

17 **Abstract**

18 Infestations by the salmon louse (*Lepeophtheirus salmonis* Krøyer) represents the major fish health
19 problem that the Atlantic salmon (*Salmo salar*) industry has to face. Sea lice infestation has a large
20 impact on the economy of fish farmers, which are looking for a cost-effective and environmentally
21 sustainable alternative to chemical or mechanical treatments to delouse fish. The biological control
22 of sea lice using the so-called cleaner fish has been individuated as a feasible delousing approach of
23 Atlantic salmons. In particular, in recent years the lumpfish (*Cyclopterus lumpus*) has been extensively
24 farmed to be used as a 'biological weapon' in salmon farming because of its effectiveness in delousing
25 also in harsh environmental conditions. However, the environmental impact of lumpfish farming is
26 still largely unknown. Thus, the present study aimed at assessing the potential environmental impact
27 of lumpfish production through a life cycle assessment (LCA) approach. Feed and electricity
28 consumption, both for 8 of the 18 evaluated midpoint indicators, are the main responsible of the
29 environmental load while for the Freshwater and Marine eutrophication about 90% of the impact is
30 related to the emission of nitrogen and phosphorous compounds by fishes. These data lay the
31 foundation for further, sustainable improvement of lumpfish farming.

32

33 **Keywords:** lumpfish, aquaculture, life cycle assessment

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35

36 1. Introduction

37 Farming of Atlantic salmon (*Salmo salar*) represents one of the most flourishing component of the
38 finfish aquaculture sector worldwide, supplying high-end markets and serving the demand for
39 capture fisheries products. Farmed Atlantic salmon has become a super commodity, as pointed out
40 by its year-round, worldwide availability, product consistency and high production volume (Eagle et
41 al., 2004; Naylor et al., 2005). However, the infestation caused by the salmon louse (*Lepeophtheirus*
42 *salmonis* Krøyer), a copepod ectoparasite that grazes on the skin and mucosal tissue of fish, causing
43 infections, osmotic stress and death (Johansen et al., 2011), represents the main issue that the
44 Atlantic salmon industry has to face (Imsland et al., 2014). Sea lice infestation has a notable impact
45 on the economy of Atlantic salmon farmers because of the costs behind the treatment procedures
46 to delouse fish, as well as the reduction of fish growth, the increase of feed waste and the decrease
47 of market quality of the final product (Powell et al., 2018). For instance, the estimated cost of sea lice
48 for Norwegian fish farmers only exceeds 150 m€/year (Bergheim, 2012).

49 The Atlantic salmon industry is struggling sea lice infestation relying on different medicinal
50 treatments, including the application of chemotherapeutic or bath treatments with hydrogen
51 peroxide and organophosphates or synthetic pyrethroids, as well as feeding fish with food medicated
52 with emamectin benzoate (Denholm et al., 2002). Although medicinal treatments are effective in
53 delousing the salmon ectoparasite, the continuous and frequent use of pyrethroids and emamectin
54 benzoate induced the development of resistance in sea lice (Igboeli et al., 2012), leading to a reduced
55 effectiveness of the treatment and 50% increased proportion of ineffective treatments from 2002 to
56 2006 (Lees et al., 2008). Alternatively, non-medicinal approaches such as sea lice skirts or traps,
57 snorkels, thermal treatments, flushers, lasers, bubble curtains were used, but many of them are still
58 in the development or investigational phases and results in delousing need to be confirmed.

59 To overcome these limitations, a cost-effective and environmentally sustainable alternative to
60 medicinal and non-medicinal treatments has been recently individuated and refers to the biological
61 control using the so-called 'cleaner fish'. The use of cleaner fish is particularly attractive because it
62 can reduce the use of chemical medications, be more cost-effective than other approaches and
63 potentially less stressful to farmed salmon (Liu and Vanhauwaer Bjelland 2014; Treasurer, 2013).
64 Labrid fish, mainly the ballan wrasse (*Labrus bergylta*) and the goldsinny wrasse (*Ctenolabrus*
65 *rupestris*) have been used to delouse Atlantic salmon in net pens for 30 years (Bjordal, 1991), because
66 they significantly reduce the prevalence of sea lice in farmed salmon (Treasurer, 2013). However,
67 the use of labrid fish has a substantial limitation because they experience winter dormancy and do
68 not feed at water temperature below 6 °C (Kelly et al., 2014), precluding their use as cleaner fish over
69 the winter (Treasurer, 2013). Thus, an alternative cleaner fish with active feeding behaviour at low
70 water temperatures has been identified in the lumpfish (*Cyclopterus lumpus*; Imsland et al.,
71 2014a,b,c; 2015a). In fact, the lumpfish continue feeding also at temperatures as low as 4 °C (Nytrø
72 et al., 2014), allowing delousing of salmon over the year. Moreover, the lumpfish can easily rear under
73 captivity and reach the appropriate size to be deployed in salmon farms in as little as 4 months, while
74 the ballan wrasse typically requires 1.5 years (Helland et al., 2014). For these reasons, the number of
75 cleaner fish used by the salmon farming industry has increased exponentially since 2008, and almost
76 26 million were used in 2015 in salmon farming in Norway alone. It has been estimated that 50 million
77 cleaner fish will be required by Atlantic salmon industry within 2020, the most of which will be
78 lumpfish. To satisfy this huge demand, commercial production of lumpfish has grown exponentially
79 in the last few years, so that 11.8 million juveniles were reared in Norway during 2015 (Norwegian
80 Directorate of Fisheries, 2015) and over than 20 million in 2016 (Nodland, 2016). Although the use of
81 lumpfish is considered a sustainable approach to reduce the environmental impact of Atlantic salmon
82 farming, to date information on the environmental impact of lumpfish farming is lacking.

83 Thus, the present study was aimed and investigating the environmental impact of one of the main
84 lumpfish farming facilities in Norway by using a life cycle assessment (LCA) approach. LCA is an ISO-
85 standardized biophysical accounting framework commonly used to compile an inventory of material
86 and energy inputs and outputs typical of all the stages of a product life cycle and to quantify its
87 contributions to a specified suite of resource use and emissions-related environmental impact
88 categories (Guinee et al., 2001).

89

90 **2. Materials and Methods**

91 ***2.1 Lumpfish farming***

92 A schematic representation of the lumpfish farming process is reported by Powell et al. (2018). In the
93 present study we relied on data from one of the main land based lumpfish farming in Norway,
94 operating in flow-through seawater system. The facility who has collaborated and supplied with data
95 for this study prefer to remain anonym for commercial reasons. The selected lumpfish production
96 plant produces more than 1 million lumpfish yearly, while the total market need is about 40-50 million
97 cleaner fish. Moreover, this plant has been one of the first to farm lumpfish as cleaner fish against
98 sea lice and have therefore established aquaculture practice that have been replicated in others
99 facility in the whole Norway. The water flow taken from deep water pass through sand filter, UV filter
100 and oxygenation before flowing into the fish tanks in order to optimize the water quality and promote
101 fish wellness and health. Moreover, the division of the farm were juvenile is produced utilize a heat
102 exchanger for increase the temperature of the water and increase the growth rate. Briefly, lumpfish
103 in the facility are farmed as follows. Wild-caught lumpfish are used as a broodstock to produce
104 juveniles to be used as cleaner fish in Atlantic salmon sea pen. Sexually-mature adults are typically
105 wild-caught during the spawning season using gill nets deployed in shallow waters (up to ~30 m deep)
106 close to the shore. In captivity, fertilization is performed through the 'dry method', that is mixing the

107 sperm with eggs and adding seawater to activate the sperm. Sperm is collected following dissection
108 of the testes, which are then macerated and passed through a sieve. Female abdomen is squeezed
109 to obtain eggs, which are transferred in small tanks to be fertilized by male sperm. Fertilized eggs are
110 quickly transferred in upwelling incubators consisting of 14 L hoppers loaded with an average 0.4 -
111 0.9 kg of eggs, corresponding to about 20,000 to 45,000 of eggs per hopper. Seawater flow rate is
112 maintained at 15 L/min during incubation, and then increased to 20 L/min when they became eyed-
113 eggs and the oxygen uptake need increase. Overall, hatching period lasts about 250 – 300 degree
114 days in the temperature range between 7 to 12 °C. Considering that the temperature in the hatching
115 division of the plant we considered is maintained at ~ 10 °C daily, the hatching of lumpfish eggs
116 requires ~ 30 days. After hatching, larvae are transferred to bigger tanks of 1 m³ dimension and feed
117 with dry food with granule size from 75 to 250 µm. Then, in the following 6 to 9 months the
118 development of the juvenile happens almost exponentially from less than 0.1 gram to an average of
119 30-35 grams. The fishes are split several times during the development phase and divided in 1 to 3
120 m³ tanks accordingly to the fish size and density need. During this growth phase the lumpfish are still
121 feed with dry food granulate with size range from 250 µm to 840 - 1,410 µm. The preferable water
122 temperature is maintained around 12 °C in order to optimize the growth rate. Almost four week
123 before releasing them in the sea cages with Atlantic salmon, the lumpfish get vaccinated and feed
124 with granule feed with size between 0.5 and 2.0 mm. When the post vaccination incubation time is
125 over, the cleaner fish are deployed into net pen in the sea together with Atlantic salmon or rainbow
126 trout (*Oncorhynchus mykiss*). An amount of lumpfish ranging between 2 and 15% of the total number
127 of Atlantic salmon individuals reared in each sea net pen are included. Considering limitations for fish
128 density of 25 kg/m³ and maximum allowable biomass of 200,000 fish per unit (Liu et al., 2016), the
129 amount of lumpfish added to the net pens can range from 4,000 and 30,000 individuals.

130 The manager of the farm is thereafter responsible to the acclimation of the lumpfish to its new
131 environment. This happen mainly through the installation of artificial plastic seaweed in the net pen,
132 whose main function is to allow the lumpfish individuals to hide themselves and attach to the
133 substrate with their ventral sucker, as their semi pelagic feature require. One other important action
134 made from the sea farm manager to acclimate the lumpfish is to feed them with granulate food in
135 the range 2.0 to 3.0 mm on a daily basis.

136

137 ***2.1 Goal and scope definition***

138 The goal of this LCA study is to evaluate the environmental impact of the lumpfish farming facility. As
139 a cleaning fish, lumpfish is used for sea lice control. Although different sea lice pest control practices
140 (i.e., biological, chemical and mechanical control) are applied, mainly in Atlantic salmon aquaculture,
141 there is a lack of information on the environmental impacts on pest control measures used in
142 salmonids aquaculture.

143 This study contributes to fill the gap of knowledge about the impact of lumpfish production and
144 provides an important information for decision makers to make more sustainable choices in lumpfish
145 farming and also in the application of treatments for sea lice control in Atlantic salmon farming.

146 The functional unit was defined as 1 kg of live weight of lumpfish. Even if the number of fish could be
147 uses as functional unit the mass was preferred because lumpfish can be used at different size and
148 weight. Besides this, this choice is in agreement with previously carried out LCA study about cleaning
149 fish (Phillis et al., 2021).

150 The system boundaries used, were from cradle to farm gate, including all processes and materials
151 that were used prior to the grow-out phase of the lumpfish within the land based farm, as well as the
152 processes and materials in the grow-out phase itself. The prior processes include the production of
153 juveniles, feed and medicinal treatments, as well as energy use and input transport.

154

155 **2.3 Life cycle inventory**

156 For the assessment of the environmental impacts of the different elements involved in the
157 production of lumpfish, primary and secondary data were used that were obtained from a wide range
158 of sources.

159 Primary data regarding the consumption of the different production factors (e.g., diesel, electricity,
160 chemicals, feed, liquid oxygen and so forth) were directly collected at the lumpfish farm by means of
161 questionnaires and by interview with the farm operators. Table 1 reports the main production factors
162 consumed during lumpfish production.

163

164 **Table 1** – Main inventory data for lumpfish production expressed for the selected FU.

Production Factors	Amount	Unit
Electricity	26.923	kWh
Diesel	31.923	g
Oxygen, liquid	43.846	g
Antibiotics	0.152	g
Feed	0.771	kg

165

166 Secondary data about the emissions related to the combustion of fuel as well as to the output of
167 nitrogen (N) and phosphorous (P) compounds as metabolic waste by fish were estimated. Fuel
168 combustion were assessed according to Spielmann et al. (2007).

169 The output of N and P metabolic wastes by fish depends by a variety of endogenous and exogenous
170 factors, such as genetics, life stage, size, rearing system and diet (Mock et al., 2019). Ammonia is
171 predominant type of N excreted, and high levels of ammonia excretion may be due to high protein
172 intake or inadequately formulated diets which provide unbalanced protein synthesis. Phosphorus
173 excretion usually accounts for 69-86% of dietary P and is associated with the sources, which are used

174 in different ways by different species (Lazzari and Baldisserotto, 2008). In this study, the emission of
175 N and P compounds were estimated according to Cho and Kaushik (1991).

176 Background data about the unitary impact of fuels, chemicals and feed were retrieved by Ecoinvent
177 v 3.6. The inventory data were processed using the software SimaPro 9.1.1.

178

179 ***2.4 Life cycle impact assessment***

180 The inventory data were converted into potential environmental impacts using the characterization
181 factors defined by Recipe LCIA method (Goedkoop et al., 2009; Huijbregts et al., 2017). In detail, the
182 following midpoint impact categories were considered:

- 183 - Global warming (GW, expressed as kg CO₂ equivalent or eq.),
- 184 - Stratospheric ozone depletion, (ODP, expressed as mg CFC11 eq.),
- 185 - Ionizing radiation (IR , expressed as kBq Co-60 eq.),
- 186 - Ozone formation, Human health, (HOPF, expressed as g NO_x eq.),
- 187 - Fine particulate matter formation, (PMFP, expressed as g PM_{2.5} eq.),
- 188 - Ozone formation, Terrestrial ecosystems, (EOFP, expressed as g NO_x eq.),
- 189 - Terrestrial acidification, (TAP, expressed as g SO₂ eq.),
- 190 - Freshwater eutrophication, (FEP, expressed as g P eq.),
- 191 - Marine eutrophication, (MEP, expressed as g N eq.),
- 192 - Terrestrial ecotoxicity, (TETP, expressed as kg 1,4- dichlorobenzene - DCB),
- 193 - Freshwater ecotoxicity, (FETP, expressed as kg 1,4-DCB),
- 194 - Marine ecotoxicity, (METP, expressed as kg 1,4-DCB),
- 195 - Human carcinogenic toxicity, (HTPc, expressed as kg 1,4-DCB),
- 196 - Human non-carcinogenic toxicity, (HTPnoc, expressed as kg 1,4-DCB),
- 197 - Land use (LU, expressed as m²a crop eq.),

198 - Mineral resource scarcity, (SOP, expressed as g Cu eq.),

199 - Fossil resource scarcity, (FFP, expressed as kg oil eq.),

200 - Water consumption (WD, expressed as m³).

201

202 3. Results

203 Table 2 reports the potential environmental impact for the selected functional unit (1 kg of live-
204 weight lumpfish) while the contribution analysis is shown in Figure 2.

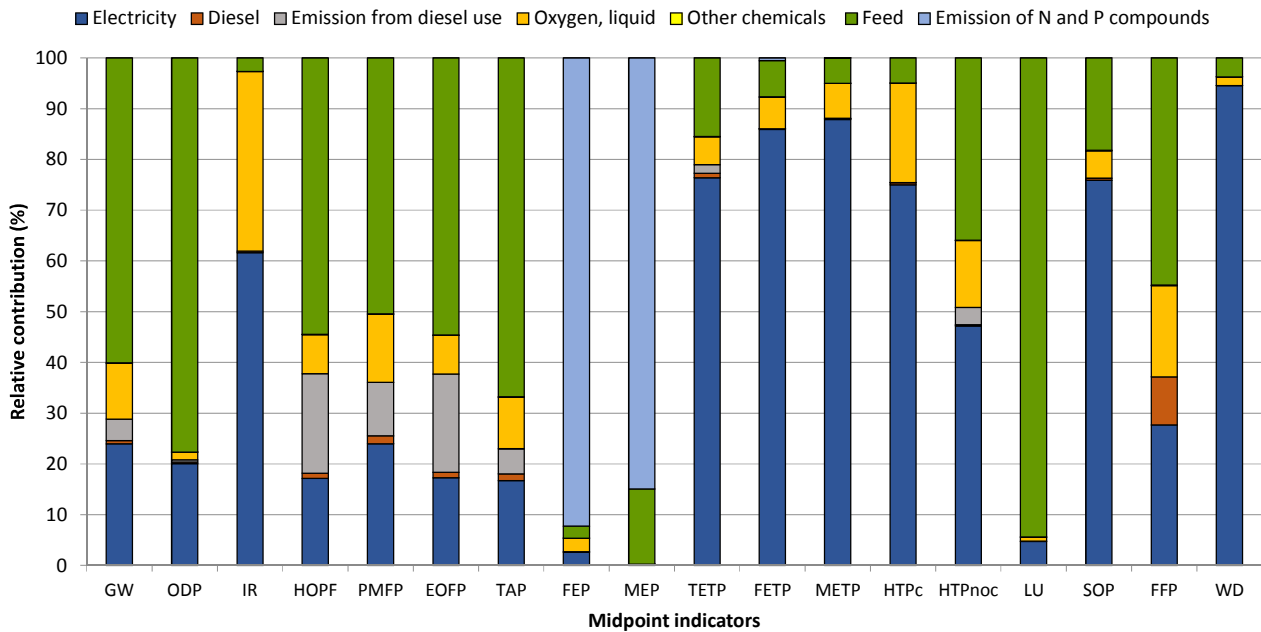
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206 **Table 2** – Absolute potential environmental impact for the selected functional unit.

Impact category	Acronym	Unit	Score
Global warming	GW	kg CO ₂ eq.	2.384
Stratospheric ozone depletion	ODP	mg CFC11 eq.	9.615
Ionizing radiation	IR	kBq Co-60 eq.	0.375
Ozone formation, Human health	HOPF	g NO _x eq.	6.501
Fine particulate matter formation	PMFP	g PM _{2.5} eq.	3.005
Ozone formation, Terrestrial ecosystems	EOFP	g NO _x eq.	6.615
Terrestrial acidification	TAP	g SO ₂ eq.	9.890
Freshwater eutrophication	FEP	g P eq.	9.535
Marine eutrophication	MEP	g N eq.	14.066
Terrestrial ecotoxicity	TETP	kg 1,4-DCB	6.252
Freshwater ecotoxicity	FETP	kg 1,4-DCB	0.196
Marine ecotoxicity	METP	kg 1,4-DCB	0.236
Human carcinogenic toxicity	HTPc	kg 1,4-DCB	0.081
Human non-carcinogenic toxicity	HTPnoc	kg 1,4-DCB	2.685
Land use	LU	m ² a crop eq.	7.258
Mineral resource scarcity	SOP	g Cu eq.	8.359
Fossil resource scarcity	FFP	kg oil eq.	0.394
Water consumption	WD	m ³	0.834

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209

210 **Figure 2** – Contribution analysis for lumpfish production in terms of midpoint indicators.

211

212 For 8 of the 18 evaluated potential environmental impacts (i.e., GW, ODP, HOPF, PMFP, EOFP, TAP,
 213 LU and FFP), feed consumption is the main responsible of the environmental load with a share of the
 214 total impact ranging from 44% for Fossil resource scarcity to 94% for Land Use. The impact of feed
 215 consumption is mainly related to the production of soybean meal, rapeseed and wheat grain.

216 For other 8 impact categories (IR, TETP, FETP, METP, HTPc, HTPnoc, SOP and WD) the main driver of
 217 the environmental results is the electricity consumption while, for the remaining two (FEP and MEP),
 218 more than 90% of the environmental load is related to the emission of N and P compounds by fishes.

219 For Freshwater eutrophication (FEP), the emission of phosphorous compounds is the main
 220 responsible of the environmental impact, while Marine eutrophication (MEP) is mainly related to
 221 ammonia emissions.

222 Except for the liquid oxygen consumption in IR (accounting for 35% of the impact) and for the
 223 emissions related to diesel fuel combustion for the Ozone formation, Human health (HOPF) and
 224 Ozone formation, Terrestrial ecosystems (EOFP) (19% of the impact), the other production factors
 225 consumed and the other emission sources play a minor role.

226 An uncertainty analysis was carried out with the Monte Carlo technique (1,000 iterations and a
227 confidence interval of 95%) to test the robustness of the achieved results. The analysis (detailed
228 results reported in the Supplementary Material in Table S2 and Figure S1) shows a low uncertainty
229 for all the evaluated impact categories except than for IR and WU where the coefficient of variation
230 is 120% and 474%, respectively.

231

232 **4. Discussion**

233 The results of the present LCA study detail the environmental impact of a lumpfish aquaculture
234 facility. As to date the lumpfish is the most abundant fish species farmed to be used as a cleaner fish,
235 and not for human consumption, the impacts originated by the production of this species cannot be
236 directly compared with that of fish that are commonly farmed to serve as food. However, our findings
237 can be compared with those from a recent study that estimated six impact categories for farmed
238 lumpfish, as well as farmed and fished wrasses, used as cleaner fish in delousing farmed Atlantic
239 salmon (Philis et al., 2021). Even if the comparison among LCA studies can be affected by different
240 system boundary and assumptions the two studies selected the same Functional unit and the same
241 LCIA method. Between the two studies, some differences emerged. In fact, the present study showed
242 that GW and MEP impacts of the facility we focused on were lower, but the LU impact was higher,
243 compared to those reported by Philis and co-authors (2021). In contrast, a similar impact in terms of
244 METP was noted. These discrepancies are due to uncertainties and differences during the LCA
245 modelling, from data collection to selection and use of data. In addition, the impact of farmed
246 lumpfish estimated by Philis and co-authors (2021) derived from data collected in different lumpfish
247 facilities, which exploiting different farming technologies and processes returned variable inventory
248 data. For instance, one of the main differences in lumpfish production processes in terms of energy
249 consumption concerns the heating of the water inside the farm in order to increase the metabolism

250 and consequently the hunger of the fish. Interestingly, the same work showed that the impact related
251 to the farming of wrasse was generally higher compared to the lumpfish mainly in terms of GW, WD
252 and METP (Philis et al., 2021). These differences can be due to the longer production cycle of wrasse
253 compared to the lumpfish (3-fold longer) and an unusually high electricity demand despite the use of
254 heat-exchangers, partly due to the longer production cycle, higher sea-water temperature
255 requirements compared to the lumpfish (Philis et al., 2021), and the use of flow-through rearing
256 technology (Brooker et al., 2018). Moreover, farmed wrasse also requires live feeding through the
257 hatching phase and is particularly prone to disease and adaptation difficulties (Helland et al., 2014).
258 However, despite our effort to compare the impacts of different cleaner fish value chains, a significant
259 gap of knowledge remains to couple life cycle emissions generated by the farming, distribution, and
260 use of the cleaner fish and their potential different delousing efficiencies in the salmon net pens
261 (Philis et al., 2021). Intra- and inter-specific differences in delousing efficiencies depend on species
262 types, behaviour, survival and adaptation rates, response to stress, growth speed, operating sea-
263 water temperature and swimming abilities (Brooker et al., 2018). Although some studies
264 demonstrated that lumpfish (Eliassen et al., 2018; Imsland et al., 2018) and wrasse (Leclercq et al.,
265 2014; Skiftesvik et al., 2013) are effective delousers, there is high level of uncertainty regarding their
266 efficiency. For this reason, in the present study we performed a comparison of the environmental
267 impacts caused by two farmed cleaner fish assuming the same delousing efficacy, but in further
268 studies this issue should need to be carefully considered.

269 Lastly, we attempted to estimate the impact of using lumpfish to produce a ton of Atlantic salmon.
270 First, the salmon production (in tons of live salmon ready for slaughterhouse) within a net pen was
271 estimated by considering the volume of a single net pen and the maximum allowed stock density of
272 salmon that can be reared in the net pen (25 kg/m^3 ; Liu et al., 2016) where conventional farming is
273 applied. Considering that the size of net pens varies among Atlantic salmon rearing facilities, we

274 suggested two different scenarios: the first one assuming the use of middle-size net pen (90 m in
275 diameter, 30 m in depth; volume = 19,347 m³) and the second one assuming a large-size net pen (160
276 m in diameter, 40 m in depth; volume = 81,528 m³). Then, as the maximum amount of salmons that
277 can be included in a net pen accounts for 200,000 specimens, we estimated the minimum (2%) and
278 the maximum (15%) amount of lumpfish that can be added to a single net pen of both sizes,
279 corresponding to 4,000 and 30,000 individuals, respectively. We calculated the impact, in terms of
280 Global warming, Terrestrial acidification and Freshwater eutrophication per ton of salmon reared in
281 middle-size and large-size net pen, as well as per ton of lumpfish (both minimum and maximum
282 amount). Lastly, we calculated the contribution of the use of lumpfish to the environmental impact
283 due to produce a ton of salmon. We considered the impact to produce a ton of Atlantic salmon in
284 terms of global warming (2793.5 kg CO₂ eq./ton), terrestrial acidification (25.1 g SO₂ eq./ton) and
285 freshwater eutrophication (66.5 g P eq./ton) as the mean of each specific endpoint according to
286 previous LCA studies of the impact of Atlantic salmon aquaculture (see Philis et al., 2019 and
287 references therein). In middle-size net pen, the addition of the minimum or maximum amount of
288 lumpfish accounted for the 1.76% and 13.23% in terms of global warming, while in large-size net
289 pens, the contribution accounted for 0.42% and 3.14%, respectively. Similarly, the share of the
290 Atlantic salmon impact related to the use of lumpfish in terms of Terrestrial acidification and
291 freshwater eutrophication was lower in the scenario that considered large-size net pen (0.19 – 1.44%
292 for g SO₂ eq.; 0.07 – 0.53% for g P eq.) than middle-size net pen (0.81 – 6.10% for g SO₂ eq.; 0.29 –
293 2.22% for g P eq.). These results seem to confirm that open net farming with large net pen volumes,
294 exceeding 60,000 m³ in one pen, are more energy- (and cost-) efficient than smaller ones (Ziegler et
295 al., 2013). However, these estimates must be considered with caution because they relied on
296 assumptions and might suffer a moderate degree of uncertainty.

297

298 ***4.1 Strategies to improve sustainability of lumpfish aquaculture***

299 Our data highlighted the manifold environmental impacts of lumpfish farming that cannot be
300 neglected and need to be explored in depth to opportunely enlarge its environmental sustainability.
301 To increase the sustainability of lumpfish aquaculture, some technical improvements of the facility,
302 as well as of the farming processes, could allow to reduce the contribution of some impact category.
303 Concerning the farming processes, a series of steps forwards, including improvements in collection
304 and transport of wild breeders, reproduction procedures and development of broodstock reared
305 entirely in captivity, might be undertaken to make lumpfish farming for sea-lice control more
306 sustainable (Powell et al., 2018). To date, nearly all the lumpfish used as cleaner fish in Atlantic salmon
307 farming industry come from wild-caught parents, which after being used as breeders are sacrificed,
308 affecting natural populations of the species. This issue is particularly relevant because the lumpfish is
309 considered a moderate to high vulnerable species (Froese and Pauly, 2014) and it has been classified
310 as near threatened (NT) in the IUCN Red List (Lorance et al., 2015). Thus, to limit the use of wild
311 breeders and to keep pace with the growing demand of lumpfish for sea lice control, the breeding
312 cycle needs to be closed in captivity (Anon, 2015) and future production needs to be derived entirely
313 from selected farmed strains (Powell et al., 2018).

314 Another strategy to reduce the collection of wild breeders concerns the cryopreservation of milt from
315 wild male (Powell et al., 2018). Sperm cryopreservation is a well-known advantageous methodology
316 for fish reproduction in aquaculture, mainly in seasonal breeders (Cabrita et al., 2005; Martínez-
317 Páramo et al., 2017), that allows the reduction of wild breeders maintaining a high fish production.
318 Few studies validated methods for cryopreserving milt of lumpfish, suggesting that this strategy can
319 be used for hatchery management in lumpfish aquaculture (Pountney et al., 2020) and to maintain a
320 stable production throughout the year of lumpfish juveniles (Norðberg et al., 2015). These
321 improvements should return two crucial outcomes. On one hand, decreasing the wild-catch of

322 breeders should reduce the environmental impacts related to fishing activities of breeders (e.g., fuel
323 consumption, release of hazardous contaminants, use of chemicals during transport to the farming
324 facility), while on the other hand, closing the breeding cycle in captivity should allow to select strains
325 with the desired compromise of property as high delousing performance, fish health and resistance,
326 growth rate and so on. In fact, lumpfish families show a dissimilar efficiency in feeding on sea-lice
327 (Imsland et al. 2016), suggesting the existence of a genetic component for sea-lice consumption that
328 might be used to select specific strains with high affinity to prey sea-lice (Powell et al., 2018).

329 Selecting strains with slow growth could be also advantageous because lumpfish show less interest
330 regarding eating sea-lice at a size of about 300–400 g (Anon 2014). Thus, strain selection should
331 reduce the amount of lumpfish to be produced, farmed and transferred to Atlantic salmon net pens
332 for delousing activities, decreasing the environmental impact of lumpfish farming. Another crucial
333 issue that could increase the sustainability of lumpfish farming concerns the re-use of individuals
334 after their deployment in Atlantic salmon net pens (Powell et al., 2018). To date, lumpfish are used
335 only for a salmon production cycle. Whilst the most of them died within net pens, the survivors are
336 generally culled because of impairment of their health status due sub-optimal rearing conditions and
337 the subsequent decrease in delousing efficiency. As this practice has been criticized because wasteful
338 and with diverse animal welfare implications (Anon 2013; Farm Animal Welfare Committee 2014),
339 lumpfish survived to the harsh conditions experienced during a salmon production cycle could be
340 used to create a broodstock of high-resistant individuals to be used in captive breeding programmes.

341 Considering the high mortality of lumpfish in the net pen, also due to predatory behaviour and bites
342 by salmons (Espmark et al., 2019), further input of fish result as necessary to guarantee the delousing
343 activity. Developing high-resistant strains from post-deployment individuals, should allow to prevent
344 new introductions of cleaner fish and, consequently, to reduce the number of individuals to be
345 farmed and the impact of farming activities.

346 The optimization of larval production might reduce the impact of lumpfish aquaculture. The selection
347 of well-adapted strains in captivity and the improvement of formulation of diets in early
348 developmental periods should contribute to reduce the high mortality occurring during larval
349 weaning, specifically during the transition from live to dry feeds (Powell et al., 2018). Improving the
350 composition of the diet and/or optimizing the amount of feed administered to larvae during post-
351 hatching periods should be particularly important considering that feed consumption has been
352 identified as the main responsible of environmental impacts in terms of midpoint indicators of
353 lumpfish farming.

354

355 **5. Conclusions**

356 The present study detailed the environmental impacts of lumpfish farming to be used as cleaner fish
357 in Atlantic salmon aquaculture. Considering the importance of the use of cleaner fish, and in
358 particular of the lumpfish, in delousing salmonids in open-water net pens, our data lay the
359 foundations to optimize the entire process of lumpfish farming, promoting a transition towards a
360 more sustainable production. For instance, considering the impacts pointed out by this study, some
361 mitigation measures could be implemented. As the main contributor to the environmental impacts
362 of lumpfish aquaculture come from feed, decreasing the feed administration or improving the feed
363 formulation could be a strategy to be implemented. At the same time, reducing the energy
364 consumption in the lumpfish facility (e.g., to heat the water) and the use of fuel (e.g., reducing the
365 distances of transport or the number of fishing operations to collect breeders) might reduce the
366 environmental impacts of lumpfish farming.

367 In addition, our findings are crucial to compare the environmental impacts of biological, mechanical,
368 and chemical treatments exploited by salmon farmers to delouse fish, as well as to estimate the
369 contribution of each single treatments to the salmon footprints.

370 Moreover, the results of this study can be used by owner and decision maker at the cleaner fish land
371 based facilities as a tool for a confrontation of lumpfish production impact with the emerging farming
372 activity of wrasse species used as supplementary cleaner fish as ballan wrasse (*Labris bergylta*).
373 Policy maker could also benefit from the outcome of this study because the LCA is a standardized
374 assessment tool for potential improvement of the juridical frame necessary to regulate the new
375 emerging delousing methods and best practice. Last but not least, this information should drive
376 salmon farmers towards the application of one or treatment mix returning the lowest environmental
377 impacts without undermining fish welfare, fish production and economic gain.

378

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Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Author contribution

Federico Haaland Gaeta: Data collection, Investigation, Writing - review & editing; **Marco Parolini**: Writing - original draft, Supervision; **Jacopo Bacenetti**: Formal analyses, Writing - review & editing