

1 Giacomo Assandri<sup>a,b,\*</sup>, Andrea Bernardi<sup>b</sup>, Arianna Schmoliner<sup>b</sup>, Giuseppe Bogliani<sup>a</sup>, Paolo Pedrini<sup>b</sup>, Mattia  
2 Brambilla<sup>b,c</sup>

3  
4 **A matter of pipes: Wryneck *Jynx torquilla* habitat selection and breeding performance in  
5 an intensive agroecosystem**

6  
7  
8 <sup>a</sup> University of Pavia, Dept. of Earth and Environmental Sciences, Via Adolfo Ferrata 9, I-27100, Pavia, Italy.

9 <sup>b</sup> MUSE. Sezione Zoologia dei Vertebrati, Corso del Lavoro e della Scienza 3, I-38123, Trento, Italy.

10 <sup>c</sup> Fondazione Lombardia per l'Ambiente, Settore biodiversità e aree protette, Largo 10 luglio 1976 1, 20822,  
11 Seveso (MB), Italy.

12  
13 \* Corresponding author at: University of Pavia, Dept. of Earth and Environmental Sciences, Via Adolfo  
14 Ferrata 9, I-27100, Pavia, Italy. e-mail: giacomo.assandri@gmail.com. telephone: +39 333 9244524.

15  
16  
17  
18 **Acknowledgements** - We are grateful to Unità Viticoltura (CTT) of Fondazione 'Edmund Mach' (and in  
19 particular to M. Venturelli, C. Ioriatti, M. Bottura, F. Ghidoni e F. Penner) for kind cooperation. A. Iemma  
20 and G. Ranghetti helped with technical issue. Parco Naturale Adamello-Brenta provided field facilities. K.  
21 Horwat revised the English.

22 **Abstract** - In intensive permanent crops, the declining Wryneck had been reported to be favoured by nest-  
23 boxes contrasting the lack of suitable breeding sites; however, it occurs in very intensive vineyard,  
24 apparently deprived of suitable nesting sites. Considering the wryneck concerning conservation status and  
25 the increasing vineyard share and intensification over Europe, investigating habitat selection and use of  
26 nesting sites in intensive vineyards is urgently required.

27 With this aim, we investigated for the first time territory-scale habitat selection in a Wryneck population  
28 inhabiting a vineyard-dominated landscape (in NE Italy) without conservation measures addressed to the  
29 species and provided a preliminary assessment of its breeding biology.

30 We investigated the effect of land-cover, management, and nest-site availability on Wryneck occurrence  
31 considering 44 territories and an equal number of control plots. In the subsequent year, Wryneck nests  
32 were surveyed in a subset of simple landscapes (<20% of semi-natural habitats). All nest boxes and holes  
33 provided by pipe beams supporting traditional *pergola* vineyards were checked twice during the breeding  
34 season.

35 According to the territory model, Wrynecks did not select particular habitat types, but set territories in  
36 sunny areas in which *pergola* vineyards were more abundant than *spalliera* ones and had a higher  
37 availability of pipe holes and in simpler landscapes.

38 Breeding attempts in pipes were few and all failed during egg deposition; conversely, 39% of the few  
39 available nest boxes were occupied with an overall breeding success of 57%. 36% of the eggs laid resulted  
40 in fledged juveniles.

41 Although pipes provide potential nesting sites and positively affect territory settlement, they finally  
42 resulted in breeding failure, potentially even exacerbating the impact of intensive agricultural management  
43 on the species (i.e. a pattern recalling an ecological trap). Nest boxes may supply safer breeding sites for  
44 Wrynecks; however, a wider assessment of the reproductive outcomes following nest-box supplementation  
45 should be planned.

46

47

48

49 **Keywords** – bare ground; cavity breeder; habitat selection; nest boxes; territory; vineyards.

50

51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65  
66  
67  
68  
69  
70  
71  
72  
73  
74  
75  
76  
77  
78  
79  
80  
81  
82  
83  
84  
85  
86  
87  
88  
89  
90  
91  
92  
93  
94  
95  
96  
97  
98  
99  
100  
101  
102  
103

## 1. Introduction

In Europe, agricultural systems harbour nearly 120 species of birds, which adapted to this habitat during the course of the millenarian development of agriculture (Tucker and Heath 1994; Krebs et al. 1999).

The European wild bird indicator shows that farmland bird populations have halved across the continent since 1980, while the populations of other bird species were generally stable or increasing (Fuller et al. 1995; Gregory et al. 2005; <http://www.ebcc.info/>). Agricultural intensification, which is likely the primary cause of direct negative effects on farmland bird populations (Newton 2004; Donald et al. 2006), impacts birds and biodiversity at two distinct but interconnected spatial scales: the landscape and the local scales. At the landscape scale, it causes the removal of marginal habitats, homogenization, and fragmentation through the conversion of grassland-like habitats into crops, increases in field size, and ultimately resulting in highly simplified landscapes. At the field scale, intensification mainly means increasing the intensity of farming practices (e.g. greater fertilizer/pesticide inputs, deep ploughing, massive use of machinery, large scale irrigation) (Tscharrntke et al. 2005; Fahrig et al. 2011). Attempts to reverse the negative trend of farmland birds in Europe mostly failed. A recent assessment showed that avian abundance and biomass are both still declining with most of this decline attributed to common avian species, in particular farmland species (Inger et al. 2014). In the years to come, farming will likely represent the single greatest source of threat to birds in Europe and worldwide (Green et al. 2005; BirdLife International 2015).

The Wryneck *Jynx torquilla* is one the farmland species that underwent the largest decline throughout Europe in the last decades (Birdlife International 2004; Sanderson et al. 2006), with an estimated decline rate of -57% in the period from 1980-2013 (<http://www.ebcc.info/trends2015.html>). Although its population remained nearly stable in the last decade (2004-2013; <http://www.ebcc.info/trends2015.html>), the species has continued to decline in several countries, including Italy (Rete Rurale Nazionale and LIPU 2015). Wryneck breeds in sunny and dry semi-open environments, in particular in extensively managed farmland; however, it is not a farmland specialist, as it exploits a variety of other habitats including open woodland, forest margins, copses, parks, and gardens, provided that two basic resources occur: cavities for breeding, since the species is a secondary cavity breeder, and ground-dwelling ants (Formicidae), as it feeds almost exclusively on those insects (Cramp 1985; Gorman 2004).

The causes for its decline are not fully understood. Habitat loss has been reported as causing the greatest impact (Gorman 2004). In farmland, agricultural intensification has been the main cause of Wryneck decline, as it leads to a reduction of nesting sites and perches used for hunting, in particular large isolated trees (Coudrain et al. 2010; Roux et al. 2015). Similarly, high-stem orchards, which used to be a preferential habitat, have disappeared in large parts of Europe (Herzog 1998; Kizos et al. 2012). In parallel, the same pattern of intensification has caused a reduction in prey availability: on one hand, agricultural intensification determines the reduction of ant abundance in the agroecosystem (by e.g. intensive use of insecticides), whereas on the other hand, the denser cover of ground vegetation due to the intensive use of fertilizers negatively affects ant detection by the species (Mermod et al. 2009). Similarly, wryneck declined in woodlands and forests. The causes of the decline in those habitat have been identified in the intensification of woodland management (which reduces nest sites), and in the thickening of forest undergrowth due to air-borne nitrogen deposition, which reduces forest openings and then impacts on the availability and accessibility of preys (Gorman, 2004; Mikusinski & Angelstam, 1997). Additionally, as is the case for other species wintering in sub-Saharan Africa, climatic and land use changes in western Sahel might have impacted the species via reduced survival (Sanderson et al. 2006; Zwarts et al. 2009). However, recent evidence suggests that Wrynecks breeding in central Europe (and probably, as a consequence of leap-frog migration, also in the Mediterranean region) mainly winter in the Mediterranean basin (van Wijk et al. 2013); thus factors acting on Palearctic breeding grounds are likely to be crucial.

A recent study carried out in Switzerland showed that Wrynecks could also persist in intensive permanent crops (orchards and vineyards), where a heterogeneous habitat matrix with some bare ground patches occurs, as long as ants and nest sites are available (Mermod et al. 2009). This study was carried out in an area where hundreds of nest boxes were installed for conservation purposes (Arlettaz et al. 2010; Zingg et

104 al. 2010), but the species had also been found to occur in abundance in intensive vineyards elsewhere  
105 (Coudrain et al., 2010; Assandri et al. 2016). However, how Wrynecks could find nesting sites in very  
106 intensive vineyards is still unknown, but this kind of knowledge is of key importance considering both the  
107 Europe-wide wryneck decline and the increasing share and intensification of vineyards over the continent.  
108 In this paper, we investigated for the first time habitat selection at the territory scale in a Wryneck  
109 population breeding in a vineyard-dominated landscape (without systematic nest-box supplementation) in  
110 north-eastern Italy in response to landscape structure and management practices. We collected  
111 information on the breeding biology of the species focusing on an area where the intensification of  
112 management practices and the conversion of natural habitats into vineyards have created largely  
113 homogeneous monocultures (Assandri 2017b), similar to other sites in the Mediterranean basin and nearby  
114 areas (Martínez-Casasnovas et al. 2010), where intensification determined harsh impacts on biodiversity  
115 (Viers et al., 2013). Considering the concerning decline of the species across its breeding range, further  
116 knowledge on its ecology is urgently needed. Our specific aims were thus i) to investigate the habitat  
117 factors driving habitat selection in vineyard-dwelling Wrynecks; it was predicted that non-vineyard habitats  
118 could be less important than the characteristics associated with vineyard management in the  
119 homogeneous landscape of the vineyard-dominated area considered; ii) to understand how this secondary  
120 cavity breeder could nest in this simple and intensively exploited landscape that is apparently deprived of  
121 suitable nesting sites (it was expected that Wrynecks have found surrogate cavities for breeding); iii) to  
122 evaluate its breeding success in this agroecosystem, and considering the intensive agriculture of the area, a  
123 reduced reproductive outcome was expected.

124

125

## 126 **2. Materials and methods**

127

### 128 **2.1 Study area**

129 The study was performed in the Trento province (South-eastern Alps, Italy), in an area characterized by  
130 partly urbanized valley bottoms intensively exploited for permanent crop cultivation and mountainsides  
131 covered by woodlands interspersed with pastures, apple orchards, and in particular, vineyards (up to 900 m  
132 a.s.l.). For further details on the study area see Assandri et al. (2016, 2017a). We focused on 3 macro-areas  
133 (Piana Rotaliana, Cembra Valley, and the eastern side of Adige Valley) with different environmental and  
134 topographic characteristics mirrored in different local agricultural practices. They displayed a strong  
135 contrast in land usage with highly intensive cultivation in the most accessible and flat areas and more  
136 extensive farming elsewhere.

137 Piana Rotaliana is an alluvial plain (200-230 m a.s.l.) with land use dominated by intensive vineyards (see  
138 Assandri et al. (2017b) for further details).

139 In Cembra Valley, vineyards are the dominant land use between 370 and 900 m a.s.l. in the southern-  
140 exposed valley side. High acclivity prevented widespread agricultural mechanization, and viticulture is  
141 associated with a massive system of terraces supported by stone walls (Agnoletti 2013).

142 The third macro-area is the eastern side of the Adige Valley from San Michele all'Adige to Trento, an  
143 exposed hilly area largely covered by vineyards (which are still expanding at the expense of the forest). This  
144 area displays an intermediate level of viticulture intensification between the two previous areas.

145 For each of the three macro-areas, eight sample sites were selected (totally accounting for 400 ha; mean  
146 sample site surface  $\pm$  sd:  $15.8 \pm 3.4$  ha; range: 10.8-22.8 ha; Figure 1), that were: i) representative of the  
147 environmental characteristics of the main area; ii) small enough to survey three of them in a morning; iii)  
148 easy to survey (i.e. without any parts hidden when walking along a pre-defined route).

149

### 150 **2.2 Territory survey**

#### 151 **2.2.1 Field mapping**

152 Territory mapping was conducted in each of the 24 sample sites during 4 visits during the breeding season  
153 in 2015 (10.04-17.04; 05.05-12.05; 29.05-05.06; 27.06-05.07).

154 During each visit, the same observer (G.A.) followed the same route inside the site, walking at a slow pace  
155 and thoroughly surveying the whole site.

156 All contacts with Wrynecks were recorded inside the sample sites as precisely as possible by using updated  
 157 aerial photographs (scale 1:2500 m) and starting from the first location. If a Wryneck moved spontaneously  
 158 (i.e. not disturbed by the observer), we also recorded its subsequent location(s) until it continued the same  
 159 activity for at least 10 minutes (e.g. feeding, staying in a hole, singing from the same perch, etc.).  
 160 Individuals' behaviours and interactions (e.g. aggressive behaviour, courtship, cavity showing, etc.) were  
 161 accurately recorded in order to ease the subsequent task of territory definition.  
 162 Three sites per day were censused from dawn to a maximum of six hours afterwards (5.30 – 11.30 a.m. in  
 163 spring), when Wryneck song activity is highest. Bad weather conditions (e.g. strong wind, rain) were  
 164 avoided.

165 Census order across sites was changed from one visit to the following, to ensure variability in the census  
 166 time within the morning.

167

### 168 **2.2.2 Definition of territories and control plots**

169 In studies dealing with resource selection by birds, several methods have been adopted to define territories  
 170 (Bibby 2000). Circular buffers (e.g. around the nest) are frequently assumed to represent the territory  
 171 defended by territorial species (e.g. Jedlikowski et al. 2016; Martinez et al. 2010).

172 Environmental variables were measured within 2.07 ha circular buffers (radius=81 m), defined by means of  
 173 a two-step procedure and representing territories defended by pairs.

174 We initially built the minimum convex polygon based on the locations attributed to the same Wryneck pair  
 175 and paying particular attention to simultaneous locations and interactions between individuals. The records  
 176 potentially attributable to migrant individuals (e.g. birds feeding in unsuitable habitats during the early part  
 177 of the study season and no longer contacted in the same site during subsequent visits) were discarded.

178 In the first step, the centroid of each polygon and the mean surface of the polygons that were based on  
 179 more than 3 points (N=12) were calculated. The latter analysis suggested an average territory size equal to  
 180 0.69 ha, corresponding to a circular plot with a radius 47 m. In cases in which a nest was found, this was  
 181 considered the final centre of the territory.

182 As a second step, we calculated the distance between the nearest neighbouring centroids (in the same  
 183 sample site) and divided it by two, obtaining a value of 115 m (N=36).

184 We finally averaged the two values obtained by the two-step procedure (radius of the circular plot  
 185 corresponding to the average polygon surface and half of the average distance between neighbouring  
 186 centroids), and assumed the result (81 m) as the final radius of a hypothetical mean Wryneck territory. We  
 187 used this distance to buffer all available centroids (N=44) and considered the resulting plots as territories in  
 188 the analyses.

189 In Switzerland, Coudrain et al. (2010) and Mermoud et al. (2009) used an average radius of 111 m based on  
 190 previous telemetry data on seven Wrynecks (Weisshaupt et al. 2011), whereas in other studies on  
 191 territorial species at a wider scale, reference values obtained from literature were applied (Martinez et al.  
 192 2010; Brambilla et al. 2010). Our method seemed to be a good trade-off between accuracy and the  
 193 inclusion of a wider area.

194 Environmental variables were measured in the 44 territories and in 44 control plots defined starting from  
 195 random points scattered within the 24 surveyed sites at locations where Wrynecks were never recorded.  
 196 Overlapping between different territories was limited to 4.44%. QGIS 2.14.2 (QGIS Development Team  
 197 2016) was used for all the spatial analyses.

198

### 199 **2.2.3 Environmental variables collection**

200 Land-cover variables were measured by means of an accurate photointerpretation of aerial photographs,  
 201 validated and updated in the field. We defined nine habitat categories: woods, apple orchards, urban areas,  
 202 hedge and tree rows, paved roads, open areas, fallow land, vineyards, and field margins and calculated  
 203 their percentage of cover for each territory. Land-cover types related to agricultural uses were measured at  
 204 the parcel scale (field with the same crop, spatial arrangement, and management characteristics, see also  
 205 Coudrain et al. (2010)), and within each vineyard parcel, management variables were collected by means of  
 206 targeted field surveys.

207 In particular, we distinguished the two vineyard trellising systems occurring in the study area: *spalliera* and  
 208 *pergola*. *Spalliera* (espalier) is the globally widespread vineyard arrangement, in which low vines (generally

209 less than 2 m) are supported by wires held between poles. Rows are generally spaced less than 2 m apart.  
 210 *Pergola* is the traditional and predominant form in the region, accounting for about 80% of the overall  
 211 vineyard surface in Trentino (Chemolli et al. 2007). It consists of tall vines (up to 2.5 m when considering  
 212 the secondary branches, growing in a dense leaf “roof”), aligned in rows spaced up to 5 m apart and  
 213 supported by a robust structure of poles and beams. Beams could be metal irrigation pipes of various  
 214 diameters (generally 7 cm) or made of wood (Figure 2). The first system is widespread in modern and  
 215 intensive vineyards, the latter in old and traditional ones. In *pergola* vineyards, some typologies of  
 216 mechanical activities (e.g. mechanical harvesting and pruning) are not possible, but there are no other  
 217 differences in terms of general management between the two types. We attributed each vineyard in the  
 218 study area to one of these two systems and used the percentage cover of *spalliera* among explanatory  
 219 variables.

220 In our study area, vineyard and apple orchard grounds are extensively covered by a dense grass sward with  
 221 the exception of the vine and tree base (a strip of about 1 m), where herbicides or mechanical grass  
 222 removal are applied (mean ground grass cover in vineyard/apple orchards of the study area: 91.5%, our  
 223 unpub. data). However, we distinguished between vineyards and apple orchards chemically  
 224 weeded/ploughed and with a full grass cover by evaluating each single parcel in the field. The percentage  
 225 of cover for chemically weeded/ploughed vineyards and orchards in each territory/control was used as an  
 226 explanatory variable.

227 For each territory, we measured the average area of the vineyard parcels included (totally or partially)  
 228 within the buffer, which is a proxy for intensive (larger fields) or extensive (smaller fields) agriculture.

229 Wrynecks are secondary cavity breeders that require holes for nesting. We quantified the availability of  
 230 potential nesting sites within each territory by counting: the number of isolated trees with a diameter at  
 231 breast height >20 cm (see Coudrain et al. 2010); the number of nest-boxes (occasionally supplied by  
 232 farmers in their fields); the number of holes provided by pipe beam ends in *pergola* vineyards. The latter  
 233 was measured, because there were anecdotal records of Wrynecks using these kinds of vineyard supports  
 234 for nesting. Since pipes are typical of intensive and modern vineyards (in traditional, non-intensive  
 235 cultivations, supports are often made of wood), we were interested in their effect in contrasting  
 236 landscapes, so we divided our 24 sample sites into two landscape typologies following Batary et al. (2011),  
 237 i.e. simple and complex landscapes. Landscapes with >20% of semi-natural areas (e.g. wood, hedgerows,  
 238 tree lines, fallow lands, meadows, field margins), that are typical of traditional agriculture were considered  
 239 as complex, whereas landscapes with <20% of those habitats were considered as simple.

240 Following that criterion, 10 sites out of 24 were classified as simple landscapes (mean cover of semi-  
 241 natural areas:  $12.02 \pm 5.80\%$ ), whereas 14 were complex (mean cover of semi-natural areas:  $29.91 \pm 5.82$   
 242 %), including all the sites on the eastern side of Adige Valley, except two (Fig. 1). In the simple landscapes,  
 243 viticulture is more intensive compared with complex landscapes where agriculture is conducted in a more  
 244 traditional and extensive way.

245 We additionally measured mean elevation and mean direct solar radiation for each territory from a 1-  
 246 m resolution digital elevation model (DEM). Solar radiation was calculated on the 21<sup>st</sup> of June using the  
 247 r.sun function from GRASS 7.0.2 and taking the shadowing effect of the topography into account (Neteler  
 248 et al. 2012). For further details on explanatory variables, see Table 1.

#### 250 **2.2.4 Statistical analysis**

251 Our explanatory variables were divided into three groups, which were separately considered when building  
 252 models: land-cover/topographic, management, and nest-site predictors (see Table 1). This approach was  
 253 adopted in similar studies when the variables’ nature differed a lot and information embedded in different  
 254 sets of variables partially overlapped (see for example Coudrain et al., 2010, Assandri et al., 2017c).

255 We carried out an accurate data exploration for each group of predictors in order to avoid common  
 256 statistical problems (e.g. collinearity), following Zuur et al. (2010).

257 Vineyard cover was included in the management group (instead of land-cover) to correct for their cover  
 258 within territories/control for plots when evaluating the effect of the vineyard management variables, as  
 259 well as to remove collinearity among covariates in the land-cover group.

260 All the explanatory variables were standardized before entering them into the models to allow  
 261 comparisons of their relative effects (Schielzeth 2010), and since recent literature has highlighted the

262 importance of this procedure to control for multicollinearity in model averaging in order to obtain reliable  
 263 predictor estimates (Cade 2015).

264 To test the effect of the covariates on Wryneck occurrence probability, we used GLMMs with a binomial  
 265 error distribution and a logit-link function. Mixed models were used due to the nested nature of our  
 266 experimental design (i.e. to account for potential non-independence of territories within the same sample  
 267 sites). The response variable was Wryneck occurrence, and explanatory variables were the ones included in  
 268 the three groups mentioned above (land-cover/topographic, management, and nest-site) and the random  
 269 intercept was the sample site. In the nest-site model, we included an interaction between pipes and  
 270 landscape typology (simple vs. complex). GLMMs were run with the R package glmmADMB (Skaug et al.  
 271 2015).

272 We worked within an information-theoretic framework (Burnham and Anderson 2002) using the dredge  
 273 function in the R package ‘MuMIn’ (Barton 2015) to build all the possible models for each set of  
 274 explanatory variables separately.

275 Models were ranked based on their AICc, and we selected only the most parsimonious models (i.e.  $\Delta AICc <$   
 276  $2$ ). We then averaged across these most supported models within each group of predictors to obtain  
 277 model-averaged coefficients, their relative standard errors, and the relative variable importance (Johnson  
 278 and Omland 2004) for each explanatory variable. In that process, ‘uninformative parameters’ (Arnold  
 279 2010), i.e. the variables included only in models comprising more parsimonious nested models, were  
 280 discarded (Richards 2008; Richards et al. 2011).

281 Finally, we built a synthetic model starting with the variables selected according to the above procedure  
 282 for each individual group, and adopting the same AICc-based ranking and model-averaging procedure  
 283 (Koleček et al. 2014). All the analyses were performed with R version 3.2.0 (R Core Team 2016).

284

### 285 **2.3 Nest survey**

286 During the breeding season of 2016, we searched for Wryneck nests within the eight sample sites of Piana  
 287 Rotaliana (all classified as simple landscapes). Territories were mapped in the same way as in 2015, but  
 288 each sample site was visited seven times between 22 March and 13 May. We then created an 81-m buffer  
 289 around each Wryneck contact location (see above), defining an area of “potential nest presence”, and then  
 290 carried out two sessions of nest-searching, one for the first brood (25 May – 10 June) and one for the  
 291 second (9-27 July), since the Wryneck is generally reported to be a double brooded species (Gorman, 2004).  
 292 Nests were searched for in the entire vineyard parcels totally or partially overlapping the eight sample site  
 293 areas and with the “potential nest presence area” using an “Explorer Premium” digital endoscope, which  
 294 allowed us to survey all the pipe holes and the nest-boxes occurring in this area. An area of 272 ha and  
 295 approximately 3,000 pipes and 18 nest-boxes occurring there were surveyed. This area is almost completely  
 296 covered by vineyards (92.3% of the land cover with the remaining part mainly constituted by roads, field  
 297 margins, and apple orchards), thus all the nests occurring within the area were virtually found.

298 After finding a nest, it was checked periodically to assess breeding success and collect reproductive  
 299 parameters (i.e. number of eggs and the relative status - intact or damaged, number of nestlings, and the  
 300 relative status -alive or dead).

301 To evaluate the environmental conditions occurring in pipes, four “Ibutton” data loggers were placed at  
 302 pipe entrances and pipes were selected close to pipes with Wryneck nests and sharing the same  
 303 characteristics with the latter (same conditions of solar exposition, same colour and material, same hole  
 304 diameter). The loggers recorded temperature every minute between the 10<sup>th</sup> May and the end of June  
 305 (thus during the period of Wrynecks’ first brood, from nest site selection to chick rearing). The minimum  
 306 and maximum pipe temperatures were compared with the air temperatures in the same period using  
 307 values measured at a meteorological station found in the study area (Maso delle Part, Mezzolombardo,  
 308 TN).

309

310

## 311 **3 Results**

312

### 313 **3.1 Territory survey**

314 The most parsimonious among land-cover/topographic models only retained the topographic variables (i.e.  
 315 solar radiation and elevation), whereas all the land-cover categories were excluded as not comprised in the  
 316 supported models ( $\Delta AICc < 2$ ), or being uninformative (see supplementary materials online, Table S1-S2).  
 317 Considering management variables, the cover of *spalliera* vineyard and the mean area of vineyard parcels  
 318 were retained, whereas chemically weeded or ploughed field cover and vineyard cover were not (Table S3-  
 319 S4).

320 Among the potential nest-site predictors, only the interaction between pipes and landscape typology was  
 321 retained, whereas isolated trees and nest boxes were discarded (Table S5-S6).

322 Among the individual groups, the land-cover/topographic best model ( $AICc = 114.6$ ) and the nest-sites  
 323 model ( $AICc = 114.3$ ) had fully comparable support, whereas the management one was less supported  
 324 ( $AICc = 118.6$ ). The synthetic model, combining variables from different groups, was slightly more supported  
 325 ( $AICc = 113$ ).

326 Despite a moderate collinearity among the retained predictors (higher  $gVIF = 4.18$ ), all of them were entered  
 327 in the final synthetic model. No odd effect imputable to collinearity was detected in the model output, and  
 328 parameter estimates were biologically meaningful and comparable with individual group model outputs, so  
 329 the statistical issue of possible collinearity was considered of minor concern in this case.

330 Support was found for nine synthetic models, and all the variables were informative with the exception of  
 331 the mean area of vineyard parcels (Table 2).

332 Overall, without considering landscape typology in interaction with other covariates, Wryneck had a higher  
 333 probability of occurrence in simple landscapes than in complex ones (Table 2).

334 Solar radiation had a positive effect on Wryneck occurrence, whereas *spalliera* vineyard cover and  
 335 elevation had negative effects, the latter being barely uninfluential, as suggested by the parameter  
 336 estimate being very close to zero (Table 2).

337 The interaction between pipes and landscape types was well supported (retained in the most parsimonious  
 338 model) and suggested a strong positive effect of pipes, especially in simple landscapes (Table 2, Figure 3).

339

### 340 **3.2 Nest survey**

341 In Piana Rotaliana during spring 2016, we defined 17 territories defended by Wryneck pairs. In 2015, the  
 342 same area hosted 25 territories. We found 11 clutches (plus a replacement clutch) during the first brood  
 343 period (6 in pipes (Figure S1) and 5 in nest boxes), and 2 clutches during the second, both in nest boxes.  
 344 Only 4 of these 11 clutches occurred inside one of the 17 defended territories.

345 All the clutches in pipes were abandoned, 4 after the deposition of the first egg, one with 3 eggs and one  
 346 with 6. In nest boxes, 3 clutches failed (due to abandonment or predation), 3 were partially successful (at  
 347 least 13 juveniles fledged over 37 eggs laid). One containing 6 eggs was predated, and subsequently the  
 348 (same?) female laid a further 4 eggs which hatched and the nestlings fledged.

349 The mean temperature in pipes, as measured by the data loggers, was  $22.0^{\circ} C$ , the overall mean of the  
 350 minimum temperature recorded was  $6.6^{\circ} C$ , and the mean of the maximum was  $43.1^{\circ} C$  ( $N = 4$ ).  
 351 Temperature measurements above  $40^{\circ} C$  accounted for the 6.7% of the sample. For comparison, the air  
 352 temperature in the same area and in the same period registered a minimum of  $3.4^{\circ} C$  and a maximum of  
 353  $33.4^{\circ} C$ .

354

355

## 356 **4. Discussion**

357

358 In our vineyard-dominated study system, Wrynecks seemed not to select particular habitat features, but  
 359 set territories in sunny areas with *pergola* vineyards with a high availability of pipe holes (i.e. potential  
 360 nesting sites) and simple landscapes, i.e. landscapes with natural remnants covering less than 20% of the  
 361 area (Batáry et al. 2011).

362 In Trentino vineyards, Wryneck abundance is positively affected by vineyard cover at a landscape scale (our  
 363 unpub. data), possibly because vineyards recall the natural habitat to which Wryneck is generally  
 364 associated with, i.e. sunny open forests in warm climates (Cramp 1985; Gorman 2004). Our results showed  
 365 that at a finer spatial scale (i.e. territory), specific vineyard characteristics likely have a greater importance  
 366 than land cover. In particular, the availability of nesting sites, which act as a limiting resource in this



367 intensive crop, seemed particularly important. Consistent results came from other studies performed in  
 368 intensive farmlands (orchards and vineyards), which showed that territory establishment and reproductive  
 369 output in Wrynecks are driven by specific limiting factors, such as food (in particular ant) availability and  
 370 detectability (critically affected by the occurrence of bare ground) and, in particular, nesting site  
 371 availability, rather than by habitat composition (Mermod et al. 2009; Zingg et al. 2010; Coudrain et al. 2010;  
 372 Weisshaupt et al. 2011).

373 The number of holes available in a territory had been reported to be correlated with territory quality in  
 374 Wrynecks and other cavity nesting species, since several holes allow birds to tune nest choice according to  
 375 changing environmental conditions or to predation pressure (Tomé et al. 2004; Zingg et al. 2010; Coudrain  
 376 et al. 2010).

377 In our study system, Wrynecks seemed to be attracted by pipe holes and defended territories with a higher  
 378 availability of these artificial structures. In fact, our extensive nest search showed that the effective  
 379 breeding attempts in pipes were very few, especially when compared with their extremely high availability  
 380 and all attempts occurred during the first part of the breeding season (i.e. the time of the first brood). All  
 381 the breeding attempts occurring in pipes failed during egg deposition, suggesting pipes were unsuitable for  
 382 reproduction. One of the possible causes for the widespread failure recorded in pipes could be the high  
 383 maximum temperature reached within them, approximately 10°C higher than outside. Additionally, a metal  
 384 pipe with a round section of 7 cm could have offered inadequate conditions for effective egg incubation.  
 385 We acknowledge our low sample size, and thus our results should be considered preliminary findings.  
 386 Nevertheless, the low number of nests in pipes despite the high number of pipe holes available and  
 387 surveyed clearly suggested that nesting in pipes is relatively rare.

388 In short, Wrynecks selected territories on the basis of the number of pipe holes, but then apparently did  
 389 not find suitable nesting sites or were forced to nest in unfavourable sites (i.e. in pipes). This behaviour  
 390 could be possibly due to juvenile/inexperienced individuals with the more experienced (or the dominant)  
 391 individuals ultimately exploiting the few favourable available nesting sites (e.g. nest boxes), and in general  
 392 this pattern was likely to be due to the general scarcity of potential breeding sites.

393 Even if the number of nest boxes was not selected among the informative predictors of Wryneck  
 394 occurrences, the nest survey revealed their key importance as breeding sites in simple landscapes with 39%  
 395 of boxes occupied. This is a high occupation rate especially considering that, in several cases, nest-boxes  
 396 occur in small clusters and that some of them were damaged or very small. Thus, several nest boxes were  
 397 likely unsuitable for the species. Additionally, nest boxes harboured the only four pairs that successfully  
 398 reared juveniles in the whole study area. The lack of effect in the analysis of territory selection could be an  
 399 effect of the clustered distribution and the very low density at which they occurred in the study area (0.055  
 400 nest-boxes/ha in 2015; 0.066 nest-boxes/ha in 2016).

401 We also found no effect of the number of isolated trees at the territory scale, even if they had been  
 402 reported to favour Wryneck territory occupancy (Coudrain et al. 2010) and farmers reported that big willow  
 403 trees (traditionally kept and pruned to produce cords to tie the vines, but now almost completely removed)  
 404 used to be occupied by Wrynecks for nesting in the study area. Also for isolated trees, the low rate at which  
 405 they occurred (0.145 trees/ha) could have led to an apparent lack of effect in the model.

406 Pipe hole distribution likely also explains why the species had both a higher probability of occurrence in  
 407 *pergola* than in *spalliera* vineyards (because the latter do not have beams with holes) and in simpler  
 408 landscapes than in complex ones. In fact, complex landscapes are characterized by more traditional and  
 409 less intensive viticulture, and vineyards are usually supported by wood beams without any (or a few) holes;  
 410 thus, the few Wrynecks in these vineyards likely depend on other cavities for nesting.

411 Wrynecks have been reported to be dependent on bare ground, since it favours ant detectability. In  
 412 previous studies in permanent agroecosystems, the availability of bare ground was a crucial predictor of  
 413 Wryneck occurrences at the foraging scale with higher occurrences of foraging Wrynecks when the extent  
 414 of bare ground was above 60% at the foraging site (Weisshaupt et al. 2011). Other studies performed at a  
 415 territory scale, reported less clear relationships between Wryneck territory occupancy and bare ground  
 416 availability, suggesting that at that scale bare ground availability might be not that crucial. Specifically,  
 417 Mermod et al. (2009) detected only a marginal effect of the extent of bare ground (with an optimum at 20-  
 418 30%), whereas Coudrain et al. (2010) found a stronger effect (with an optimum at 50%). In our study area,  
 419 the ground in vineyards and apple orchards is almost completely covered by grass with a percentage of

420 bare ground almost invariably much lower than the optimum reported in the above cited territory-scale  
 421 studies. In fact, no more than 10-15% of bare ground occurs in vineyards/orchards in the study area and  
 422 only in fields in which some herbicides or mechanical grass removal were applied. We still expected to find  
 423 a relatively higher amount of fields with bare ground in Wryneck territories, but models suggested a lack of  
 424 a noticeable effect. Although the methodology used to assess the bare ground extent in our study differed  
 425 from the one adopted in the cited Swiss studies, apparently, in our study system the crucial drivers of  
 426 territory selection are others (e.g. nesting site availability) than bare ground availability. Further studies in  
 427 the area should assess the fine-scale foraging habits of the Wryneck to understand how this species could  
 428 find prey and thus persist in this habitat, which presents a much lower extent of bare ground than the  
 429 optimum reported by other studies in similar habitats (Mermod et al. 2009, Coudrain et al. 2010). In  
 430 particular, factors affecting prey (in particular ant) detectability and abundance should be investigated to  
 431 further shed light on the species ecology in this agroecosystem.

432 In conclusion, the territory setting for Wryneck in the study area was strongly dependent on a vineyard  
 433 feature (pipe availability) characterising a simple (and intensive) landscape, which possibly provides nesting  
 434 sites. In fact, this feature was unsuitable for reproduction and the species bred successfully only in the few  
 435 nest boxes that sparsely occurred in the study area.

436 In areas characterized by intensive agriculture and in the absence of specific conservation measures (e.g.  
 437 nest box provisioning, Zingg et al. 2010), such as the one investigated here, the persistence of a viable  
 438 Wryneck population is likely to be strongly challenged. Our results suggested that particular structural  
 439 features of the farmland, which provide potential nesting sites (e.g. pipes) attracted individuals, but turned  
 440 out to be unsuitable for reproduction, and could potentially even exacerbate the impact of intensive  
 441 management on the species, possibly leading to an ecological trap.

442 Results suggested that the primary limiting factor for Wryneck was nest-site availability, thus the prime  
 443 conservation measure for its conservation should be the provisioning of nest boxes.

444 Nest box provisioning is a popular conservation measure to enhance secondary cavity nesting birds when  
 445 natural cavities are lacking, e.g. due to the removal of large trees in response to agricultural intensification  
 446 (Newton 1994a; Newton 1994b). However, when the general environmental conditions are not suitable for  
 447 a species, providing nest boxes may create ecological traps (Mänd et al. 2005; Klein et al. 2007). Hence, this  
 448 conservation measure should be tested on defined sample areas and followed by an assessment of the  
 449 reproductive outcome in those areas. Furthermore, nest boxes should be provided in an adequate number  
 450 to allow Wrynecks to choose the boxes in the most suitable territories.

451

## 452 **References**

- 453 Agnoletti M (2013) Italian Historical Rural Landscapes. Springer  
 454 Arlettaz R, Schaub M, Jérôme F, et al (2010) From publications to public actions: when conservation  
 455 biologist bridge the gap between research and implementation. *Bioscience* 60:835–842.  
 456 Arnold TW (2010) Uninformative parameters and model selection using Akaike's information criterion. *J*  
 457 *Wildl Manage* 74:1175–1178. doi: 10.2193/2009-367  
 458 Assandri G, Bogliani G, Pedrini P, Brambilla M (2016) Diversity in the monotony? Habitat traits and  
 459 management practices shape avian communities in intensive vineyards. *Agric Ecosyst Environ*  
 460 223:250–260. doi: 10.1016/j.agee.2016.03.014  
 461 Assandri G, Bogliani G, Pedrini P, Brambilla M (2017a) Assessing common birds' ecological requirements to  
 462 address nature conservation in permanent crops: Lessons from Italian vineyards. *J Environ Manage*  
 463 191:145–154. doi: 10.1016/j.jenvman.2016.12.071  
 464 Assandri G, Giacomazzo M, Brambilla M, et al (2017b) Nest density, nest-site selection, and breeding  
 465 success of birds in vineyards: Management implication for conservation in a highly intensive farming  
 466 system. *Biol Conserv* 205:23–33. doi: <http://dx.doi.org/10.1016/j.biocon.2016.11.020>  
 467 Assandri G, Bogliani G, Pedrini P, Brambilla M (2017c). Insectivorous birds as 'non-traditional' flagship  
 468 species in vineyards: Applying a neglected conservation paradigm to agricultural systems. *Ecological*  
 469 *indicators*, in press. doi: 10.1016/j.ecolind.2017.05.012  
 470 Barton C (2015) MuMIn: Multi-Model Inference. R package version 1.13.4.  
 471 Batáry P, Báldi A, Kleijn D, Tscharntke T (2011) Landscape-moderated biodiversity effects of agri-  
 472 environmental management: a meta-analysis. *Proc R Soc B Biol Sci* 278:1894–1902. doi:

- 473 10.1098/rspb.2010.1923
- 474 Bibby CJ (2000) Bird census techniques. Elsevier
- 475 BirdLife International (2015) European Red List of Birds.
- 476 Birdlife International (2004) Birds in Europe: population estimates, trends and conservation status. BirdLife  
477 International, Cambridge, UK
- 478 Brambilla M, Casale F, Bergero V, et al (2010) Glorious past, uncertain present, bad future? Assessing  
479 effects of land-use changes on habitat suitability for a threatened farmland bird species. *Biol Conserv*  
480 143:2770–2778. doi: 10.1016/j.biocon.2010.07.025
- 481 Brambilla M, Pedrini P (2011) Intra-seasonal changes in local pattern of Corncrake *Crex crex* occurrence  
482 require adaptive conservation strategies in Alpine meadows. *Bird Conserv Int* 21:388–393.
- 483 Brambilla M, Rubolini D (2009) Intra-seasonal changes in distribution and habitat associations of a multi-  
484 brooded bird species: implications for conservation planning. *Anim Conserv* 12:71–77. doi:  
485 10.1111/j.1469-1795.2008.00226.x
- 486 Burnham KP, Anderson DR (2002) Model Selection and Multimodel Inference: A Practical Information-  
487 Theoretic Approach. Springer Science & Business Media
- 488 Cade BS (2015) Model averaging and muddled multimodel inferences. *Ecology* 96:2370–2382. doi:  
489 10.1890/14-1639.1
- 490 Chemolli M, Rizzo M, Bona E, Tonon C (2007) Vigneti e aziende viticole. *Terra Trent* 4:12–18.
- 491 Coudrain V, Arlettaz R, Schaub M (2010) Food or nesting place? Identifying factors limiting Wryneck  
492 populations. *J Ornithol* 151:867–880. doi: 10.1007/s10336-010-0525-9
- 493 Cramp S (1985) The Handbook of Birds of Europe, the Middle East and North Africa. Vol. IV. Oxford  
494 University Press, New York
- 495 Donald PF, Sanderson FJ, Burfield IJ, van Bommel FPJ (2006) Further evidence of continent-wide impacts of  
496 agricultural intensification on European farmland birds, 1990–2000. *Agric Ecosyst Environ* 116:189–  
497 196. doi: 10.1016/j.agee.2006.02.007
- 498 Fahrig L, Baudry J, Brotons L, et al (2011) Functional landscape heterogeneity and animal biodiversity in  
499 agricultural landscapes. *Ecol Lett* 14:101–112. doi: 10.1111/j.1461-0248.2010.01559.x
- 500 Fuller RJ, Gregory RD, Gibbons DW, et al (1995) Population declines and range contractions among lowland  
501 farmland birds in Britain. *Conserv Biol* 9:1425–1441. doi: 10.1046/j.1523-1739.1995.09061425.x
- 502 Gorman G (2004) Woodpeckers of Europe. Bruce Coleman, Chalfont St Peter, UK
- 503 Green RE, Cornell SJ, Scharlemann JPW, Balmford A (2005) Farming and the fate of wild nature. *Science* (80-  
504 ) 307:550–555. doi: 10.1126/science.1106049
- 505 Gregory RD, van Strien A, Vorisek P, et al (2005) Developing indicators for European birds. *Philos Trans R*  
506 *Soc Lond B Biol Sci* 360:269–88. doi: 10.1098/rstb.2004.1602
- 507 Herzog F (1998) Streuobst: A traditional agroforestry system as a model for agroforestry development in  
508 temperate Europe. *Agrofor Syst* 42:61–80. doi: 10.1023/A:1006152127824
- 509 Inger R, Gregory RD, Duffy JP, et al (2014) Common European birds are declining rapidly while less  
510 abundant species' numbers are rising. *Ecol Lett* 18:28–36. doi: 10.1111/ele.12387
- 511 Jedlikowski J, Chibowski P, Karasek T, Brambilla M (2016) Multi-scale habitat selection in highly territorial  
512 bird species: Exploring the contribution of nest, territory and landscape levels to site choice in  
513 breeding rallids (Aves: Rallidae). *Acta Oecologica* 73:10–20. doi: 10.1016/j.actao.2016.02.003
- 514 Johnson JB, Omland KS (2004) Model selection in ecology and evolution. *Trends Ecol Evol* 19:101–108. doi:  
515 10.1016/j.tree.2003.10.013
- 516 Kizos T, Plieninger T, Harald S, Petit C (2012) HNV permanent crops: olives, oaks, vines, fruit and nut trees.  
517 In: Oppermann R, Beafoy G, Gwyn J (eds) High Nature Value Farming in Europe - 35 European  
518 Countries, Experiences and Perspectives. Verlag Regionalkultur, pp 70–84
- 519 Klein Á, Nagy T, Csörgő T, Mátics R (2007) Exterior nest-boxes may negatively affect Barn Owl *Tyto alba*  
520 survival: an ecological trap. *Bird Conserv Int* 17:263–271. doi: 10.1017/S0959270907000792
- 521 Koleček J, Schleuning M, Burfield IJ, et al (2014) Birds protected by national legislation show improved  
522 population trends in Eastern Europe. *Biol Conserv* 172:109–116. doi: 10.1016/j.biocon.2014.02.029
- 523 Krebs JR, Wilson JD, Bradbury RB, Siriwardena GM (1999) The second Silent Spring? *Nature* 400:611–612.  
524 doi: 10.1038/23127
- 525 Mänd R, Tilgar V, Löhmus A, Leivits A (2005) Providing nest boxes for hole-nesting birds - Does habitat

- 526 matter? *Biodivers Conserv* 14:1823–1840. doi: 10.1007/s10531-004-1039-7
- 527 Martínez N, Jenni L, Wyss E, Zbinden N (2010) Habitat structure versus food abundance: the importance of  
528 sparse vegetation for the common redstart *Phoenicurus phoenicurus*. *J Ornithol* 151:297–307. doi:  
529 10.1007/s10336-009-0455-6
- 530 Martínez-Casasnovas JA, Ramos MC, Cots-Folch R (2010) Influence of the EU CAP on terrain morphology  
531 and vineyard cultivation in the Priorat region of NE Spain. *Land use policy* 27:11–21. doi:  
532 10.1016/j.landusepol.2008.01.009
- 533 Mermod M, Reichlin TS, Arlettaz R, Schaub M (2009) The importance of ant-rich habitats for the persistence  
534 of the Wryneck *Jynx torquilla* on farmland. *Ibis* 151:731–742. doi: 10.1111/j.1474-919X.2009.00956.x
- 535 Mikusinski G, Angelstam P (1997) European woodpeckers and anthropogenic habitat change: a review.  
536 *Vogelwelt* 118: 277–283.
- 537
- 538 Neteler M, Bowman MH, Landa M, Metz M (2012) GRASS GIS: A multi-purpose open source GIS. *Environ*  
539 *Model Softw* 31:124–130. doi: 10.1016/j.envsoft.2011.11.014
- 540 Newton I (2004) The recent declines of farmland bird populations in Britain: An appraisal of causal factors  
541 and conservation actions. *Ibis* 146:579–600. doi: 10.1111/j.1474-919X.2004.00375.x
- 542 Newton I (1994a) The role of nest sites in limiting the numbers of hole-nesting birds: a review. *Biol Conserv*  
543 70:265–276.
- 544 Newton I (1994b) Experiments on the limitations of bird breeding densities: a review. *Ibis* 136:397–411.
- 545 QGIS Development Team (2016) QGIS Geographic Information System. Open Source Geospatial Foundation  
546 Project.
- 547 R Core Team (2016) R: A language and environment for statistical computing.
- 548 Rete Rurale Nazionale, LIPU (2015) Uccelli comuni in Italia. Aggiornamento degli andamenti di popolazione  
549 e del Farmland Bird Index per la Rete Rurale Nazionale dal 2000 al 2014.
- 550 Richards SA (2008) Dealing with overdispersed count data in applied ecology. *J Appl Ecol* 45:218–227. doi:  
551 10.1111/j.1365-2664.2007.01377.x
- 552 Richards SA, Whittingham MJ, Stephens PA (2011) Model selection and model averaging in behavioural  
553 ecology: The utility of the IT-AIC framework. *Behav Ecol Sociobiol* 65:77–89. doi: 10.1007/s00265-010-  
554 1035-8
- 555 Roux DS Le, Ikin K, Lindenmayer DB, et al (2015) Enriching small trees with artificial nest boxes cannot  
556 mimic the value of large trees for hollow-nesting birds. *Restor Ecol* 1–7. doi: 10.1111/rec.12303
- 557 Sanderson FJ, Donald PF, Pain DJ, et al (2006) Long-term population declines in Afro-Palaearctic migrant  
558 birds. *Biol Conserv* 131:93–105. doi: 10.1016/j.biocon.2006.02.008
- 559 Schielzeth H (2010) Simple means to improve the interpretability of regression coefficients. *Methods Ecol*  
560 *Evol* 1:103–113. doi: 10.1111/j.2041-210X.2010.00012.x
- 561 Skaug H, Fournier D, Bolker BM, et al (2015) Generalized Linear Mixed Models using “AD Model Builder.”
- 562 Tomé R, Bloise C, Korpimäki E (2004) Nest-site selection and nesting success of Little Owls (*Athene noctua*)  
563 in Mediterranean woodland and open habitats. *J Raptor Res* 38:35–46.
- 564 Tschardt T, Klein A-M, Kruess A, et al (2005) Landscape perspectives on agricultural intensification and  
565 biodiversity - ecosystem service management. *Ecol Lett* 8:857–874. doi: 10.1111/j.1461-  
566 0248.2005.00782.x
- 567 Tucker GM, Heath MF (1994) *Birds in Europe: their conservation status*. BirdLife International, Cambridge,  
568 UK
- 569 van Wijk RE, Schaub M, Tolkmitt D, et al (2013) Short-distance migration of Wrynecks *Jynx torquilla* from  
570 Central European populations. *Ibis (Lond 1859)* 155:886–890. doi: 10.1111/ibi.12083
- 571 Viers JH, Williams JN, Nicholas K a., et al (2013) Vinecology: pairing wine with nature. *Conserv Lett* 6:287–  
572 299. doi: 10.1111/conl.12011
- 573 Weisshaupt N, Arlettaz R, Reichlin TS, Tagmann-Loiset A, Schaub M (2011) Habitat selection by foraging  
574 Wrynecks *Jynx torquilla* during the breeding season: identifying. *Bird Study* 58:111–119.
- 575 Zingg S, Arlettaz R, Schaub M (2010) Nestbox design influences territory occupancy and reproduction in a  
576 declining, secondary cavity-breeding bird. *Ardea* 98:67–75. doi: 10.5253/078.098.0109
- 577 Zuur A, Ieno EN, Elphick CS (2010) A protocol for data exploration to avoid common statistical problems.

578           Methods Ecol Evol 1:3–14. doi: 10.1111/j.2041-210X.2009.00001.x  
579       Zwarts L, Bijlsma R, van der Kamp J, Wymenga E (2009) Living on the edge: wetlands and birds in a changing  
580           Sahel. KNNV Publishing, Zeist, the Netherlands  
581

582 *Figure captions*

583

584 **Fig.1** Location of the study area in Italy and the sample sites in the Trento province. The vineyard cover is shown in violet, and the  
585 24 sites (numbered from 1 to 24) in which Wryneck territory mapping was performed are in black. The three macro-areas described  
586 in the study area section are also shown. Landscape classification of the sites: 1-9, and 12: simple landscape; 10,11, and 13-24:  
587 complex landscape.

588

589 **Fig.2** Comparison of different beams found in *pergola* vineyards in Trentino. A. Metal beam generally found in simple landscapes  
590 made from an irrigation pipe (diameter 7 cm). In the inset: detail of the hole provided by this kind of beam. B. Wooden beam  
591 generally found in complex landscapes.

592

593 **Fig.3** Graphical representation of the effect of pipes on the probability of Wryneck occurrence in contrasting landscapes (complex  
594 and simple) as predicted by the averaged synthetic models. Other predictors included in the models are kept constant at their  
595 mean value. 95% confidence intervals of the mean are shown in light grey. Quite large 95% confidence intervals are due to the  
596 choice of conducting model averaging on mixed models and to the presence of several outliers..N=88  
597

598 **Table 1.** List of variables used in the analysis and their mean value and range in territories and control plots.  
599

Variable name	Description	TERRITORIES (Mean and range)	CONTROLS (Mean and range)
<b>Land-cover/topographic variables</b>			
Woods	% cover of woodlands (large majority of broadleaved woodlands).	2.75 (0-28.00) %	5.14 (0-35.78) %
Apple orchards	% cover of intensive apple orchards.	5.02 (0-37.77) %	3.40 (0-55.67) %
Urban areas	% cover of urban areas (including isolated houses).	1.58 (0-23.70) %	2.68 (0-32.02) %
Hedge and tree rows	% cover of hedge and tree rows, defined as linear clusters of shrubs and/or trees, which were less than 15-m wide, isolated in the farmed landscape or originating from woodland remains but clearly isolated from the main woodland area.	1.41 (0-8.91) %	2.25 (0-11.72) %
Paved roads	% cover of paved roads.	2.01 (0-6.02) %	2.84 (0-7.88) %
Open areas	% cover of open areas (fields, meadows, extirpated wood crops).	2.11 (0-17.58) %	1.52 (0-19.12) %
Fallow land	% cover of fallow land (also including abandoned vineyards invaded by shrubs).	1.43 (0-24.65) %	1.25 (0-12.27) %
Field margins	% cover of field margins (also including unpaved roads and small rural buildings).	9.58 (0.97- 23.08) %	12.03 (0.12- 27.82) %
Altitude	Mean territory altitude	318 (204-727) m	409.58 (205- 675) m
Solar radiation	Mean territory solar radiation on 21 <sup>th</sup> June	8741 (8123- 8968) W/m <sup>2</sup>	8546 (7584- 8973) W/m <sup>2</sup>
<b>Management variables</b>			
Vineyards	% cover of vineyards	73.60 (30.70- 95.46) %	68.10 (22.78- 94.25) %
<i>Spalliera</i>	% cover of <i>spalliera</i> vineyards	7.12 (0-96.33) %	22.47 (0-100) %
Chemically weeded or ploughed fields	% cover of permanent crop fields (vineyards and apple orchards) with chemically weeded or ploughed rows (i.e. fields with at least some bare ground).	63.42 (0-100) %	50.68 (0-100) %
Mean area of vineyard parcels	Mean area of vineyard patches overlapping with a territory.	4200 (1005- 8394) m <sup>2</sup>	3295 (544- 15966) m <sup>2</sup>
<b>Nest-site variables</b>			
Isolated trees	Number of isolated trees (diameter at breast height >20 cm) per territory.	0.25 (0-3)	0.34 (0-3)
Nest boxes	Number of nest boxes per territory.	0.11 (0-3)	0.16 ± 0.48 (0-2)
Pipes	Number of holes originated by <i>pergola</i> pipe beam ends per territory.	20.47 (0-46)	11.79 (0-26)
Landscape typology	Categorical. Two levels: complex and simple.	-	-

602 **Table 2.** Most supported synthetic GLMM models on Wryneck occurrence. Models are ranked according to Akaike's information  
 603 criterion corrected for small sample size (AICc) and only models within an interval of  $\Delta\text{AICc} < 2$  are shown. The difference in AICc  
 604 from the best-supported model ( $\Delta\text{AICc}$ ), Akaike's weights ( $w_i$ ), and -2 log-likelihood values (logLik) are also given. Negative (-) or  
 605 positive (+) relationships between predictors and Wryneck occurrences are shown. For variable acronyms, see Table 1. N=88.  
 606

Model	df	logLik	AICc	$\Delta\text{AICc}$	$w_i$
solar radiation (+) + pipes (-) *landscape typology	6	-	113	-	0.16
		49.996			
solar radiation (+) + pipes (+)	4	-		0.10	0.15
		52.325			
<i>spalliera</i> (-) + pipes (-)*landscape typology	6	-		0.24	0.14
		50.118			
solar radiation (+) + <i>spalliera</i> (-) + landscape typology	5	-		0.56	0.12
		51.427			
solar radiation (+) + landscape typology	4	-		0.88	0.10
		52.713			
solar radiation (+) + <i>spalliera</i> (-)	4	-		1.17	0.09
		52.860			
pipes (-) *landscape typology	5	-		1.27	0.09
		51.786			
elevation (-) + irradiation (+)	4	-		1.60	0.07
		53.073			
<i>spalliera</i> (-) + landscape typology	4	-		1.96	0.06
		53.254			

607

608



609 **Table 3.** Model averaged standardized parameter (based on models with  $\Delta AICc < 2$ ) and relative variable importance of predictors  
 610 (measured considering the sum of the Akaike weights over the most supported models in which that variable appears) from  
 611 synthetic models of Wryneck occurrence. Covariates are ranked according to cumulative weights. The reference value for  
 612 landscape typology and its interaction is “complex landscape”, thus “pipes” coefficient refers to complex landscapes and  
 613 “pipes\*landscape” is the correction for simple ones. For variable acronyms see Table 1. N=88.  
 614

Variable	$\beta$	SE	$\sum w_i$
intercept (reference value for complex landscape)	-	0.737	-
solar radiation	0.458	0.394	0.71
landscape typology (simple)	0.871	0.859	0.68
pipes	-	0.769	0.55
spalliera	-	0.329	0.42
pipes*landscape (simple)	0.683	1.031	0.39
elevation	-	0.544	0.07
	0.389		

615