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3 Traditionally managed landscapes do not prevent amphibian decline and the
4 extinction of paedomorphosis

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19 ORIGINAL RESEARCH ARTICLE

20

21 *Abstract.* Eco-cultural landscapes are assumed to be favourable environments for the
22 persistence of biodiversity, but global change may affect differently their terrestrial and aquatic
23 components. Yet, few long-term studies have examined how multiple, global change stressors
24 may affect wetland biodiversity in such environments. Facultative paedomorphosis is a
25 spectacular example of intra-specific variation, in which biphasic (metamorphosing) amphibians
26 coexist with fully aquatic conspecifics which do not metamorphose (paedomorphs).
27 Paedomorphosis is seriously threatened by global change stressors, but it is unknown to what
28 extent traditional management will allow its long-term persistence. Here, we tested the effects of
29 alien species introductions while taking into account land-use and climate changes on the
30 distribution of two polymorphic newt species (*Ichthyosaura alpestris* and *Lissotriton graecus*) in
31 Montenegro by using a 68-year data set and Bayesian mixed models integrating complex spatial
32 and temporal structures. We found that, despite the persistence of natural landscapes,
33 metamorphs dramatically declined, and paedomorphs were nearly extirpated, losing 99.9% of
34 their aquatic area of occupancy and all the major populations. Fish introduction was the main
35 determinant of decline for both phenotypes. Climate and the presence of crayfish further
36 contributed to the decline of metamorphs, which started later and was less dramatic than that of
37 paedomorphs. The near extinction of paedomorphosis on a country-wide scale shows how
38 invasive species determine broad scale impacts, which can be even stronger than other global
39 change stressors, and underlines the need for immediate management actions to avoid the
40 extinction of a unique developmental process, paedomorphosis.

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42 *Key words:* Alien species, amphibian decline, biodiversity loss, climate change, fish
43 introductions, eco-cultural landscape, freshwater habitats, global change, invasive species, land-
44 use, paedomorphosis, traditional landscape

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46

INTRODUCTION

47 Traditional cultural landscapes are the result of centuries long co-evolution between the social
48 and the ecological systems. Within the context of Europe, these systems are thought to have
49 exceptional natural values, several protected species and habitats being maintained by the
50 traditional agricultural and forestry practices (Hartel et al. 2010, Fischer et al. 2012, Rotherham
51 2015). Aquatic habitats are typical components of traditional landscapes. These habitats can
52 serve a wide range of social functions, from recreation to the provisioning of water for humans
53 and livestock (Hammit et al. 1994, Boix et al. 2012, Hartel and von Wehrden 2013). They are
54 also essential for the persistence of a large diversity of native species because water offers vital
55 resources, such as a place for life, reproduction, development of aquatic larval stages, foraging
56 and hydration (Dudgeon et al. 2006). The Convention of Biological Diversity highlighted the
57 need to protect waters, but there is a lack of data on the long-term evolution of the status of
58 freshwater environments (Williams et al. 2004, Vorosmarty et al. 2010, Abell et al. 2017).
59 However, there is evidence that global changes may have a greater effect on freshwater fauna
60 than on terrestrial fauna (Ricciardi and Rasmussen 1999, Bignal and McCracken 2000). It is
61 therefore important to determine the ecological transitions occurring in these habitats,
62 particularly with respect to the globalisation of anthropogenic activity, land use and climatic
63 features (Stoate et al. 2009, Gordon et al. 2010, Plăiașu et al. 2012, Hartel et al. 2014). Among

64 freshwater species, amphibians, and more particularly newts, often dominate freshwater habitats
65 such as ponds and mountain lakes (Schabetsberger and Jersabek 1995). Because many
66 amphibian species exhibit a typical biphasic life stage, they are particularly threatened by
67 environmental perturbations acting on both aquatic and terrestrial environments (Semlitsch
68 2003). One of the important causes of decline in aquatic habitat is the introduction of alien
69 species, such as fish and crayfish (Knapp and Matthews 2000, Orizaola and Braña 2006, Pilliod
70 et al. 2010, Bucciarelli et al. 2014, Havel et al. 2015, Miró et al. 2018). In response to these
71 introductions, populations can collapse or subsist through dispersal to the terrestrial habitat and
72 the use of alternative breeding patches, provided they are available (Winandy et al. 2015, Tiberti
73 2018).

74 Many newt and salamander species express paedomorphosis, a developmental trait which
75 makes them fully aquatic, as it involves the retention of larval traits (e.g. gills) at the adult stage
76 (Fig 1). This is believed to be important in the adaptation and evolution of species (Gould 1977,
77 Denoël et al. 2005b, Bonett and Blair 2017). Whereas it is obligate in some species which lose
78 the ability to metamorphose, it is a polyphenism in other species (facultative paedomorphosis)
79 and, thus, involves the coexistence of both a paedomorphic and a metamorphic phenotype as a
80 response to environmental drivers (Oromi et al. 2016, Mathiron et al. 2017). Yet, a genetic basis
81 of paedomorphosis was found in some species (Voss and Shaffer 1997) and constraining
82 environments can promote metamorphosis over paedomorphosis across generations (Semlitsch
83 and Wilbur 1989). Monitoring the state and temporal change of these particular populations can
84 help inform us about local and global threats to their habitats as well as on both developmental
85 processes. Indeed, the decline of paedomorphs can result from severe, but localized perturbation
86 in the aquatic habitat, while a decline of both phenotypes likely reveals longer-term effects and a

87 larger perturbation as metamorphs can survive, at least temporarily, detrimental waters (Denoël
88 et al. 2009).

89 Until recently, no long-term studies have examined how multiple, global change stressors
90 affect the diversity of ponds and mountain lakes on a country-wide scale in areas dominated by
91 eco-cultural landscapes. Montenegro is a good candidate in which to investigate the ecological
92 transition of freshwater environments in traditionally managed landscapes because of its
93 geographic location within the Mediterranean hotspots, its richness in many endemic species, its
94 low populated areas and the area's prevalence of traditional agricultural and wild landscapes
95 (Griffiths et al. 2004, Mittermeier et al. 2004). Moreover, among Balkan countries, Montenegro
96 has historically been known to host numerous paedomorphic populations and endemic taxa
97 (Radovanović 1951, 1961, Džukić et al. 1990) that have been surveyed across decades. Fish
98 introductions were linked to declines in some amphibian populations following a two-step
99 process, involving first the loss of paedomorphosis followed thereafter by metamorphosis. Yet,
100 previous surveys showed the local persistence of the paedomorphs in some sites and the
101 generalized persistence of the metamorphs up to the early 2000's (Breuil 1985, Denoël et al.
102 2005a, Denoël et al. 2009). From this point, it is unknown if both could have recovered, or
103 whether declines continued following the same trend. Because many factors, such as land-use
104 and climatic changes, can affect natural populations (Walls 2009, Ficetola et al. 2010), it is
105 important to differentiate their contribution to the ongoing declines from the effects of alien
106 species introductions.

107 In this study, we took advantage of a data set covering more than 60 years to assess the
108 impact of multiple threats on the freshwater habitats of the traditionally managed landscapes of
109 Montenegro (Breuil 1985, Džukić et al. 1990, Denoël et al. 2009). We analysed representative

110 sites across the entire Montenegro to determine four phenomena: (1) whether pond and lake
111 environments suffer biodiversity loss – focusing here on amphibians – even if their surrounding
112 terrestrial environment suffers limited habitat loss, (2) the current status and decline of all known
113 populations of paedomorphic newts, including those of endemic taxa as historically described in
114 Montenegro, (3) whether common terrestrially-adapted phenotypes (i.e. metamorphs) are
115 similarly threatened as are rare aquatic phenotypes (i.e. paedomorphs) and (4) the identification
116 of the potential drivers of decline, focusing on alien species introductions while taking into
117 account land-use and climate changes.

118

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METHODS

120

Studied localities and species

121 We focused on the two newt species in which paedomorphosis is the most frequently expressed
122 in the Balkans: the alpine newt (*Ichthyosaura alpestris*) and the Greek smooth newt (*Lissotriton*
123 *graecus*) (Amphibia, Salamandridae) (Fig 1). These two species were, until recently, classed
124 within the genus *Triturus*, whereas *L. graecus* was formerly considered as a subspecies of
125 *Lissotriton vulgaris* (Wielstra et al. 2018). In this study, we considered the 23 main localities in
126 which paedomorphosis is expressed in Montenegro (Ćirović 2009; pers. obs., Denoël et al. 2009)
127 and for which historical data are available (Appendix S1: Tables S1 and S2). These data came
128 from a combination of our own observations (see also Džukić et al. 1990) and bibliographical
129 data, including data representing type description (Radovanović 1951 for the oldest record). An
130 examination of the conserved specimens (Institute for Biological Research ‘Siniša Stanković’,
131 Belgrade, Serbia) was also done in 2017 to complete our data set and confirm the historical

132 presence of paedomorphosis (Appendix S1: Table S3 and Fig S1) (Džukić et al. 2015).
133 Combining bibliographic and direct observations, our data covered 68 years (1948–2016), with
134 direct surveys covering the period 1970-2016.

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136

Sampling procedure

137 Sampling techniques varied across sites due to the differences among the studied freshwater
138 habitats (lakes, ponds and wells). To limit observer effects across sites, one of the authors (MD)
139 was involved in all the surveys of 2010's, as well as in a number of the historical ones. In all
140 sites, we first used visual sampling techniques by walking around all the water bodies. Then, we
141 used dip-netting of either visually encountered specimens or those which could be taken blindly
142 (i.e. without initial sight) from the most favourable areas (e.g. aquatic vegetation). Moreover,
143 when possible, aquatic rocks were turned over to look for hidden newts. Most surveys were done
144 during one or two visits (several hours minimum per visit), but up to two weeks were spent in the
145 most important historical sites. In the deep Bukumirsko Lake, we also used 'minnow' traps in all
146 the benthic micro-habitats, including the deepest ones. Traps were built with plastic 1.5 L water
147 bottles from which the neck of the bottle was inverted to point toward the interior of the bottle.
148 Some historical sites (particularly #7 and #15) were also highly prospected as part of in depth
149 studies on the evolutionary ecology of newt paedomorphosis (e.g. Kalezić et al. 1996). All
150 sampling followed ethical standards, and the studied newts were released directly back into their
151 habitat of capture after each census.

152 Non-observing a species in a site during a particular survey can mean that the species is
153 absent, or that it is present but has remained undetected, and not taking into account this issue

154 can result in biased inference (MacKenzie et al. 2017). A subset of 11 sites were surveyed
155 multiple times during the same year (average: 2.6 surveys per site; range: 2-7), by means of the
156 multiple techniques used in all the surveys. We therefore used occupancy analysis to assess the
157 reliability of the presence/absence pattern (MacKenzie et al. 2017). Occupancy models were run
158 using the unmarked package in R 3.3 (Fiske and Chandler 2011), assuming constant detection
159 probability across all surveys. Occupancy analysis suggested that the per-visit detection
160 probability was very high for both paedomorphic and metamorphic newts. For paedomorphs, the
161 estimated detection probability was ~100% (SE = 0.001), while for metamorphs the estimated
162 detection probability was 90% (SE = 7) per visit. Therefore, one single survey was generally
163 enough to ascertain newt occupancy with high reliability and imperfect detection was not a major
164 issue in our data set.

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166

Newt identification

167 Newts were identified according to species and phenotype. Paedomorphs were distinguished
168 from metamorphs by the presence of external gills and open gill slits. The adulthood of each
169 individual was established on the basis of a well-developed cloaca (Denoël 2017). Occurrence
170 data on both paedomorphs and metamorphs at different times allowed us to analyse the timing
171 of their decline. We also recorded the number of individuals caught, but, as these data were not
172 available historically for most of the sites, we used only occurrence data. The analysis of
173 ecological correlates of paedomorphosis versus metamorphosis was performed previously
174 (Denoël et al. 2009); therefore, the present study focused only on the populations which
175 expressed paedomorphosis in at least one survey. All these populations had also historically
176 expressed metamorphosis.

177

178

Habitat characterization

179 We determined the persistence of the water bodies and recorded the presence of introduced alien
180 species, i.e. fish and crayfish. We calculated the extent of occurrence (EOO) as the minimum
181 convex polygon for each species and the area of occupancy (AOO) as the sum of areas of cells,
182 considering 2 x 2 km cells as recommended by the IUCN (IUCN 2017). We also measured the
183 surface area and maximum water depth of each pond and lake, and calculated the aquatic area of
184 occupancy (AAOO) as the sum of all occupied surface areas. Finally, we classed water bodies in
185 three categories: lakes, ponds and wells.

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Land use

188 To determine temporal changes in terrestrial landscapes, we extracted the cover percentage from
189 the Corine Land Cover maps (<https://land.copernicus.eu>). We considered natural habitats (Corine
190 Land Cover classes 311-324 and 332) and agricultural lands (211-242). These layers constitute
191 most of the local land use which can represent favourable versus potentially unfavourable
192 habitats for newts, respectively. In Montenegro, the natural and traditionally managed lands are
193 typically characterized by low densities of cattle, such as sheep and cows that graze either in
194 semi-open landscapes constituted of bushes and patches of woods at mid-elevations and alpine
195 meadows at the highest altitudes, and with very limited use of crops, then contrasting with
196 intensive agriculture as commonly found in Western Europe (Lakovic et al. 2016). Impacts on
197 natural lands by farmers have been generally low, as seen by resilience patterns of natural
198 vegetation over the last decades (Nyssen et al. 2014). There were no urban areas, or almost none,

199 around the studied ponds and lakes according to Corine (Fig 2). The Corine Land Cover
200 programme was started in 1985 by the European Community to generate digital land-use/land-
201 cover maps covering the European continent. Analyses were based on four Corine Land Cover
202 maps: the Corine 1990 (generated using satellite images taken during 1986-1988); Corine 2000
203 (images from 1999-2001), Corine 2006 (images from 2005-2007) and Corine 2012 (images from
204 2011-2012). All the measures of landscape features were computed at two radii (100 and 1,000
205 m) around the periphery of the water bodies. These radii were chosen because they have been
206 shown to act on newt distribution (Denoël and Ficetola 2007, Denoël et al. 2013). The 100 m
207 radius is expected to have the most importance when directly in contact with water and because
208 most metamorphic newts stay in close proximity to water during their terrestrial phase
209 (Semlitsch 1998, Jehle and Arntzen 2000).

210

211

Climate

212 As climatic variables, we considered the mean annual temperature and the total annual
213 precipitation during the period before each survey for each locality. Climate was calculated as
214 the mean value considering the year of the survey and the four years preceding years. Climatic
215 data were extracted from the Climatic research Unit (CRU) 4.01 climate grids (Harris et al.
216 2014). The CRU 4.01 climate grids contain the monthly values of precipitation and temperature
217 for 1901–2016, on the basis of data collected from meteorological stations over the entire globe.
218 These data have a coarse resolution (resolution: 0.5°, i.e. approx. 50 km for the study area)
219 compared to the distance among nearby sites, and no high-resolution time series of climatic data
220 were available for the study area. Therefore, we downscaled the CRU on the basis of the
221 CHELSA high-resolution layers of mean annual temperature and calculated the total annual

222 precipitation (30 arc-seconds resolution) (Karger et al. 2017), using the change factor approach
223 (Diaz-Nieto and Wilby 2005). The average data of the CHELSA climatology are mean values
224 over the 1979-2013 period. Therefore, we calculated the change factor between the CHELSA
225 data and the average CRU data for the 1979-2013 period, and then used them to downscale the
226 resolution of annual layers. Given that downscaling might produce bias, we also repeated
227 analyses using the CRU data (without downscaling), and obtained nearly identical results.

228

229

Statistical analyses

230 We used Bayesian mixed models with binomial error distribution to assess the factors
231 determining changes in the distribution of paedomorphic and metamorphic newts in the 23 water
232 bodies, while taking into account multiple typologies of non-independence among observations.
233 As independent variables, we considered six parameters which can determine newt distribution.
234 Two variables represented the distribution of non-native predators (the presence of introduced
235 fish and crayfish); two variables represented the climate during the years preceding the survey
236 (mean temperature and total annual precipitation); two represented landscape composition during
237 the period of survey (cover of natural vegetation and of agricultural land within 100 m; obtained
238 from the Corine Land Cover data from the closest period). All independent variables were scaled
239 (mean = 0 and variance = 1) before analyses. The correlation between independent variables was
240 weak ($|r| < 0.5$ in all pairwise correlations), suggesting that collinearity was not a major issue for
241 our analyses. Landscape variables were only available since 1985; therefore, we repeated
242 analyses twice: we first analysed the entire data set (1948-2016), not considering land use, and
243 then analysed only data collected after 1984, including landscape as an additional variable.

244 We used first-order autoregressive models to take into account temporal autocorrelation.
245 We included site identity as a random factor. Furthermore, we used a conditional autoregressive
246 term to take into account spatial non-independence between nearby sites. We repeated these
247 analyses two times. In a first set of analyses, we aimed to assess the overall causes of the decline
248 of paedomorphs and metamorphs, without focusing on interspecific differences; therefore, we
249 included species identity as a random factor. In a second set of analyses, we tested whether there
250 were interspecific differences in the response to threats. Therefore, we included species identity
251 as a fixed factor, and we also tested the interactions between species identity and the threatening
252 factors. We used the Integrated Nested Laplace Approximations (INLA) approach to fit mixed
253 models and calculate 95% credible intervals (CIs). INLA is a computationally effective and
254 extremely powerful alternative to Markov Chain Monte Carlo to run Bayesian models, which is
255 particularly appropriate for data sets with complex spatial and temporal dependencies (Bivand et
256 al. 2015, Bivand et al. 2017, Rue et al. 2017).

257

258

RESULTS

259

Historical distribution of paedomorphosis

260 Both phenotypes (i.e. paedomorphs and metamorphs) have been described in 23 localities of
261 Montenegro (11 for the alpine newt and 12 for the Greek smooth newt: Fig 2, Appendix S1:
262 Tables S1 and S2). Examination of the museum collections of 19 historical localities and field
263 verifications in the four remaining localities confirmed the sexual maturity of gilled newts, i.e.
264 their paedomorphosis in all of the 23 studied populations (Appendix S1: Table S3, Fig S1). This
265 sample represents almost all the populations of Montenegro (all major ones), as very few

266 additional mentions of occurrence were available (i.e., three records in museum, only one in the
267 literature: (Denoël et al. 2005a)) but not usable here given the lack of data.

268 The aquatic habitats where paedomorphs were historically found encompassed mountain
269 lakes (30%), ponds (52%) and wells (17%) (Appendix S1: Fig S2). Sixty-four percent of
270 paedomorphic alpine newts were found in lakes, whereas all paedomorphic Greek smooth newts
271 were found in ponds and wells. The lakes, ponds and wells had a mean \pm SE surface area of $79 \pm$
272 24 m^2 , $432 \pm 129 \text{ m}^2$ and $18 \pm 3 \text{ m}^2$, respectively. All aquatic habitats were permanent and deep
273 (mean \pm SE water depth of lakes = $14.0 \pm 4.3 \text{ m}$, ponds = $2.2 \pm 0.3 \text{ m}$ and wells = $2.6 \pm 0.8 \text{ m}$).
274 See Appendix S1: Table S1 for details on the aquatic habitats.

275

276 *Land-use and climate changes*

277 The cover of forests and shrubland was, on average, $> 60\%$ if measured within a 100-m buffer,
278 and was, on average, $> 70\%$ if measured within a 1000-m buffer (Appendix S1: Table S4; Fig
279 S3). The agricultural cover was generally limited (Appendix S1: Table S4). In 2012, landscape
280 changes were generally limited. Agricultural cover remained stable and natural vegetation
281 showed an increase of approximately 18%, which corresponded to a decrease of mosaics
282 between agricultural and natural lands (Fig 2; Appendix S1: Fig S3; Table S4). Only one of the
283 studied aquatic habitats was destroyed during the study period.

284 Climatic parameters showed strong variation throughout the study period. Total annual
285 precipitation showed strong variation among these years, but we did not detect obvious long-
286 term trends (Appendix S1: Fig S4). For temperature, in the study sites, there was a trend towards
287 warmest mean annual temperature, particularly after 1980. The average annual temperature of

288 the last years was roughly 1.5 °C higher than average temperatures of the 1970's (Appendix S1:
289 Fig S4).

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291

Freshwater habitat deterioration

292 Major disturbance was found in 57% of the studied Montenegrin freshwater habitats (100% of
293 lakes, 25% of ponds and 25% of wells). Considering the specific habitats of alpine and Greek
294 smooth newts, 73% and 42% of the water bodies were disturbed, respectively. The most frequent
295 observed disturbance was the presence of aquatic alien species, notably fish (52%) and crayfish
296 (17%), and these disturbances occurred over the whole of Montenegro (Fig 3a; Appendix S1:
297 Table S5). In habitats with fish there were, on average, 2.5 fish species present (range: 1-6
298 species), belonging to Salmonidae, Cyprinidae and Ictaluridae; all crayfish were Astacidae. All
299 lakes had introduced fish, and all localities with invasive crayfish also had introduced fish.
300 Introductions took place in all types of habitats, except wells. Only one case of crayfish
301 introduction was observed in the 2000's, the others occurring after 2010.

302

303

Paedomorphic newt decline

304 Paedomorphs of both the alpine and Greek smooth newts have declined progressively since their
305 historical discovery (Fig 4 and Appendix S1: Table S6). During the 2000's and 2010's surveys,
306 paedomorphs were found in 57% and 22% of the historical localities, respectively (Figs 2 and 4;
307 Appendix S1: Table S2). For paedomorphic alpine newts, in the 2000's, only 36% of historical
308 populations persisted, and only 18% of populations persisted in the 2010s. For paedomorphic
309 Greek smooth newts, in the 2000's, 75% of historical populations persisted, and only 25% of

310 populations persisted in the 2010's. No paedomorphs subsisted in lakes, while they persisted in
311 17% of rocky ponds and in 75% of wells.

312 The extent of occurrence (EOO), the area of occupancy (AOO based on 2 x 2 km grid
313 cells) and the aquatic area of occupancy (AAOO based on the surface of water bodies) dropped
314 considerably in both studied species since their historic record began. Alpine newts lost 99.99%
315 of EOO, 81.81 % of AOO and 99.99% of AAOO, while Greek smooth newts lost 97.2% of
316 EOO, 83.33% of AOO, and 95.41% of AAOO (Fig 5). During the last decade (2000's to 2010's),
317 the decline in EOO, AOO and AAOO was 99.99%, 73.68% and 50% in the alpine newt, and
318 95.40%, 92.82% and 77.78%, respectively, for the Greek smooth newt (Fig 5).

319 Bayesian autoregressive mixed models showed that fish presence was the main driver of
320 paedomorph extirpation. When we analysed the entire period, there was a strong negative
321 relationship between fish and paedomorph presence in ponds, while the CIs of climatic and land
322 cover variables, as well as crayfish presence, widely overlapped zero (Fig 6). Results were
323 similar when we focused on the period 1985-2016, and we did not detect any relationship
324 between paedomorph persistence and landscape variables (Fig 6).

325 When we considered species as a fixed factor, we observed a slightly lower persistence
326 rate for Greek smooth newts, but the CIs overlapped zero (Appendix S1: Table S7). Models were
327 similar to the ones considering species identity as a random factor (Appendix S1: Table S7).
328 When we added the interactions between species and independent variables to this model, none
329 of them was significant, suggesting a similar response to the different stressors across species.

330

331

Metamorphic newt decline

332 Although metamorphs remained present in all sites in the early 2000's, they were only found in
333 52% of localities in the 2010's (lakes: 43%, ponds: 50%, wells: 75%). They declined
334 significantly later than paedomorphs in both species (non-overlap of CI bands: see Fig 4).

335 Some metamorphic newts could still be found in sites with fish, but they were fully or
336 almost fully absent in places where crayfish had been introduced. Auto-regressive mixed models
337 showed that invasive species and climate can act jointly, determining the extirpation of
338 metamorphic newts. When we analysed the whole period, there was a strong negative
339 relationship between fish and metamorph presence in ponds. The effect size of climate and
340 crayfish presence was weaker than the effect size of fish presence. Nevertheless, the 95% CIs of
341 these variables did not overlap zero, suggesting a negative relationship between metamorph
342 persistence and crayfish presence, and between metamorph presence and mean annual
343 temperature during the years before the survey (Fig 6). Results were identical when we focused
344 on the period 1985-2016. Fish presence showed the strongest, negative effect size; crayfish
345 presence and temperature during the previous years showed negative effect sizes with 95% CIs
346 not overlapping zero, while we did not detect any relationship between metamorph persistence
347 and precipitation or landscape variables (Fig 6).

348 When we considered species as a fixed factor, we observed a lower persistence rate for
349 Greek smooth newts (Appendix S1: Table S7). The credible intervals of the two interactions
350 (species \times crayfish presence and species \times cover of natural habitat) did not overlap zero. Crayfish
351 presence more strongly affected alpine than Greek smooth newts, as 75% of crayfish occurrences
352 affected alpine newts. An interaction between natural habitat and species identity suggested that
353 alpine newts suffered from the loss of surrounding habitat more than Greek newts.

354

DISCUSSION

355

356 Whereas terrestrial environments showed limited alterations in the studied eco-cultural
357 landscapes, the associated aquatic landscapes have been deeply affected over the last decades.
358 The main driver of change was the introduction of alien aquatic species, which persisted and
359 expanded over the last decades. This resulted in a rapid biodiversity loss as shown by the high
360 decline of amphibian populations. Even worse, the fully aquatic and rare paedomorphic
361 phenotypes are now on the edge of extinction on a country-wide scale. Although Montenegro
362 was historically a hotspot of the unique evolutionary process which is paedomorphosis (Džukić
363 et al. 1990), this is no longer the case. Because newts and, even more specifically, paedomorphs
364 used to be the native top predators of these freshwater habitats, the current situation is, therefore,
365 more than just alarming, suggesting that many of these aquatic ecosystems may be globally
366 impacted.

367 In Western Europe, terrestrial lands are deeply affected by anthropogenic change, and the
368 alteration and disappearance of freshwater habitats have been well-documented on a variety of
369 scales (Wood et al. 2003, Ferreira and Beja 2013, Arntzen et al. 2017). In many cases, the
370 remaining aquatic habitats became less favourable for the establishment of native aquatic fauna
371 (Ficetola et al. 2011, Denoël et al. 2013). In contrast, it is often assumed that alpine
372 environments and other traditionally managed lands continue to offer valuable resources to
373 sustain biodiversity (Hartel et al. 2010, Rotherham 2015). Although agricultural practices are
374 rapidly changing in Eastern European countries (Pătru-Stupariu et al. 2016) (but see Miró et al.
375 2018), there was no increase of agricultural lands in the proximity of ponds and lakes in the
376 present study. In most cases, the landscape remained natural with even a decrease of some
377 agricultural typologies (i.e. agriculture mosaics), a sign of land abandonment and reduced use of

378 natural resources (see also Nyssen et al. 2014, Lakovic et al. 2016). However ponds, which were
379 historically used primarily for water sources for humans and livestock are now often redirected
380 to new objectives, such as fish stocking. This exemplifies a major and worrying change of
381 thinking on the value and use of water resources in rural environments which can have
382 significant negative impacts on biodiversity.

383

384 *Declines over the last decades*

385 Previous reports highlighted the decline of amphibians, including newts, at the global scale,
386 (Denoël 2012, Dufresnes and Perrin 2015, Drechsler et al. 2016, Arntzen et al. 2017), and
387 paedomorphic newts have suffered particularly dramatic population losses (Denoël et al. 2005a).
388 In Montenegro, the first report of paedomorph extirpation dates back to the 1980's (Breuil 1985),
389 whereas the first global survey in the early 2000's showed a strong decline which altered the
390 distribution of paedomorphs, but caused no changes for metamorphs (Denoël et al. 2009). Our
391 results show that paedomorphs are now only present in one-fifth of the historical sites. Only a
392 few paedomorphs were observed in the remaining populations, suggesting that the large
393 historical populations no longer exist. With a maximum of 13 paedomorphs found at a single site
394 (all populations considered) in 2016, the situation contrasts radically with what existed more than
395 30 years ago, when populations of up to one thousand paedomorphs could be found in ponds and
396 in lakes (Džukić 1981, Kalezić and Džukić 1985, Kalezić and Džukiž 1986, Kalezić et al. 1989).
397 The proportion of paedomorphs within populations also dropped significantly. Indeed, some
398 populations were predominantly paedomorphic in historical surveys (e.g. 86% paedomorphs in
399 Bukumirsko jezero: Kalezić et al. 1989), whereas paedomorphs were either extinct or found in
400 very low proportions during the latest survey (i.e. maximum 7% in Rutešića voda), therefore

401 showing an extinction of all populations which were mainly paedomorphic. We also point out
402 that the observed declines are not due to turn-over in metapopulations (Cruickshank et al. 2016)
403 because there were no potential population source of paedomorphs nearby the study populations.

404 The present survey showed a much more serious decline than in other diversity hotspots
405 where both phenotypes declined but persisted (Crochet et al. 2004, Denoël et al. 2005a, Denoël
406 and Winandy 2015). Countries near Montenegro were also known to host populations of
407 paedomorphic newts (Džukić et al. 1990, Denoël et al. 2001, Papaioannou et al. 2015,
408 Sotiropoulos et al. 2017). However, for the alpine newts, the most important populations also
409 disappeared in Bosnia and Herzegovina and Slovenia (Denoël et al. 2005a). Some of the extinct
410 paedomorphic populations were unique and were historically described as endemic subspecies
411 (Radovanović 1951, 1961). This is the case of *Ichthyosaura alpestris montenegrina* in
412 Bukumirsko Jezero, *I. a. serdara* in Zminičko Jezero and *I. a. piperiana* in Kapetonovo and
413 Manito Jezero. In all cases, paedomorphs disappeared from the type localities, whereas
414 metamorphs also vanished or largely declined (see also Džukić 1995, Denoël et al. 2005a). Such
415 continuous decline overcomes thresholds defined by the International Union for the
416 Conservation of Nature (IUCN) to identify critically-endangered taxa (IUCN 2017). The fact that
417 localised extinctions have also been reported for lakes inhabited by paedomorphs in other
418 families, such as ambystomatids, including the axolotl (*Ambystoma mexicanum*), which is
419 endemic to a single lake and is now nearly extinct (Whiteman and Howard 1998, Contreras et al.
420 2009), further stresses the high conservation concern of paedomorphosis.

421

422

Anthropogenic pressures on freshwater habitats and amphibians

423 Multiple factors drive biodiversity loss on the global scale. Invasive species, land use conversion
424 and climate change are among the most threatening factors (Sala et al. 2000, Bosch et al. 2018).
425 The introduction of non-native species, particularly fish, was the main driver of the decline of
426 both metamorphic and paedomorphic newts (Fig 4). Fish stocking started in alpine lakes of
427 Montenegro in the 1970's, and continued in the last decades, particularly in ponds. Aquatic
428 conditions were often adequate for many fish species, and the presence of young-of-the-year
429 indicated reproduction at several sites. The lakes which were already fished in the early 2000's
430 continued to be stocked with fish, and multiple species were often introduced in the same lake.
431 Introductions involved fish as varied as cyprinids, catfish and salmonids, with up to six species
432 found in a single lake. The presence of such alien fish communities is very detrimental to native
433 species because they are expected to occupy varied niches of the freshwater habitats and
434 profoundly impact native food webs (Havel et al. 2015, Gallardo et al. 2016, Cabrera-Guzmán et
435 al. 2017). For instance, the larger and more predatory species can impact newts by predation, as
436 shown by the stomach contents of 11 salmonids, which altogether contained 15 paedomorphic
437 alpine newts at the time of introduction in Bukumirsko Jezero (Denoël et al. 2005a; pers. obs.).
438 In addition to consumptive interactions, fish can also be competitors and be aggressive towards
439 newts and salamanders (Zambrano et al. 2010, Winandy and Denoël 2015). An interesting
440 pattern evidenced by the present study is that metamorphs manage to persist in ponds and lakes
441 longer than paedomorphs in the presence of fish (Fig 3). This differential response of both
442 phenotypes could be explained by different factors. First, the fully aquatic life of paedomorphs
443 exposes them to fish all year round and prevents them from leaving water bodies with fish by
444 dispersal on land, whereas metamorphs are able to do so (Winandy et al. 2015). Second, it has
445 been suggested that metamorphs in water bodies with fish could come from nearby fishless ones,

446 i.e. through a source-sink dynamic process (Breuil 1985). Third, paedomorphic newts are more
447 pelagic than metamorphs and, therefore, are more easily preyed upon by large predators such as
448 salmonids (Denoël et al. 2005b, Lejeune et al. 2018). Unfortunately, the longer persistence of
449 metamorphs proved to be, in most cases, only a transient state, given that in the last years they
450 showed a pattern of decline comparable to the one of paedomorphs in the 2000's.

451 In addition to fish, the last survey detected crayfish, which are a new threat to newts.
452 Although not seen historically and found in only one pond a dozen of years ago (Ćirović 2009;
453 pers. obs.), crayfish have now been introduced in four localities, including both ponds and lakes.
454 Crayfish are amongst the most important invasive components of freshwater ecosystems
455 worldwide, with the potential to have significant detrimental effects on whole communities due
456 to their omnivorous feeding behaviour and intense burrowing activity (Rodríguez et al. 2005,
457 Gherardi 2010). Their recent invasion in numerous countries raises important concerns for
458 amphibian and, even more, newt conservation as crayfish can greatly affect their survival
459 (Gamradt and Kats 1996, Cruz et al. 2006, Ficetola et al. 2011, Ficetola et al. 2012). The effect
460 of crayfish in Montenegro was particularly detrimental as shown by the disappearance of both
461 newt phenotypes following crayfish introduction.

462 Finally, mixed models showed that climate change affected the metamorphic phenotype
463 in the studied populations. Although drought events can affect populations of paedomorphs in
464 other areas (Semlitsch and Gibbons 1985, Denoël and Winandy 2015, Mathiron et al. 2017), it is
465 less likely the case in Montenegro because most ponds and lakes are very deep, making them
466 permanent aquatic habitats. In contrast, high temperatures and low precipitation were associated
467 with a decrease in the occurrence of the metamorphs, the terrestrially-adapted phenotypes.
468 Despite the fact that microhabitats could buffer the impact of climate change (Scheffers et al.

469 2014), this result comprises recent evidence that global warming could affect amphibian
470 populations (Walls 2009, Bonett et al. 2014, Ficetola and Maiorano 2016).

471

472 *Is there still hope for amphibians and paedomorphosis?*

473 Despite perturbations and declines, many species are able to show resilience after threat removal
474 (Knapp et al. 2001). This is the case in amphibians (Vredenburg 2004, Knapp et al. 2007, Knapp
475 et al. 2016, Milligan et al. 2017), and even paedomorphic newts, after fish removal (Denoël and
476 Winandy 2015). Indeed, as facultative paedomorphosis is a polyphenism, some metamorphic
477 dispersers from populations where facultative paedomorphosis is expressed can give birth to
478 individuals which can mature as paedomorphs (Denoël and Winandy 2015, Oromi et al. 2016).
479 However, selection against paedomorphosis can alter its expression (Semlitsch and Wilbur
480 1989), which might decrease the likelihood of the resilience of paedomorphic populations after a
481 long period of time of counter selection, as is probably the case in Montenegro. Moreover, as
482 there are no remaining populations of paedomorphs nearby to the studied populations, colonizers
483 would come only from populations of metamorphs (see also Oromi et al. 2018). In fact, fish were
484 extirpated in one of the mountain lakes (Ridsko Jezero), but paedomorphs did not recover as only
485 metamorphic newts were found during the last surveys.

486 Fish removal could be performed in both ponds and lakes, but this often requires long-
487 term actions (Knapp and Matthews 1998, Tiberti et al. 2017). These management aspects were
488 therefore not in the scope of the present study whose primary aim was to show the ongoing
489 decline and to raise awareness that the extinction of paedomorphs, and consequently of the
490 process itself, is likely to occur globally if no conservation measures are taken. In developing

491 countries, such as Montenegro, management may be particularly challenging. On the one hand,
492 the recent fishing interest of local people and the abandonment of the traditional use of ponds
493 (i.e. providing water sources for humans and cattle) make it likely that new introductions will
494 occur after management. On the other hand, no significant management activities or funding
495 have been devoted as yet to freshwater biodiversity conservation in Montenegro. Thus, local
496 enforced protection along with awareness raising should be promoted in parallel with all removal
497 actions. For instance, fishing bans helped in preventing new fish introductions in Pyrenean alpine
498 lakes (Miró and Ventura 2013).

499 It is possible that some new populations of paedomorphs could still be discovered.
500 However, all major alpine lakes have already been intensively surveyed in Montenegro (Denoël
501 et al. 2009); therefore, it is likely that no new major populations would be found. In principle, it
502 is also possible that paedomorphosis could still be expressed later in some of the studied ponds,
503 despite the absence of observation during the last survey. However, only direct management
504 through fish removal could help to restore populations as ponds and lakes are usually permanent.

505

506

Conclusions

507 Land use and running waters are targeted by major legislation and funding programmes
508 (EU_Water_Framework_Directive 2000, Young et al. 2005, Rounsevell et al. 2006). However,
509 small lentic freshwater habitats, such as ponds and mountain lakes are often disregarded (Nicolet
510 et al. 2009, Oertli et al. 2010, Boix et al. 2012, Sayer 2014). As shown here, freshwater habitats
511 can show a higher degradation than terrestrial landscapes, even within areas which are
512 considered traditionally managed. In particular, introduced species were shown in the present

513 case to have a stronger effect on amphibian decline and on the almost extinct emblematic
514 paedomorphic phenotype than be those of climate change (see also Ficetola et al. 2018).
515 Therefore, it is urgent that ponds and small lakes receive as much attention as large lakes and
516 rivers in conservation policy (Davies et al. 2008). Such a target would need a better recognition
517 of these habitats for preserving intra and interspecific biodiversity value and an improvement of
518 the ecosystemic value of freshwater habitats without alien fish introductions.

519

520

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533

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851 FIG. 1. Paedomorphic alpine newt (A, *Ichthyosaura alpestris*) and Greek smooth newt (B,
852 *Lissotriton graecus*). This phenotype is characterized by the retention of larval traits, such as
853 gills, at the adult stage. Both pictures depict a male. Photos: Mathieu Denoël.

854

855 FIG. 2. Amphibian distribution across time: Historical (1951-2004), intermediate (2002-2005)
856 and recent (2016) distribution of paedomorphic newts in Montenegro. Circles: alpine newts
857 (*Ichthyosaura alpestris*), squares: Greek smooth newts (*Lissotriton graecus*). Black symbols:
858 persistence of paedomorphosis until the last sampling, grey symbols: persistence until the
859 intermediate surveys and open symbols: disappearance before the intermediate survey.
860 Background: Corine Land Cover 2012 (dark green: natural vegetation, light green: mosaics
861 between natural and agricultural vegetation, yellow: agriculture, red: urbanisation, blue: water).

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864 FIG. 3. Main local threats to freshwater habitats (lakes and ponds) of Montenegro: (a) proportion
865 of sites with introduced fish and astacid crayfish and those destroyed; (b) proportion of sites
866 fished with cyprinids, salmonids and ictalurids.

867

868 FIG. 4. Decline of paedomorphic and metamorphic alpine newts and Greek smooth newts in
869 traditional landscapes of Montenegro. (a-b): Mean observed frequency \pm SE; (c-d) Frequency

870 fitted using generalized mixed models. Blue: paedomorphs, red: metamorphs. Shaded areas
871 represent 95% Confidence Intervals.

872

873 FIG. 5. Global change in the amount of habitat use in Montenegro by paedomorphic alpine newts
874 (blue bars) and Greek smooth newts (red bars): (a) extent of occurrence, (b) area of occupancy
875 (based on 2 x 2 km grid cells) and (c) aquatic area of occupancy (based on surface area of water
876 bodies) at the historical maximum (all known populations), in the 2000's and in the 2010's.

877

878 FIG. 6. Bayesian mixed models of the effect of environmental and climatic variables on the
879 decline of paedomorphic and metamorphic newts. Whiskers denote 95% Credible Intervals.

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884 Fig 1.

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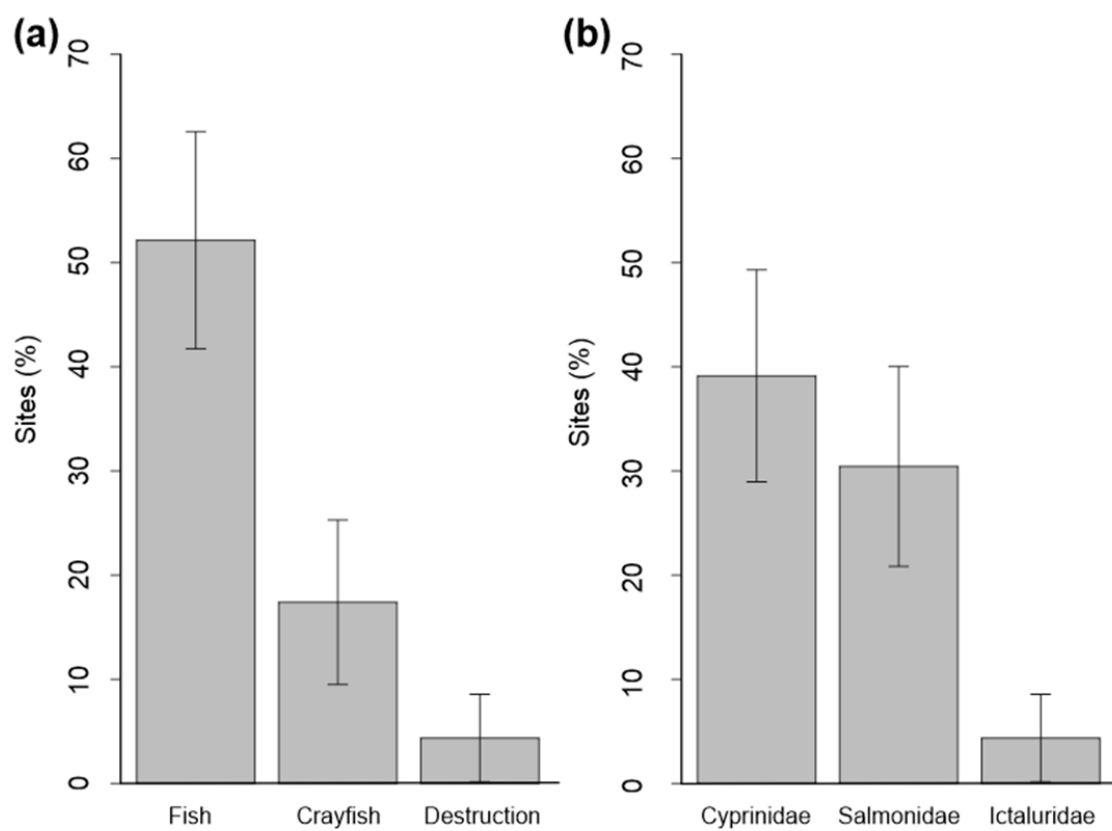
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Fig 3.



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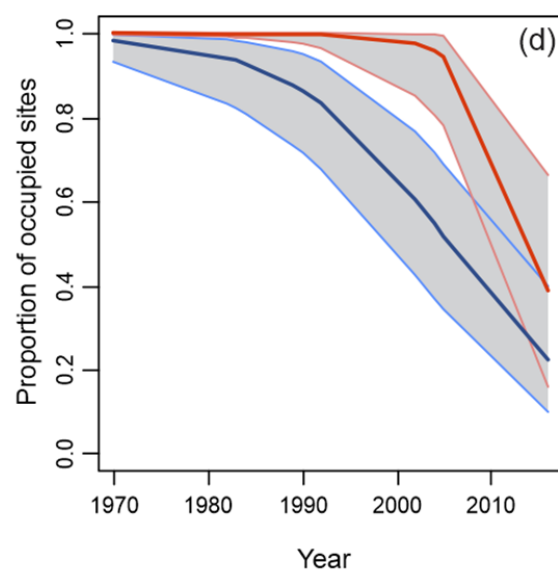
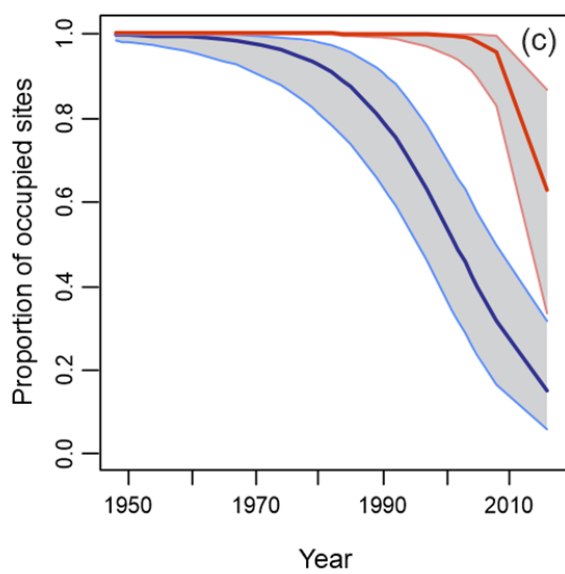
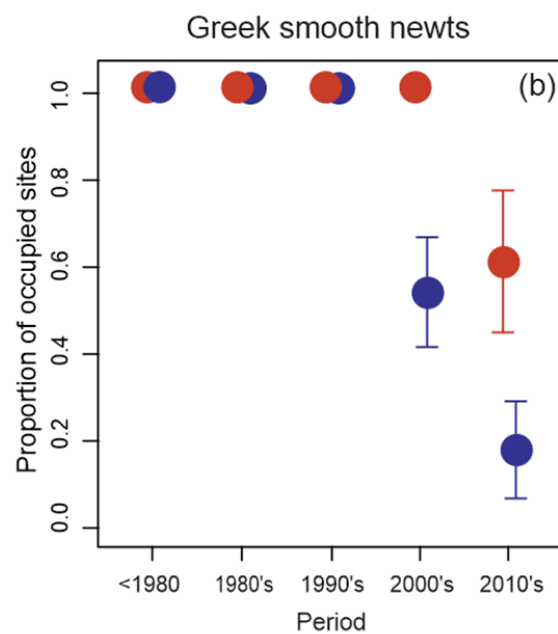
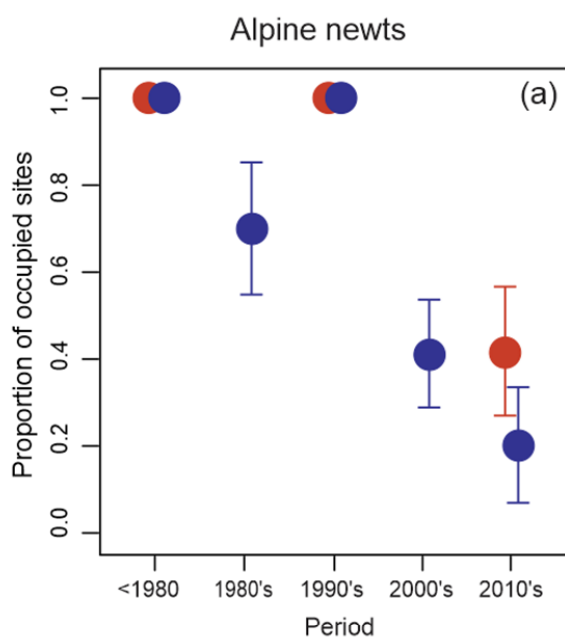
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917 Fig 4.



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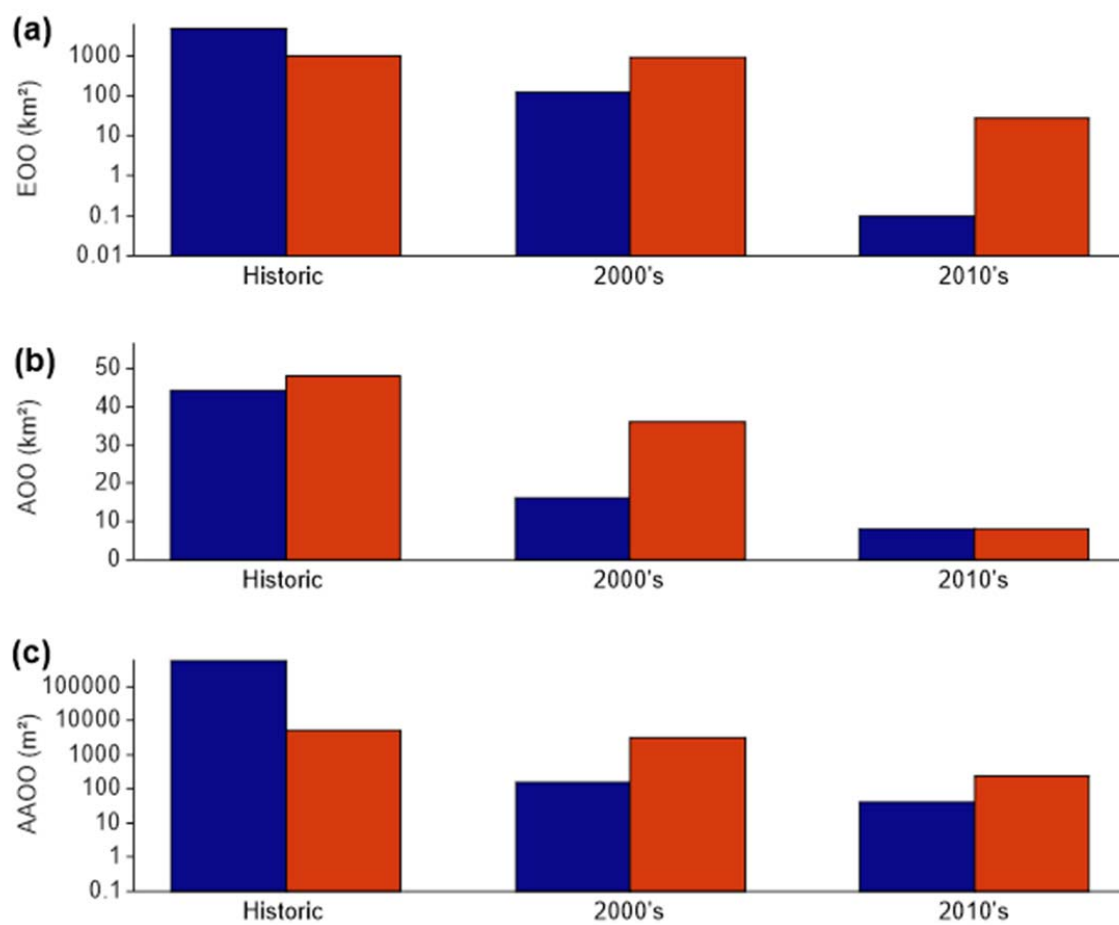
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926 Fig 5.



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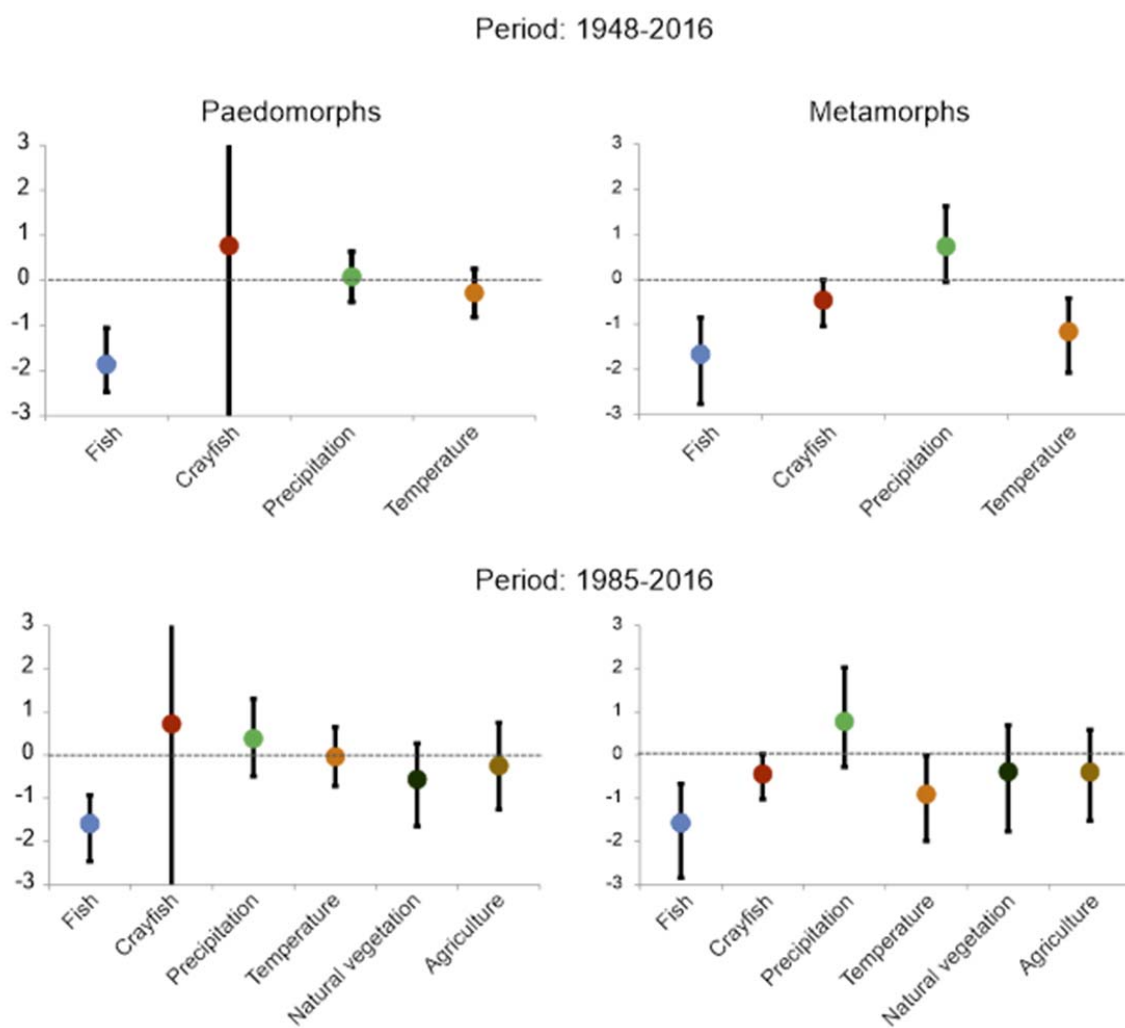
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939 Fig 6.



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SUPPORTING INFORMATION

APPENDIX S1

950

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952

953 TABLE S1. Location and abiotic aquatic habitat features of the populations of paedomorphic
954 alpine newts (*Ichthyosaura alpestris*) and Greek smooth newts (*Lissotriton graecus*) in
955 Montenegro.

956

957 TABLE S2. Historical and recent occurrence of paedomorphosis in the alpine newts (*Ichthyosaura*
958 *alpestris*) and Greek smooth newts (*Lissotriton graecus*) in Montenegro.

959

960 TABLE S3. Examined specimens of alpine newts (*Ichthyosaura alpestris*) and Greek smooth
961 newts (*Lissotriton graecus*) in the collection of the IBISS Museum (Institut za Biološka
962 Istraživanja 'Siniša Stanković', Belgrade).

963

964 TABLE S4. Corine Land Cover surrounding ponds and lakes with records of paedomorphosis in
965 alpine newts (*Ichthyosaura alpestris*) and Greek smooth newts (*Lissotriton graecus*) in
966 Montenegro.

967

968 TABLE S5. Alteration of the aquatic habitat in the populations of paedomorphic alpine newts
969 (*Ichthyosaura alpestris*) and Greek smooth newts (*Lissotriton graecus*) in Montenegro:
970 occurrence of introductions of fish and crayfish, and habitat destruction.

971

972 TABLE S6. Occurrence (proportion of occupied sites) of paedomorphs and metamorphs across
973 time in the studied localities: data categorized by decades for the populations of alpine newts
974 (*Ichthyosaura alpestris*) and Greek smooth newts (*Lissotriton graecus*) in Montenegro.

975

976 TABLE S7. Results of generalized mixed models on the occurrence through time of paedomorphic
977 and metamorphic alpine newts (*Ichthyosaura alpestris*) and Greek smooth newts (*Lissotriton*
978 *graecus*) in Montenegro. Models include species identity as an additional fixed factor. CI:
979 Credible Intervals.

980

981 FIG S1. Extinct paedomorphs, historically described as endemic subspecies of alpine newts of
982 Montenegro. (a) *Ichthyosaura alpestris montenegrinus* from Bukumirsko Lake, (b) *I. a.*
983 *piperiana* from Kapetanovo Lake and (c) *I. a. serdara* from Zminičko Lake. Specimens
984 examined in the museum collection of the Institut za Biološka Istraživanja ‘Siniša Stanković’,
985 Belgrade, Serbia (reference 2422, 1502: holotype, 1479: holotype, respectively). All pictured
986 specimens are females showing developed gills and mature cloaca. Photos: Mathieu Denoël.

987

988 FIG. S2. Large variety of freshwater lentic habitats of Montenegro in which alien species
989 (specifically fish) were introduced and in which the large populations of paedomorphic newts
990 fully vanished, whereas metamorphs either declined or got extirpated: (a) Bukumirsko, (b)
991 Zminičko, (c) Kapetanovo (in the background) and Manito Lakes, respectively, the type

992 localities of the alpine newt *Ichthyosaura alpestris montenegrina*, *I. a. serdara* and *I. a.*
993 *piperiana*, and (d) Velika Osjecenica pond, historically used by the Greek smooth newt
994 *Lissotriton graecus*. Photos: M. Denoël (2016).

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997 FIG S3. Maps of Montenegro with Corine Land Cover across time. (a) 1990, (b) 2000, (c) 2006
998 and (d) 2012. Dark green: natural vegetation, light green: mosaics between natural and
999 agricultural vegetation, yellow: agriculture, red: urbanisation and blue: water; black dots: studied
1000 populations (historical presence of paedomorphosis in newts).

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1002

1003 FIG S4. Climatic variations (precipitations and temperature) in the studied localities of
1004 paedomorphosis in the alpine newts (*Ichthyosaura alpestris*) and Greek smooth newts
1005 (*Lissotriton graecus*) in Montenegro. Climatic data were extracted from the CRU 4.01 climate
1006 grids (Harris, Jones, Osborn, & Lister, 2014). Box plots show the median (dark midline),
1007 percentiles 25–75 and minimum-maximum value not higher than 1.5 IQR.