

1 **Screening of diesters of ortho phthalic acid in printed baby bibs in the European market**

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8

9 **Abstract**

10 A screening exercise targeting ortho-phthalates in twenty two plastic baby bibs collected  
11 in European market was performed. A GC-MS method was used with a limit of detection  
12 ranging from 0.1 to 0.6 mgkg<sup>-1</sup> bib. The most frequently detected phthalates were di-n-  
13 butyl phthalate and di-iso-butyl phthalate, in nearly all samples. The latter was detected  
14 often in higher concentration than the first one, suggesting its intended use. Overall, the  
15 highest levels were detected for benzyl butyl phthalate, with 6 samples presenting  
16 concentration from 13 - 47 mgkg<sup>-1</sup> and one sample with the highest value of 65 mgkg<sup>-1</sup>.  
17 Results indicate that several non-authorised phthalates are being used either in the  
18 plastic or in the printing inks. Worst case migration calculations indicate that DEP, DAP  
19 and DIBP exceed the 0.01 mgkg<sup>-1</sup>, and therefore, determination of experimental  
20 migration is needed to conclude on compliance of baby bibs with the European and  
21 Swiss legislation. Bibs are considered, accordingly to European legislation, as FCMs and  
22 therefore, they should comply with the applicable rules, restrictions and limits. These  
23 articles should be included in surveillance plans, focusing in monitoring bibs  
24 composition, migration and the application by industry of good manufacturing practices.

25

26 **Keywords:** phthalates, plastic baby bibs, potential migration, plasticisers

27

28 Introduction

29 Diesters of ortho-phthalic acid (dialkyl or alkyl esters of 1,2-benzenedicarboxylic acid),  
30 commonly referred to as phthalates, are amongst the most common organic  
31 contaminants in both the environment and consumer products. Phthalates are not  
32 chemically bound to polymer chains and therefore they may migrate into food at rates  
33 and levels, which like for other migrants, depend on the contacting conditions (Poças,  
34 2018).

35 Exposure to phthalates is well known to have adverse effects on the health and  
36 development of humans, especially for children. Humans can be exposed to these  
37 phthalates through various routes, the most concerning being ingestion (EFSA, 2019).  
38 Infants and young children, due to their hand-to-mouth behaviour when handling  
39 different articles, are potentially more exposed than adults. Furthermore, these groups  
40 of population have specific physiological and developmental characteristics, as well as  
41 nutritional needs and food consumer patterns different than adults. These differences  
42 justify a particular attention in risk assessment of chemicals they are exposed to. This  
43 has impact on the rules and criteria applicable to foods and specific materials and  
44 articles intended for children consumption, use and handling.

45 Phthalates are used as plasticizers in many plastics applications including packaging and  
46 other food-contact materials (FCM), such as tubing and equipment parts used in food  
47 processing, handling and storage (Pereira et al. 2019) and are also used in non-plastic  
48 articles, such as printing inks, coatings and adhesives. Higher molecular weight  
49 phthalates, such as di-2-ethylhexyl phthalate (DEHP), are primarily used as plasticizers  
50 to soften polyvinyl chloride (PVC) products, while the lower molecular weight  
51 phthalates, such as diethyl phthalate (DEP), di-n-butyl phthalate (DBP), and butyl benzyl  
52 phthalate (BBP), are widely used as solvents and technical support agents or production  
53 aids such as for example pasting agent for catalysts (Cao et al., 2010). There is a trend  
54 for replacing ortho phthalates by other plasticizers families, such as terephthalates,  
55 benzoates, citrates, trimellitates, carboxyl adipates, phosphates and polymeric  
56 plasticizers.

57 Several phthalates are authorised for use in plastic FCM under the European harmonised  
58 legislation (EU, 2011). From those, DEHP, DBP and BBP are identified as substances of  
59 very high concern (SVHC) under the REACH Regulation due to their toxicity for  
60 reproduction and endocrine disrupting properties for human health. DEHP is  
61 additionally considered to be endocrine disruptor for the environment (EFSA, 2019). The  
62 Commission Regulation (EU) 2018/2005 (updates the REACH - Regulation (EC) No  
63 1907/2006) restricts the level of the phthalates DEHP, BBP, DBP and di-iso-butyl  
64 phthalate (DIBP) in any plasticized material in articles used by consumers or in indoor  
65 areas, including toys and childcare products, after June 2020. However, this restriction  
66 does not apply to food contact materials that are regulated by Regulation (EU) No  
67 10/2011. It is worthy to note that the Regulation (EU) 2018/2005 also updated the  
68 definition of “childcare” to the following: “childcare article’ shall mean any product  
69 intended to facilitate sleep, relaxation, hygiene, the feeding of children or sucking on  
70 the part of children”. Plastic baby bibs are used for assisting on feeding infants and  
71 toddlers, so stakeholders may wrongly assume that bibs should comply with the REACH  
72 regulation only and ignore the need for complying with the food contact rules, which is  
73 more restrictive. While the REACH regulation defines limits for the concentration of  
74 phthalates in the plasticised material in articles (individually or in any combination, no  
75 greater than 0.1 % (m/m)), the FCM legislation sets lower concentration limits, but  
76 additionally sets restrictions regarding the type of article and migration limits.

77 According European framework legislation, baby bibs are considered a food contact  
78 material (FCM) as they can reasonably be expected to be brought into contact with food  
79 or to transfer their constituents to food under normal or foreseeable conditions of use  
80 (European Union Regulation nº 1935/2004). During its use, the baby spilt food and saliva  
81 on the bib, which is then typically fed again into the baby mouth, by the care-taker  
82 (Galbiati et al., 2020). Additionally, babies can chew and suck the bib. This product has  
83 not been targeted in surveys regarding phthalates in food contact materials. The  
84 objective of this work was to screen plastic baby bibs, collected in the European market,  
85 for ortho phthalates, to determine the potential of migration and to pinpoint the need  
86 for addressing these articles in surveillance plans.

87 Material and methods

88 *Samples*

89 Twenty-two bibs were purchased from all around Europe: 4 in Italy, 5 in Spain, 5 in  
90 Portugal, 3 in Ireland, 2 in Slovenia, 2 in Germany and 1 from Denmark. All samples were  
91 purchased directly in specialized or non-specialized stores and none was acquired by e-  
92 commerce channel. In order to prevent any cross contamination, samples were  
93 manipulated and stored wrapped in aluminum foil during the transport to the laboratory  
94 in Portugal. The material of the bibs was confirmed by (DSC) Dynamic Scanning  
95 Calorimetry (Shimadzu Corporation), through the determination of thermal transitions.  
96 Most of the bibs were made of polyethylene/vinyl acetate (PEVA) and a few of them  
97 were in polyamide (PA) and PE. The material grammage was in average  $1.5 \text{ gdm}^{-2}$ .

98 *Phthalates analysed*

99 The phthalates analysed are described in Table 1. From these DBP, BBP, DEHP, DINP and  
100 DIDP are authorized under the Regulation nº 10/2011 applicable to plastics in contact  
101 with food, with the following respective specific migration limits (SMLs): 0.3, 30, 1.5 and  
102 9  $\text{mgkg}^{-1}$  for the sum of both DINP and DIDP. DAP is authorized with a migration level  
103 that should be non-detectable, which in practice is assumed as  $0.01 \text{ mgkg}^{-1}$  as for the  
104 non-authorized phthalates. The standards were supplied by Fluka® (DMP, DEP, DIBP,  
105 DNOP, DIDP), Sigma Aldrich® (DAP, DBP, DEHP, DCHP), Chem Service® (DHP), Jayflex®  
106 (BBP, DINP) and AccuStandard® (DIAP). The identification of the detected phthalates  
107 was performed by matching spectra with library NIST MS Search (Version 2.2, 2014),  
108 following confirmation with standards.

109 *Background contamination management*

110 To minimise the analytical background levels of those phthalates that are ubiquitous,  
111 the following measures were taken:

112 - All the material used to handle the samples were cleaned between samples to avoid  
113 cross contamination;

114 - Glassware rinsed with hexane and acetone; then heated at 160 °C (overnight) prior to  
115 use;

116 - Disposable glass centrifuge test tubes used for extraction and concentration steps;

117 - Chromatographic analyses undertaken in a different laboratory from the one used for  
118 preparation of analytical standards and handling the samples;

119 - Residual contamination in blank injections subtracted from samples;

120 - Solvent injection after calibration curve analysis and between every samples;

121 - Triple super clean gas filter installed into gas supply of gas chromatograph.

#### 122 *Qualitative analysis of phthalates*

123 Ca 1 g of each sample was cut (3x3 mm), weighted and extracted with 10 ml  
124 dichloromethane (DCM) with the internal standard (Benzophenone d10) at 0.05 mgkg-  
125 1. Extraction conditions were: 40 °C for 24 hours with occasional stirring, followed by  
126 ultrasonic bath for 30 min, centrifugation (2500 rpm for 7 min) and filtration (0,45 µm  
127 PTFE membrane) before chromatography analysis. A blank of DCM was also prepared.

#### 128 *Quantitative analysis of phthalates*

129 Calibration curves, using deuterated phthalates (Sigma-Aldrich®) as internal standards,  
130 were prepared according to the regulatory status of the specific phthalate, in the range  
131 of 0-0.4 mgL<sup>-1</sup> for all phthalates, except for DIDP and DINP that the range of 0-2 mgL<sup>-1</sup>  
132 was applied. DBP-d4 (CAS n° 93952-11-5) was used as internal standard for DMP, DEP,  
133 DAP, DIBP and DBP, DEHP-d4 (CAS n° 93951-87-2) was used for DHP, BBP and DEHP, and  
134 DNOP-d4 (CAS n° 93952-13-7) for DNOP, DINP, DIDP. The calibration solutions were  
135 prepared in the extraction solvent (DCM).

136 Ca 1 g of each sample in replicate was cut, weighted and extracted with 10 ml  
137 dichloromethane (DCM) with the internal standard solution of deuterated phthalates at

138 0.25 mgL<sup>-1</sup>. The same extraction conditions as for qualitative analyses were applied  
139 following chromatography analysis. A blank of DCM was also prepared.

#### 140 *Analysis by gas chromatography GC-MS*

141 The analyses were performed in a system Bruker® Scion 456 TQ GC/MS, with the  
142 following operation conditions in the qualitative analyses: vector gas Helium at 1  
143 ml/min; injection temperature 320 °C; column Supelco® SLB-5ms (30m x 0,25 mm I.D.  
144 Df= 0,25 µm); oven temperature 40 °C for 5 min, rate of 10 °C/min up to 320 °C and 320  
145 °C for 25 min; volume of injection of 1 µl, splitless time 0,5 min. In the quantitative  
146 analyses slightly different conditions were applied: oven temperature 80 °C for 1 min,  
147 rate of 10 °C/min up to 320 °C and 320 °C for 8 min. Between each sample and each  
148 standard solution, a run with DCM was made in order to clean the equipment.

149 The mass detector was operated in the electronic impact mode at 70 eV; transfer line  
150 temperature 300 °C and source temperature 320 °C. Full scan between 33 and 700 m/z  
151 for the qualitative analyses and selected ion monitoring (SIM) mode according to Table  
152 1 for the selected phthalates.

153

## 154 Results and Discussion

### 155 *Method performance*

156 Figure 1 shows a chromatogram of the calibration solution: the total ion current (TIC)  
157 and extracted-ion chromatograms according to each phthalate. The main parameters of  
158 the method performance: limit of detection (LOD), limit of quantification (LOQ) and  
159 linear range are reported in Table 2 for each phthalate. A good linearity of response was  
160 achieved in the range considered. The coefficients of determination ( $R^2$ ) were higher  
161 than 0,999 for all phthalates. LOD and LOQ were calculated from the calibration curve,  
162 as 3 and 10 times the standard deviation of the intercept, respectively and, when  
163 transformed into the bib concentration base, were in the order 0.1 to 0.2 mgkg<sup>-1</sup> for all  
164 phthalates except for DEHP, DINP and DIDP for which the values ranged between 0.3  
165 and 0.6 mgkg<sup>-1</sup>.

166 *Phthalates quantification in bibs*

167 The phthalates concentration in bibs are shown in Table 3. The values are expressed as  
168 concentration in  $\text{mgkg}^{-1}$  bib  $\pm$  standard deviation. A multi-factor ANOVA was performed  
169 (Fisher Least Significant Difference; p-value  $<0.05$ ; different letters in the same column  
170 mean significant differences).

171 The most frequently detected phthalates were DBP and DIBP, in nearly all samples, and  
172 DEP and BBP in 16 out of 22 samples. The highest levels were detected for BPP, with 6  
173 samples presenting concentration from 13 to  $47 \text{ mgkg}^{-1}$  and one sample with the highest  
174 value of  $65 \text{ mgkg}^{-1}$ . DIBP was quantified in one sample at  $30 \text{ mgkg}^{-1}$ . DBP and DEP  
175 concentration averaged around  $2 \text{ mgkg}^{-1}$ . DEHP was detected in 6 samples, with values  
176 up to  $24 \text{ mgkg}^{-1}$ .

177 It seems there is a fair correlation between DIBP and DBP (Pearson coefficient 0.895),  
178 with concentration values for DIBP always higher than the ones for DBP in most samples  
179 (Figure 2). This seems to suggest the intended use of DIBP either in the main plastic or  
180 in the printing inks in spite of being not authorized in plastics FCM, according to  
181 European harmonized legislation for plastics. It is neither authorized in printing inks,  
182 according to Swiss legislation (FSVO, 2019). DIBP is authorized by U.S. Food and Drug  
183 Administration (FDA) in adhesives to be used (i) separated from the food by a functional  
184 barrier, (ii) in dry foods or (iii) when in contact with fatty and aqueous foods only through  
185 edges and seams of laminates. DIBP is also addressed in the German BfR  
186 Recommendations on Food Contact Materials to be used in plasticizer-free PVC and  
187 copolymers and in unsaturated polyester resins but only as pasting agent for catalysts  
188 (BfR, 2019) and with the same restrictions applicable to DBP in accordance with the Reg.  
189 (EU) No 10/2011 (EU, 2011). Recently, the European Food Safety Authority (EFSA)  
190 updated its opinions published in 2005 on DBP, BBP, DEHP, DINP and DIDP authorised  
191 for use as plasticisers and technical support agents in plastic for food contact (EFSA,  
192 2019). However, EFSA did not consider DIBP because it is not authorised for use in plastic  
193 FCM. Nevertheless, EFSA recognized that DIBP can substantially add to the overall  
194 exposure of consumers to phthalates because of similar intake estimates compared to  
195 DBP. Furthermore, biomonitoring studies have indicated a trend towards increased

196 human exposure to DIBP as a consequence of replacement for DBP (Zota et al. 2014).  
197 Therefore, it is worthy to note that the use of DIBP in FCMs and consequent exposure  
198 should be better characterized in spite of being a phthalate not authorized for food  
199 contact.

200 DHP and DNOP were not detected in any of the samples analysed and DMP was detected  
201 in one sample at low level. DAP was detected in 2 samples out of 22, in one sample at  
202 ca 1 mgkg<sup>-1</sup> and in another sample at a level of 5 mgkg<sup>-1</sup>. This result may also be worthy  
203 to further develop because DAP has a migration level that should be non-detectable.  
204 DINP was detected in 3 samples with the highest concentration of 7 mgkg<sup>-1</sup> and DIDP  
205 was detected in one sample only at 2.5 mgkg<sup>-1</sup>.

#### 206 *Phthalates migration estimates*

207 The results of phthalates concentration in the bibs were then translated into the  
208 migration assuming a total mass transfer and the conventional ratio of 6 dm<sup>2</sup> kg<sup>-1</sup> food.  
209 These are shown in Figure 3 as a radar plot of the concentrations (after normalizing with  
210 logarithms of the concentration values). Each line represents a bib sample and a  
211 comparison is made with the specific migration limits reported in Regulation (EC) No  
212 10/2011 (black dotted line).

213 Worst case migration estimates for BBP are well below the SML (30 mgkg<sup>-1</sup>) for all  
214 samples and the same applies to DBP (0.3 mgkg<sup>-1</sup>), DEHP (1.5 mgkg<sup>-1</sup>), DINP and DIDP  
215 (sum 9 mgkg<sup>-1</sup>). The non-authorized phthalates DMP, DHP and DNOP also have worst-  
216 case estimated migration levels lower than the applicable limit of 0.01 mgkg<sup>-1</sup> for all bibs  
217 tested.

218 The worst-case estimated migration for DEP and for DIBP, which are non-authorized  
219 phthalates, exceed the applicable value of 0.01 mgkg<sup>-1</sup> in 9 and 16 out of 22 samples,  
220 respectively. The maximum estimated migration was 0.1 mgkg<sup>-1</sup> for DEP and 0.24 mgkg<sup>-1</sup>  
221 for DIBP and both occurred in one sample only.

222 As indicated, DIBP was the phthalate with higher number of undue occurrences and also  
223 showing the highest levels of estimated migration. A recent systematic review,



224 corroborating previous studies, provided evidence that DIBP causes male reproductive  
225 and developmental toxicity and slight evidence for female reproductive toxicity, among  
226 other effects, supporting DIBP as a children's health concern (Yost et al., 2019).  
227 Furthermore, studies on the association of early life exposure to metabolites of several  
228 phthalates, including DBP and DIBP, with obesity and cardiovascular risks in childhood  
229 suggest that early life exposure may affect child growth and adiposity - mono-isobutyl  
230 phthalate (MiBP), the metabolite of DIBP was associated with higher total cholesterol  
231 levels at 4–6 years (Vafeiadi et al., 2018).

232 Calculated values for migration are recognized to overestimate the real migration  
233 values, but nevertheless are indicative of the need for surveillance particularly because  
234 bibs are often wrongly not seen by industry as food contact materials, and surveillance  
235 actions are not systematically acted. In fact, the nature of use of the article in question  
236 (baby bibs) is relatively uncertain regarding the contact conditions that affect migration,  
237 namely regarding the time of use. Temperature of contact can be well estimated as this  
238 is not much different than the baby body temperature. The time of contact depends on  
239 the frequency and duration of meals but the practice of keeping the bib on the baby  
240 between meals is not uncommon (Rajbux et al, 2020). According to the JRC Guidelines  
241 on testing conditions for articles in contact with foodstuffs (Beldi et al., 2019), the  
242 migration test would be performed with simulants A (10% ethanol v/v), B (3% acetic acid  
243 (w/v) and D2 (vegetable oil), with 0.5 hours of contact at 40 °C and under repeat use  
244 conditions. However, it is well known that most often babies were the bib for much  
245 longer time and they chew and suck the material, therefore increasing the potential for  
246 intake of migrants.

247

## 248 Conclusions

249 This work focused on screening plastic baby bibs to detect and identify phthalates and  
250 to determine their potential migration. Results indicate that several non-authorized  
251 phthalates are being used either in the plastic or in the printing inks. The most frequently  
252 phthalates detected were DBP and DIBP, the latter present often in higher concentration

253 than the first one, suggesting its intended use. Worst case migration calculations  
254 indicate that DEP, DAP and DIBP exceed the 0.01 mgkg<sup>-1</sup>, and therefore, determination  
255 of experimental migration is needed to conclude on compliance of baby bibs with the  
256 European and Swiss legislation. Bibs are considered, accordingly to European legislation,  
257 as FCMs and therefore, they should comply with the applicable rules, restrictions and  
258 limits. These articles should be included in surveillance plans, focusing in monitoring bibs  
259 composition, migration and the application by industry of good manufacturing practices.

## 260 References

261 G. Beldi; N. Jakubowska; P. Robouch; E. Hoekstra.2019. Testing conditions for  
262 kitchenware articles in contact with foodstuffs - Part 1: Plastics, JRC 116750

263 BfR. 200. Database BfR Recommendations on Food Contact Materials, Accessed via:  
264 [http://bfr.ble.de/kse/faces/DBEmpfehlung\\_en.jsp](http://bfr.ble.de/kse/faces/DBEmpfehlung_en.jsp).

265 Cao X-L. 2010. Phthalate Esters in Foods: Sources, Occurrence, and Analytical Methods.  
266 *Comprehensive Reviews in Food Science and Food Safety*, 9, 21-43

267 EFSA CEP Panel (EFSA Panel on Food Contact Materials, Enzymes and Processing Aids),  
268 Silano V, Barat Baviera JM, Bolognesi C, Chesson A, Cocconcelli PS, Crebelli R, Gott DM,  
269 Grob K, Lampi E, Mortensen A, Rivièrè G, Steffensen I-L, Tlustos C, Van Loveren H, Vernis  
270 L, Zorn H, Cravedi J-P, Fortes C, Tavares Poças MF, Waalkens-Berendsen I, Wölfle D,  
271 Arcella D, Cascio C, Castoldi AF, Volk K and Castle L, 2019. Scientific Opinion on the  
272 update of the risk assessment of di-butylphthalate (DBP), butyl-benzyl-phthalate (BBP),  
273 bis(2-ethylhexyl)phthalate (DEHP), di-isononylphthalate (DINP) and di-  
274 isodecylphthalate (DIDP) for use in food contact materials. *EFSA Journal*  
275 2019;17(12):5838, 85 pp. <https://doi.org/10.2903/j.efsa.2019.5838>

276 EU COMMISSION REGULATION (EU) No 10/2011 of 14 January 2011 on plastic materials  
277 and articles intended to come into contact with food, OJ L 12, 15.1.2011, p. 1

278 FSVO (Federal Food Safety and Veterinary Office). Ordinance of the FDHA on materials  
279 and articles intended to come into contact with food-stuffs, 3 October 2019, RO 2019  
280 3371

281 Galbiati E., Pereira J., Selbourne M.C., Poças F. 2020. Photoinitiators use in printed baby  
282 bibs and their migration into Tenax<sup>®</sup> by gas chromatography-mass spectrometry  
283 (GC/MS). *Packaging Technology and Science*. <https://doi.org/10.1002/pts.2546>

284 Pereira J., Selbourne M.C., Poças F. 2019. Determination of phthalates in olive oil from  
285 European market, *Food Control*, 98 (4), 54-60  
286 <https://doi.org/10.1016/j.foodcont.2018.11.003>

- 287 Poças, F. (2018). Migration from packaging and food contact materials into foods.  
288 Reference module in food science. Elsevier97800810059651–18.  
289 <https://doi.org/10.1016/B978-0-08-100596-5.21460-1>.
- 290 Rajbux, C., Pereira, J., Selbourne, C., Costa-Pinto, A.R., Poças, F. 2020. Assessment of  
291 baby Bibs. GC-MS screening, migration into saliva and insight of toxicity with QSAR tools.  
292 Food Control. <https://doi.org/10.1016/j.foodcont.2019.106951>
- 293 Vafeiadi M, Myridakis A, Roumeliotaki T, Margetaki K, Chalkiadaki G, Dermitzaki E,  
294 Venihaki M, Sarri K, Vassilaki M, Leventakou V, Stephanou EG, Kogevinas M and Chatzi  
295 L (2018). Association of Early Life Exposure to Phthalates With Obesity and  
296 Cardiometabolic Traits in Childhood: Sex Specific Associations. Front. Public Health  
297 6:327. doi: 10.3389/fpubh.2018.00327
- 298 Yost, E., Euling S., Weaver, J., Beverly,B.,Keshava, N., Mudipalli, A., Arzuagab,A.,  
299 Blessinger, T., Dishaw, L., Hotchkiss, A., Makris, S. 2019. Hazards of diisobutyl phthalate  
300 (DIBP) exposure: A systematic review of animal toxicology studies. Environment  
301 International 125, 579–594.
- 302 Zota AZ, Calafat AM, Woodruff TJ. 2014. Temporal trends in phthalate exposures:  
303 findings from the National Health and Nutrition Examination Survey, 2001–2010.  
304 Environ Health Perspect 122:235–241; <http://dx.doi.org/10.1289/ehp.1306681>