LWS felt-based is the one without impacts in the construction phase because there are not electric energy consumptions associated. In this case, the highest environmental impacts in the use phase are due to the use of nitrogen fertilizer.



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Fig. 5: Environmental hotspots for the felt-based LWS

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522 The results obtained show the impact of the systems due to the materials used. This impact
523 could be reduced by a sustainable choice of materials. Specifically, the profile with the highest
524 impact is the LWS made with felt layers, due to the support system around 40% and the
525 fertilizers around 50% of the total impact.

526 In general, both systems can reduce impact by selecting a more sustainable material for the
527 support structure and other components such as the type of substrate and fertilization. In both
528 cases, reductions can be achieved with small changes. The impact categories analysed show
529 similar results, with some notable differences due mainly to the use of materials such as
530 aluminium and fertilization. For instance, for felt-based LWS, the most impactful categories are
531 freshwater eco-toxicity, land use and climate change, as the substrate needs to be fertilized ten
532 times in a 10-year lifespan. For the mineral, fossil and renewable resource depletion, both LWS
533 have a high impact. The same trend is perceptible for the freshwater eco-toxicity.

534

535 The relative comparison between the two systems studied is reported in **Fig.6**. For each
536 evaluated impact category, the LWS with the greatest impact is set equal to 100% while the
537 second one is proportionally called. LWS made with felt layers demonstrates the greatest
538 environmental burden for all impact categories assessed, except for the depletion of water
539 resources. This is consistent with the study of Ottelè et al., [8], which conducted a life cycle
540 analysis comparing conventional brick solutions with continuous and modular plant facades,
541 including systems made of plastic and felt. Great differences were found in the impact

542 categories studied for each alternative plant façade. In that case, the results were influenced by
 543 the type of material used for each system.

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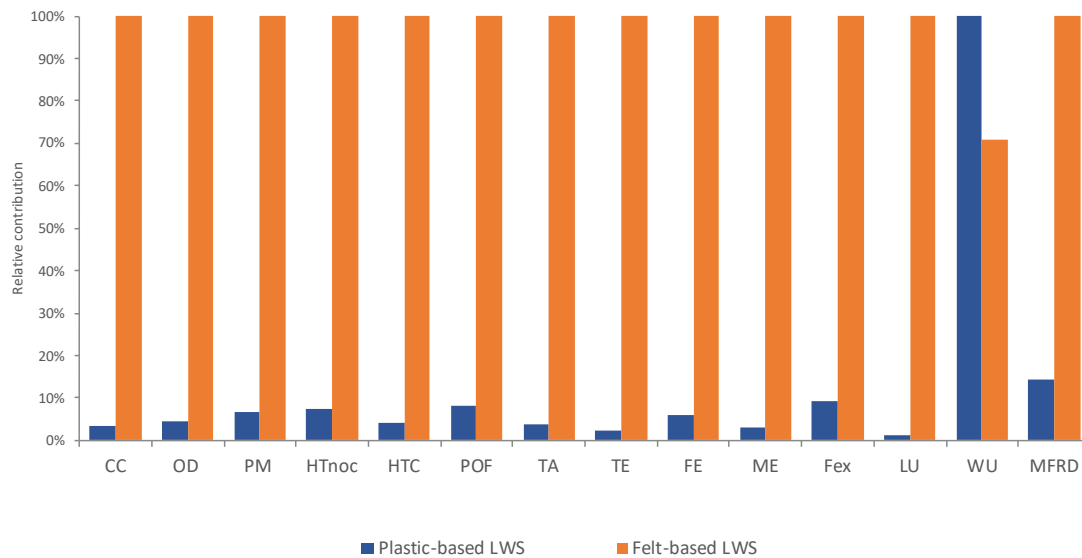
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Fig.6: Comparison between the two LWS

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556
 557 Among the evaluated impact categories, water resource depletion is the only one for which the
 558 LWS made by plastic shows a higher impact, this is linked to the irrigation needs of the system.

559 For the other categories, it is clear that the LWS based on felts is the one with the highest
 560 environmental impact due to the composition of the materials used and the fertilization.

561 However, despite their environmental impact, the two LWS can counteract them through its
 562 reduction in energy consumption and temperatures.

563 Other authors [8,32,39] have reached similar results considering the entire life cycle of the
 564 systems and studying vegetable façade systems different from ours. It has been demonstrated

565 that, even if we do not consider the whole life cycle and exclude some phases, the results agree
 566 that the performance of the systems is the same whenever there is the presence of substrate

567 and vegetation. Thus, the environmental impact will depend on the materials used for
 568 construction, and the substances used during maintenance according to the type of substrate.

569 Besides, they argue that from the results of the LCA, it is possible to make improvements in the
 570 systems, which in some cases mean that the benefit is twice as great as the impact they can

571 generate. This benefit is related to the temperature reduction potential.

572

573 Life Cycle Impact Assessment

574 This section aims to weight the results of the entire analysis. The most impacting phases are
575 shown for each category in Figure 7. The data represents the impact caused for 1m² of LWS.

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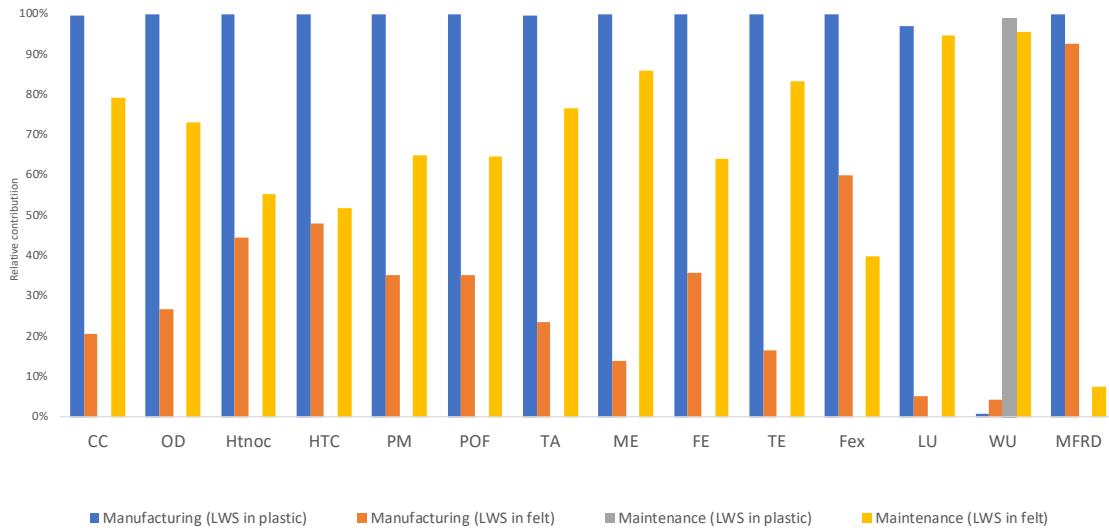
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Fig.7: Impact categories per LWS studied. A comparison based upon LCA results.

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588 The phase with the greatest weight in the process is the manufacturing phase, linked to the
589 materials and assembly processes. The results were analyzed by comparing the systems. The
590 impact of LWS made of plastic during manufacturing is notable due to the electricity
591 consumption and the use of aluminum for the anchoring system. In the case of the climate
592 change impact category, the difference is almost 80%. While the felt-based LWS has its 100%
593 during the maintenance phase, due to the fertilizers used during its life cycle.

594

595 Limitations and future perspectives

596 In this study, it is assumed that the two living wall systems have the same thermal and
597 environmental performances and the behavior of a plant façades during their life cycle is out of
598 the system boundaries of this LCA study. On the one hand, as there are no monitoring data for
599 the systems studied, there is no possibility of verifying their performance. In the same line of
600 ideas, today there are no tools in which it is possible to simulate the reduction of energy

601 consumption and temperatures to obtain a value. Also, the benefits of plant façades go far
602 beyond the effect of thermal insulation; fundamental effects such as evapotranspiration, shade,
603 acoustic insulation and the fixation of dust particles would be out of the study.

604 This study analyzed the living wall life cycle impact only in the phase of manufacturing,
605 construction, and maintenance, to identify how the selection of materials affects, which is
606 associated with an important series of environmental benefits.

607 Unlike other studies [8,31,32,40] in which these technologies and their materials are studied to
608 identify how they affect their energy performance. These parameters should be explored in
609 future comprehensive studies. However, even if the use phase is not included in the system
610 boundary the achieved results can be useful. In fact, the study, quantifying the environmental
611 impact and identifying the environmental hotspots (i.e., the process mainly responsible of the
612 environmental impact) of the two LWS, is the starting point for a subsequent optimization.

613 **Conclusions**

614 This study helps designers and technology developers to understand the potential and the
615 environmental concerns associated to LWS. Also, it is a starting point for identifying the best
616 option on the market by understanding the impacts of the various lifecycle phases through the
617 LCA approach.

618 The materials used to build an LWS have a significant environmental impact when installed in a
619 building. From the incorporated and operational energy of a building, the role of the materials is
620 fundamental, as it can be reduced depending on the proper selection of the materials.

621 Life cycle analysis of living wall systems considers several aspects, including integration into the
622 building envelope, the selection of materials with low environmental impact and the
623 consideration of other impacts, which can contribute to the correct decision when incorporating
624 it as a sustainable technical solution.

625 The results of the LCA performed highlight the environmental impact of two LWS: a modular
626 system made with solid plastic boxes and pre-cultivated vegetation inserted in cavities, and a
627 system based on layers of felt with pre-cultivated vegetation inserted in pockets, both with
628 aluminium anchoring system.

629 From the research during the three selected phases, it is clear that each LWS has strengths
630 and weaknesses:

- 631 - plastic-based LWS shows lower impact during the manufacturing, construction, and
632 maintenance phases.
- 633 - The environmental impact of plastic-based LWS shows a lower impact respect to the
634 felt-based LWS due to the low mass of materials used. This impact could be reduced
635 further reduced by replacing materials like polyester with other recycled textiles and
636 recycled aluminium for the system anchors.
- 637 - The felt-based LWS has an aluminium support that deeply affects the environmental
638 load. With this regard, to improve the system towards a more environmentally
639 sustainable one the design and research activities should focus on the identification of
640 less impacting materials. Besides this, the use of fertilizers during the life cycle involves
641 a significant impact, a less impacting option would be the use of an organic fertilisers or
642 leguminous crops.

643

644 Greening the building envelope with LWS taking into account the materials involved is a key
645 step in selecting a solution that leads to an environmentally friendly performance. This study
646 highlighted that the use of recycled materials, organic substrates, and low environmental impact
647 materials are part of the sustainable strategies for the design of these systems. These should
648 be considered as key strategies for the environment, sustainability, and low energy
649 consumption of LWS, throughout their life cycles.

650

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655 **References**

- 656 [1] Y.A. Izrael, S.M. Semenov, O.A. Anisimov, Y.A. Anokhin, A.A. Velichko, B.A. Revich,
657 I.A. Shiklomanov, The Fourth Assessment Report of the Intergovernmental Panel on
658 Climate Change: Working Group II contribution, 2007.
659 doi:10.3103/S1068373907090014.
- 660 [2] M. Weißenberger, W. Jensch, W. Lang, The convergence of life cycle assessment and

- 661 nearly zero-energy buildings: The case of Germany, *Energy Build.* 76 (2014) 551–557.
662 doi:10.1016/j.enbuild.2014.03.028.
- 663 [3] I. für Bauforschung, *Energetische Gebäudemodernisierung Fraunhofer IRB*, Stuttgart.,
664 2010.
- 665 [4] M.K. Dixit, J.L. Fernández-Solís, S. Lavy, C.H. Culp, Need for an embodied energy
666 measurement protocol for buildings: A review paper, *Renew. Sustain. Energy Rev.* 16
667 (2012) 3730–3743. doi:https://doi.org/10.1016/j.rser.2012.03.021.
- 668 [5] M. Iommi, The mediterranean smart adaptive wall. An experimental design of a smart
669 and adaptive facade module for the mediterranean climate, *Energy Build.* 158 (2018)
670 1450–1460. doi:https://doi.org/10.1016/j.enbuild.2017.11.025.
- 671 [6] K. Perini, M. Ottel , E.M. Haas, R. Raiteri, O.M. Ungers, Greening the building envelope
672 , facade greening and living wall systems, *Open J. Ecol.* 1 (2011) 1–8.
673 doi:10.4236/oje.2011.11001.
- 674 [7] N. Dunnett, N., Kingsbury, *Planting Green Roofs and Living Walls.*, Portland, Or., 2008.
- 675 [8] M. Ottel , K. Perini, A.L.A.L.A.A. Fraaij, E.M.M. Haas, R. Raiteri, Comparative life cycle
676 analysis for green faades and living wall systems, *Energy Build.* 43 (2011) 3419–3429.
677 doi:10.1016/j.enbuild.2011.09.010.
- 678 [9] S. Gourdji, Review of plants to mitigate particulate matter, ozone as well as nitrogen
679 dioxide air pollutants and applicable recommendations for green roofs in Montreal,
680 Quebec, *Environ. Pollut.* 241 (2018) 378–387. doi:10.1016/j.envpol.2018.05.053.
- 681 [10] J. Klingberg, M. Broberg, B. Strandberg, P. Thorsson, H. Pleijel, Influence of urban
682 vegetation on air pollution and noise exposure – A case study in Gothenburg, Sweden,
683 *Sci. Total Environ.* 599–600 (2017) 1728–1739. doi:10.1016/j.scitotenv.2017.05.051.
- 684 [11] K. Perini, M. Ottel , S. Giulini, A. Magliocco, E. Roccotiello, Quantification of fine dust
685 deposition on different plant species in a vertical greening system, *Ecol. Eng.* 100 (2017)
686 268–276. doi:10.1016/j.ecoleng.2016.12.032.
- 687 [12] L. Mariani, S.G. Parisi, G. Cola, R. Laforteza, G. Colangelo, G. Sanesi, Climatological
688 analysis of the mitigating effect of vegetation on the urban heat island of Milan, Italy, *Sci.*
689 *Total Environ.* 569–570 (2016) 762–773. doi:10.1016/j.scitotenv.2016.06.111.
- 690 [13] S. Sheweka, N. Magdy, The living walls as an approach for a healthy urban

- 691 environment, *Energy Procedia*. 6 (2011) 592–599. doi:10.1016/j.egypro.2011.05.068.
- 692 [14] L. Pan, L.M. Chu, Energy saving potential and life cycle environmental impacts of a
693 vertical greenery system in Hong Kong: A case study, Elsevier Ltd, 2015.
694 doi:10.1016/j.buildenv.2015.06.033.
- 695 [15] K. Perini, P. Rosasco, Cost-benefit analysis for green facades and living wall systems,
696 *Build. Environ.* 70 (2013) 110–121. doi:10.1016/j.buildenv.2013.08.012.
- 697 [16] UK Green Wall Association, UK Guide to Green Walls, (2013).
- 698 [17] C.Y. Jim, H. He, Estimating heat flux transmission of vertical greenery ecosystem, *Ecol.*
699 *Eng.* 37 (2011) 1112–1122. doi:10.1016/j.ecoleng.2011.02.005.
- 700 [18] J. Coma, G. Pérez, C. Solé, A. Castell, L.F. Cabeza, New green facades as passive
701 systems for energy savings on Buildings, *Energy Procedia*. 57 (2014) 1851–1859.
702 doi:10.1016/j.egypro.2014.10.049.
- 703 [19] M. Manso, J.P. Castro-Gomes, Thermal analysis of a new modular system for green
704 walls, *J. Build. Eng.* 7 (2016) 53–62. doi:10.1016/j.jobbe.2016.03.006.
- 705 [20] S. Nadia, S. Noureddine, N. Hichem, D. Djamila, Experimental study of thermal
706 performance and the contribution of plant-covered walls to the thermal behavior of
707 building, *Energy Procedia*. 36 (2013) 995–1001. doi:10.1016/j.egypro.2013.07.113.
- 708 [21] G. Pérez, L. Rincón, A. Vila, J.M. González, L.F. Cabeza, Behaviour of green facades in
709 Mediterranean Continental climate, *Energy Convers. Manag.* 52 (2011) 1861–1867.
710 doi:10.1016/j.enconman.2010.11.008.
- 711 [22] F. Olivieri, L. Olivieri, J. Neila, Experimental study of the thermal-energy performance of
712 an insulated vegetal façade under summer conditions in a continental mediterranean
713 climate, *Build. Environ.* 77 (2014) 61–76. doi:10.1016/j.buildenv.2014.03.019.
- 714 [23] U. Mazzali, F. Peron, P. Romagnoni, R.M. Pulselli, S. Bastianoni, Experimental
715 investigation on the energy performance of Living Walls in a temperate climate, *Build.*
716 *Environ.* 64 (2013) 57–66. doi:https://doi.org/10.1016/j.buildenv.2013.03.005.
- 717 [24] B. Fava, J.A.; Consoli, F.; Dension, R.; Dickson, K.; Mohin, T.; Vigon, A Conceptual
718 Framework for Life-Cycle Impact Assessment, *Soc. Environ. Toxicol. Chem. SETAC.*
719 (1993).
- 720 [25] J.A. Fava, Will the next 10 years be as productive in advancing life cycle approaches as

- 721 the last 15 years?, *Int. J. Life Cycle Assess.* 11 (2006) 6–8.
- 722 [26] H.K. Jeswani, A. Azapagic, P. Schepelmann, M. Ritthoff, Options for broadening and
723 deepening the LCA approaches, *J. Clean. Prod.* 18 (2010) 120–127.
724 doi:<https://doi.org/10.1016/j.jclepro.2009.09.023>.
- 725 [27] M.M. Khasreen, P.F.G. Banfill, G.F. Menzies, Life-cycle assessment and the
726 environmental impact of buildings: A review, *Sustainability.* 1 (2009) 674–701.
727 doi:10.3390/su1030674.
- 728 [28] I.O. for S. (ISO), *Environmental Management e Life Cycle Assessment e Requirements*
729 *and Guidelines. ISO 14044.*, n.d.
- 730 [29] M. Asif, T. Muneer, R. Kelley, Life cycle assessment: A case study of a dwelling home in
731 Scotland, *Build. Environ.* 42 (2007) 1391–1394.
732 doi:<https://doi.org/10.1016/j.buildenv.2005.11.023>.
- 733 [30] R. Broun, G.F. Menzies, Life cycle energy and environmental analysis of partition wall
734 systems in the UK, *Procedia Eng.* 21 (2011) 864–873.
735 doi:10.1016/j.proeng.2011.11.2088.
- 736 [31] L. Kosareo, R. Ries, Comparative environmental life cycle assessment of green roofs,
737 *Build. Environ.* 42 (2007) 2606–2613. doi:10.1016/j.buildenv.2006.06.019.
- 738 [32] H. Altan, N. John, J. Yoshimi, T. Ilyas, M. Galadari, Comparative life cycle analysis of
739 green wall systems in the uk, (n.d.).
- 740 [33] C. Ingrao, A. Matarazzo, C. Tricase, M.T. Clasadonte, D. Huisingh, Life Cycle
741 Assessment for highlighting environmental hotspots in Sicilian peach production
742 systems, *J. Clean. Prod.* 92 (2015) 109–120.
743 doi:<https://doi.org/10.1016/j.jclepro.2014.12.053>.
- 744 [34] T. Malmqvist, M. Glaumann, S. Scarpellini, I. Zabalza, A. Aranda, E. Llera, S. Díaz, Life
745 cycle assessment in buildings: The ENSLIC simplified method and guidelines, *Energy.*
746 36 (2011) 1900–1907. doi:10.1016/j.energy.2010.03.026.
- 747 [35] I.O. for S. (ISO), *Environmental Management e Life Cycle Assessment e Principles and*
748 *Framework. ISO 14040.*, n.d.
- 749 [36] B. Wernet, G., Bauer, C., Steubing, B., Reinhard, J., Moreno-Ruiz, E., and Weidema,
750 *The ecoinvent database version 3 (part I): overview and methodology.*, (2016).

- 751 <http://link.springer.com/10.1007/s11367-016-1087-8>.
- 752 [37] N.C.N.H. Wong, A.Y. Kwang Tan, Y. Chen, K. Sekar, P.Y. Tan, D. Chan, K. Chiang,
753 N.C.N.H. Wong, Thermal evaluation of vertical greenery systems for building walls,
754 *Build. Environ.* 45 (2010) 663–672. doi:10.1016/j.buildenv.2009.08.005.
- 755 [38] European Commission, International Reference Life Cycle Data System (ILCD)
756 Handbook - General guide for Life Cycle Assessment - Provisions and Action Steps,
757 2011. doi:<http://dx.doi.org/10.2788/94987>.
- 758 [39] M. Manso, J. Castro-Gomes, B. Paulo, I. Bentes, C.A. Teixeira, Life cycle analysis of a
759 new modular greening system, *Sci. Total Environ.* 627 (2018) 1146–1153.
760 doi:10.1016/j.scitotenv.2018.01.198.
- 761 [40] C. Ingrao, F. Scrucca, C. Tricase, F. Asdrubali, A comparative Life Cycle Assessment of
762 external wall-compositions for cleaner construction solutions in buildings, *J. Clean. Prod.*
763 124 (2016) 283–298. doi:10.1016/j.jclepro.2016.02.112.

764

765 **Nomenclature**

- 766 Kg CO² eq - Climate change (CC)
- 767 Kg CFC-¹¹ eq - Ozone depletion (OD)
- 768 Kg PM_{2.5} eq – Particulate matter formation (PM)
- 769 CTU_h – Human toxicity-no cancer effect (HT_{noc})
- 770 CTU_h – Human toxicity-cancer effect (HTC)
- 771 Kg NMVOC eq – Photochemical ozone formation (POF)
- 772 molc H⁺ eq – Terrestrial acidification (TA)
- 773 molc N eq – Terrestrial eutrophication (TE)
- 774 Kg P eq - Freshwater eutrophication (FE)
- 775 Kg N eq – Marine eutrophication (ME)
- 776 CTU_e – Freshwater ecotoxicity (FEx)
- 777 Kg C deficit – Land use (LU)
- 778 M³ water eq – Water resource depletion (WU)
- 779 Kg Sb eq – Mineral and fossil resource depletion (MFRD)
- 780 LWS – Living wall system

781 LCA – Life cycle assessment

782 LCI – Life cycle inventory

783