

LETTER

Eco-evolutionary drivers of avian migratory connectivity

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Abstract

Migratory connectivity, reflecting the extent by which migrants tend to maintain their reciprocal positions in seasonal ranges, can assist in the conservation and management of mobile species, yet relevant drivers remain unclear. Taking advantage of an exceptionally large (~150,000 individuals, 83 species) and more-than-a-century-long dataset of bird ringing encounters, we investigated eco-evolutionary drivers of migratory connectivity in both short- and long-distance Afro-Palaearctic migratory birds. Connectivity was strongly associated with geographical proxies of migration costs and was weakly influenced by biological traits and phylogeny, suggesting the evolutionary lability of migratory behaviour. The large intraspecific variability in avian migration strategies, through which most species geographically split into distinct migratory populations, explained why most of them were significantly connected. By unravelling key determinants of migratory connectivity, our study improves knowledge about the resilience of avian migrants to ecological perturbations, providing a critical tool to inform transboundary conservation and management strategies at the population level.

KEYWORDS

avian migration, bird conservation, Convention on Migratory Species, EURING, European-African migrants, migration ecology, migratory connectivity, mobile species, ringing encounters

INTRODUCTION

Animal migration is a widespread phenomenon which has evolved as an adaptive response to spatiotemporal variations in resources (Dingle & Drake, 2007). Ongoing environmental changes are disrupting migration strategies (Romano et al., 2023) and threatening migratory populations at the global scale, with dramatic declines in the abundance of migratory species and an

increasing number of formerly migratory populations becoming resident (Buchan et al., 2020; Morganti, 2015; Pulido & Berthold, 2010; Robinson et al., 2009; Runge et al., 2014; Shaw, 2016; Visser et al., 2009; Wilcove & Wikelski, 2008). Migratory connectivity quantifies how strongly individuals tend to maintain their reciprocal positions between seasonal ranges where they spend different phases of their annual cycle (Webster et al., 2002). Strong migratory connectivity occurs when

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individuals tend to maintain their reciprocal positions in seasonal ranges, implying limited population mixing, whereas weak migratory connectivity reflects spatial rearrangement of individuals, indicating higher population mixing between breeding and nonbreeding ranges (Webster et al., 2002). This ecological property offers the possibility to assist in the conservation of migratory species ranging from invertebrates (Gao et al., 2020) to marine mammals (Dunn et al., 2019) and has become critical to understand how ecological perturbations in the nonbreeding range may affect the fitness and population dynamics of migrants (Taylor & Norris, 2010). Particularly among birds, strongly connected migratory populations have been suggested to be more vulnerable to environmental changes because any differential environmental change in the nonbreeding grounds may affect an entire (sub)population (Ambrosini et al., 2019; Briedis & Bauer, 2018). Conversely, the negative consequences of environmental alterations may be buffered in loosely connected populations because only part of a breeding population would experience such changes (Ambrosini et al., 2019; Briedis & Bauer, 2018). Hence, understanding the processes affecting the population dynamics of migratory species requires improved knowledge of the ecological and evolutionary causes of migratory connectivity (Beresford et al., 2019; Boulet & Norris, 2006; Patchett et al., 2018).

Although bio-energetic principles and geographical effects that affect the strength of avian migratory connectivity have been identified (Finch et al., 2018; Somveille et al., 2021), the relevant eco-evolutionary drivers are poorly understood. In particular, no interspecific comparative analysis has been conducted for disentangling the potential evolutionary and ecological drivers of the extent of migratory connectivity in different species. Some temporal and spatial avian migration patterns may be inherited and are generally under strong selective pressures (Åkesson & Helm, 2020; Gu et al., 2021; Liedvogel et al., 2011), leading to hypothesize the existence of a phylogenetic signal on the strength of migratory connectivity because selective pressures and migration costs can be shared between closely-related species. Additionally, since avian migration is largely influenced by life-history traits such as niche specialization (Reif et al., 2016; Romano et al., 2023) and body mass-dependent energetic costs of aerial locomotion (Hein et al., 2012), these and other similar ecological factors could play critical roles in determining how birds redistribute between breeding and nonbreeding ranges.

Here, we took advantage of an over-a-century-long collection of ringing encounters of European breeding birds to investigate the eco-evolutionary determinants of avian migratory connectivity. To our knowledge, this study involves the largest dataset ever used to investigate migratory connectivity in the animal kingdom. We worked at the scale of geographical populations along a *continuum* of migratory behaviours ranging from short- to long-distance

migrants, and from partially to fully migratory species, testing a set of hypotheses aimed at explaining intra- and interspecific variation in the strength of connectivity. First, we determined migratory connectivity for a larger number of species and populations compared to previous studies. Then, we tested for phylogenetic, geographic and life-history effects on migratory connectivity.

Despite the available theory predicts well the observed patterns of avian migratory connectivity strength at continental scales (Finch et al., 2018; Somveille et al., 2021), multi-species information based on empirical data collected within the European-African migration system is scarce. Over the last decades, European populations of migratory birds, especially long-distance ones, have declined substantially, likely due to habitat loss or deterioration in the nonbreeding grounds (Beresford et al., 2019; Howard et al., 2020; Sanderson et al., 2006; Vickery et al., 2014). Migratory connectivity seems to play a key role in such declines (Patchett et al., 2018), and previous studies call for gaining knowledge on how birds mix between breeding and nonbreeding grounds to assist the conservation of European migrants (Beresford et al., 2019). Furthermore, long-distance migrants in the European-African system seem to have already responded to past climatic perturbations occurring in Africa through the evolution of low migratory connectivity as a bet-hedging adaptation (Cresswell, 2014; Patchett et al., 2018).

We focused on the Afro-Palaearctic migration system and used data concerning ~150,000 individuals from 83 species and 32 avian families. Our predictions originate partly from the evidence provided by earlier research, but also extend toward unexplored effects of life-history traits stemming from ecological theory, and are summarized in [Table 1](#) (see also [Appendix SI](#), for details). More specifically, our hypotheses about the effects of geographic drivers hinge on the biological principle of optimizing both energy expenditure (migration cost) and energy acquisition (migration benefit) by which individuals should redistribute in seasonal ranges in the most energy-efficient way, following the theory of avian migratory connectivity (Somveille et al., 2021). According to the optimization of migration cost, we predicted a weaker migratory connectivity for species having a farther and smaller nonbreeding range ([Table 1](#); [Appendix SI](#)). By unravelling eco-evolutionary determinants of avian migratory connectivity, we aim at improving our understanding of the resilience of migratory birds to ecological perturbations, providing a critical tool to inform transboundary conservation and management strategies.

MATERIALS AND METHODS

Migratory connectivity analysis

Initially, we assessed the strength of migratory connectivity in 137 bird species by filtering more than 12 million ringing encounters obtained from the EURING

TABLE 1 Hypotheses about the potential eco-evolutionary drivers of avian migratory connectivity. See Appendix S1, for an extended explanation.

Eco-evolutionary driver	Hypothesis and prediction	Supporting references
Geographical		
Migration distance	Migration distance is a proxy for the cost of relocating between seasonal grounds, with the shortest available path corresponding to the travel cost optimization. When migration distance increases, migration cost is more similar between alternative migration paths; in turn, more mixing (lower connectivity) is likely to occur in the population (Figure S1, Appendix S1). Empirical data indeed suggested low connectivity in long-distance migrants. Thus, migratory connectivity is expected to decrease with migration distance.	Gilroy et al. (2016); Finch et al. (2018); Somveille et al. (2021)
Land availability (i.e., nonbreeding latitude, for Afro-Palearctic migrants)	As connectivity increases for populations with larger availability of suitable land in their nonbreeding range, geographical constraints related to the shape of the African continent should force populations of Afro-Palearctic migrants to mix more at the southernmost latitudes (where smaller land mass is available), showing lower connectivity. Similarly, for populations overwintering in Europe, a stronger connectivity could arise from their segregation in the different peninsulas of southern Europe or in Mediterranean Africa (where larger land mass is available). Thus, in the European-African migration system, connectivity is expected to decrease with an increasing propensity to winter farther south.	Finch et al. (2018)
Nonbreeding population spread	The nonbreeding population spread is the mean inter-individual pairwise distance in the nonbreeding population range and is a proxy of how close individuals spread out, on average, in their nonbreeding range. While accounting for migration distance, spreading over a larger nonbreeding range should minimize the cost of migration (Figure S1, Appendix S1), thus lower mixing (stronger connectivity) is likely to occur. Empirical data also supported a stronger migratory connectivity for species with larger nonbreeding range spread. Therefore, migratory connectivity is expected to increase with increasing nonbreeding population spread. Despite breeding population spread may exert a similar effect, it is not considered here because it would be redundant when both nonbreeding population spread and the ratio between breeding and nonbreeding population spread are considered.	Gilroy et al. (2016); Finch et al. (2018); Sarà et al. (2019); Somveille et al. (2021)
Relative population spread	The relative population spread is the ratio between the breeding population spread and the nonbreeding population spread, and it represents the extent to which populations occupy larger or smaller nonbreeding ranges relative to the range occupied in the breeding period (e.g. relative population spread = 1 corresponds to breeding and nonbreeding ranges of the same size). Theoretically, individuals should mix more (i.e., migratory connectivity should decrease) when the nonbreeding spread is smaller than the breeding one. However, mixing of individuals can be similarly promoted also when the breeding spread is smaller than the nonbreeding spread. Thus, a quadratic effect of relative population spread on migratory connectivity is expected, which should peak when it approaches 1.	A corollary of the hypothesis on nonbreeding population spread
Life-history		
Body mass	Although bird species with a larger body mass tend to migrate farther, which may lower connectivity, they also tend to live longer, which should foster the social transmission of knowledge about routes and nonbreeding sites across generations and thus increase migratory connectivity. Once migration distance is accounted for, larger species should be those with the stronger connectivity.	Hein et al. (2012); Teitelbaum et al. (2016); Foss-Grant et al. (2018)
Habitat breadth	In birds, habitat generalists are more likely to exhibit migration propensity or to migrate longer distances. Habitat specialists would not be advantaged to spread over different nonbreeding ranges. A lower habitat breadth could thus act as a driver of stronger migratory connectivity.	Reif et al. (2016)

(Continues)

TABLE 1 (Continued)

Eco-evolutionary driver	Hypothesis and prediction	Supporting references
Geographical		
Dietary breadth	In birds, dietary breadth is strongly correlated with habitat breadth. Similarly to habitat specialists, species with a narrower diet breadth could not receive selective benefits in spreading over different nonbreeding grounds, which would suggest relatively stronger migratory connectivity.	Reif et al. (2016)
Passerine or non-passerine	Passerine juveniles usually migrate separately from adults, whilst non-passerine juveniles tend to follow adults during their first migration. The former thus tends to redistribute stochastically over a wider nonbreeding area than the latter, due to the larger unpredictability of conditions experienced upon the first migration (serial residency hypothesis). Therefore, passerines could be expected to retain lower migratory connectivity than non-passerines.	Cresswell (2014)
Evolutionary		
Phylogeny	Migration patterns shaping migratory connectivity such as route choice and nonbreeding target destinations may be genetically inherited and often are under strong selective pressures. They also mirror cost optimization pathways and selective pressures that could be more similar in closely-related species. Thus, a phylogenetic signal in the strength of migratory connectivity could be expected.	Åkesson and Helm (2020); Gu et al. (2021)

databank (du Feu et al., 2016; <https://euring.org>) and spanning over more than one century (1900–2019), as described in Appendix S2. Selected data included 371,090 individuals (range: 20–36,506 individuals/species; mean \pm SE: 2708 \pm 480 individuals/species) encountered between 1917–2019.

For each species, we assessed the strength of migratory connectivity according to Ambrosini et al. (2009). The method was developed using ringing encounters and has been used consistently to estimate migratory connectivity at various geographical scales, and for both single- and multi-species analyses (e.g., Ambrosini et al., 2016; Finch et al., 2018; Knight et al., 2021; Sarà et al., 2019; Somveille et al., 2021). Using individual locations in breeding and nonbreeding grounds, we calculated two seasonal distance matrices. The strength of migratory connectivity was quantified through the Mantel correlation coefficient (r_M) between seasonal matrices, whereby a strong correlation between the matrices indicates that individuals tend to maintain reciprocal positions, avoiding seasonal mixing. The probability of a positive connectivity was tested by a one-tailed permutation test (Ambrosini et al., 2009) because negative values in the strength of migratory connectivity are not biologically meaningful (Cohen et al., 2018). A bootstrap procedure was also used to estimate the 95% CIs of r_M values (Cohen et al., 2018). For those species showing significant connectivity, eight k-mean cluster analyses (pre-defined number of clusters: 2–9) were performed to identify clusters of individuals that tend to gather in separate groups in the breeding and nonbreeding ranges (Ambrosini et al., 2009). The best clustering structure was identified as that maximizing the overall average silhouette width (*oasw*; Rousseeuw, 1987), a metric showing the best performance among clustering validity indices (Arbelaitz et al., 2013). Species with a strong clustering structure

(*oasw* \geq 0.5) suggest the presence of geographical populations within the same species that differ in their migration strategy, showing a distinct combination of seasonal ranges (Figure 1a,b). Thus, for species showing a strong clustering structure, we re-calculated r_M on each cluster separately and used these values as migratory connectivity estimates rather than the value calculated using all the data available for that species. For species that were not spatially clustered into distinct migratory clusters (*oasw* $<$ 0.5), we considered all individuals as belonging to a single geographical population (Figure 1c,d), independently of the significance of r_M . Hence, we refer to a ‘geographical population’ as either (1) the ensemble of all the ring encounters of a species, if the species showed non-significant connectivity or a weak clustering structure by which all individuals can be considered to belong to the same migratory population, or (2) each of the clusters identified by the migratory connectivity analysis if the species showed a significant connectivity and a strong clustering structure.

Birds show large inter-individual variability in migration propensity, often including both migratory and resident phenotypes even within the same population (Chapman et al., 2011; Gilroy et al., 2016). Thus, a biologically representative investigation of the drivers of avian migratory connectivity should be better pursued across the ‘migratoriness’ continuum, including populations with co-occurrence of migratory and non-migratory phenotypes, as well as spanning a wide range in migration distance. We classified the geographical populations identified by the migratory connectivity analysis as strongly or weakly migratory based on the extent of the overlap of their breeding and nonbreeding ranges, which showed a bimodal distribution (Appendix S3). Then, we discarded the species without strongly migratory populations in our dataset, as they are usually considered sedentary species

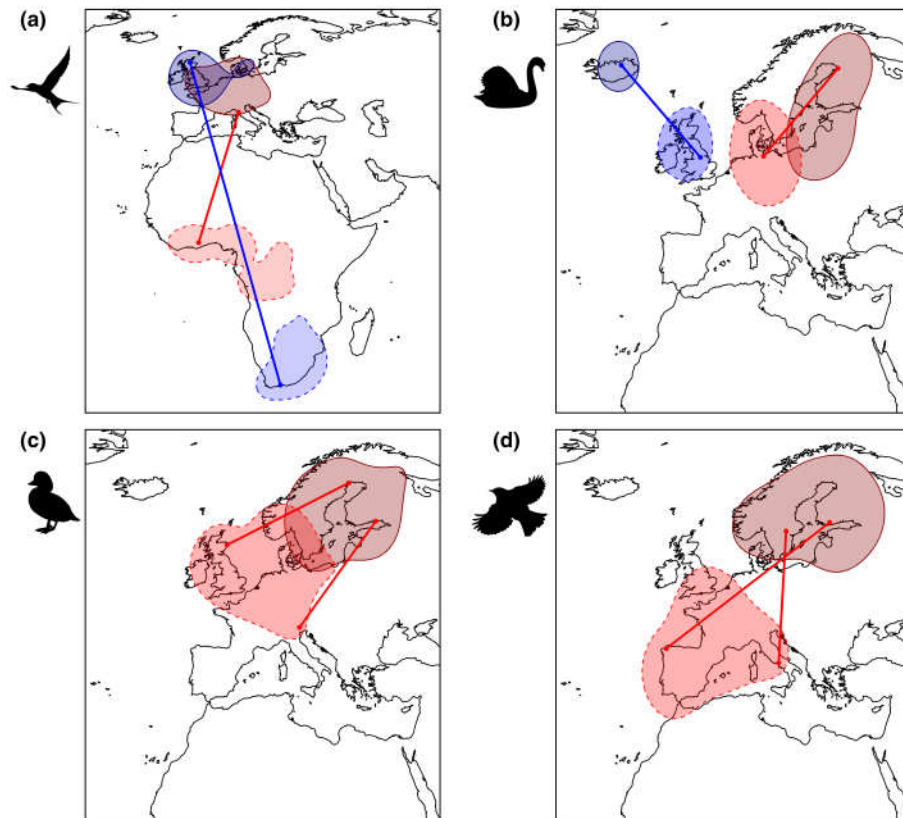


FIGURE 1 Migratory connectivity maps of representative bird species showing, at the species level (a) weak connectivity and strong clustering (barn swallow *Hirundo rustica*: $r_M=0.15$, $p=0.004$, $n=96$, $oasw=0.61$); (b) strong connectivity and strong clustering (whooper swan *Cygnus cygnus*: $r_M=0.87$, $p<0.001$, $n=509$, $oasw=0.76$); (c) moderate connectivity and weak clustering (common goldeneye *Bucephala clangula*: $r_M=0.40$, $p=0.003$, $n=69$, $oasw=0.34$); (d) no connectivity (redwing *Turdus iliacus*: $r_M=0.03$, $p=0.13$, $n=683$). Lines connect individual breeding sites and nonbreeding destinations of two randomly selected individuals per species, while differently coloured kernel areas encompass 95% ring encounters of the relevant geographical populations (solid contour: breeding, dotted contour: nonbreeding). Migratory connectivity maps for all the species and populations analysed here can be accessed at <https://migrationatlas.org/> (Spina et al., 2022). Species silhouettes are taken from <http://phylopic.org> (see Acknowledgments for credits).

in the Afro-Palearctic migration system. Additionally, we only retained populations having ≥ 30 individuals re-encountered, because lower samples may not provide robust connectivity estimates (Ambrosini et al., 2009). Our final dataset included 150,909 individuals (191 populations of 83 species; range: 30–27,479 individuals/species and 1–9 populations/species; mean \pm SE: 1818 ± 429 individuals/species and 2.30 ± 0.21 populations/species; Appendix S4). In a second step, we re-ran the analyses by removing weakly migratory populations, retaining 120,377 individuals (150 populations of 83 species).

Phylogenetic comparative analysis

A phylogenetic comparative analysis was conducted using the *metafor* R package (Viechtbauer, 2010). Because r_M values are correlation coefficients, they were transformed into Z_r using Fisher transformation. We then fitted a phylogenetic mixed model where the variance components of the random part allow calculating how much variance is attributable to the phylogeny (phylogenetic heritability, H^2)

while considering multiple Z_r values for the same species and accounting for the fixed effects included in the model (H^2 is equivalent to Pagel's λ ; Nakagawa & Santos, 2012). H^2 was calculated as the ratio between the variance due to phylogeny and all the variance components in the model (Nakagawa & Santos, 2012). To account for phylogeny, we built a 50% majority rule-consensus tree using 10,000 phylogenetic trees (Hackett et al., 2008) retrieved from www.birdtree.org, as recommended for avian comparative studies (Rubolini et al., 2015).

In a second model, we included as fixed effects a set of moderators that may influence migratory connectivity according to our hypotheses (Table 1). Geographical predictors were calculated using the positions of the ringing encounters used in the analyses. For each population, we considered the mean (orthodromic) migration distance, the mean nonbreeding latitude (measured as positive or negative degrees from the equator), the nonbreeding population spread (mean inter-individual pairwise distance in the nonbreeding population range; Finch et al., 2018) and both the linear and quadratic effects of a metric reflecting the relative spread of the breeding and nonbreeding populations

(i.e. the ratio between the breeding and the nonbreeding population spread, both calculated as above). Life-history traits (body mass, habitat and dietary breadths) were compiled from the literature (Appendix S5). Finally, we included a dichotomous moderator indicating whether a species was a passerine or a non-passerine species. We initially explored the relationships between Z_r values and predictors and applied appropriate transformations whenever we detected nonlinear effects (Appendices S6 and S7). All predictors (including the binary one) were scaled, and we found no multicollinearity between predictors (Pearson's $|r| \leq 0.55$). Exploratory analysis also suggested an heterogeneity of variance in Z_r values between passerines and non-passerines (Appendix S7). Hence, we allowed for heterogeneity of variance between these groups by entering the binary predictor as inner variable in the random part of the model and setting a diagonal covariance structure for the variance–covariance matrix. This allowed the model to estimate different variances for each level of this predictor. We also scaled Z_r values by the inverse of their variance (equal to $N - 3$, where N is the number of individuals in a geographical population). Degrees of freedom were calculated with the containment method that offers a better control of the Type I error rate and produces confidence intervals with closer-to-nominal coverage rates (<https://wviechtb.github.io/metafor/reference/index.html>). Models were fitted using REML, and t-values were used as measures of effect size. Residual heterogeneity was tested through the Q_E -test (Viechtbauer, 2010).

Birds travelling farther distances have been shown to profit from a better access to annual resource fluctuations (Somveille et al., 2019). Hence, our model was also re-fitted to explicitly account for this benefit by replacing nonbreeding latitude with the annual resource surplus available to geographical populations, quantified according to Somveille et al. (2019). The normalized difference vegetation index (NDVI) was used as a general proxy of resource availability (Bonnet-Lebrun et al., 2021; Somveille et al., 2019). For both seasonal ranges, we calculated the difference between the mean NDVI when the population is present and the mean NDVI when the population is absent, and took their sum as a proxy of the annual resource surplus, a relative measure of the net gain in resource availability due to migration (Appendix S8, for details). Similarly to Somveille et al. (2019), we considered only land bird species for this analysis ($N = 145$ populations of 66 species), because NDVI is not representative of resource availability at sea.

RESULTS

Overall, migratory connectivity was moderate and significantly larger than zero (estimated $Z_r = 0.471 \pm 0.119$ SE, $t_{82} = 3.965$, $p < 0.001$, corresponding to $r_M = 0.439$, 95% CI: 0.232–0.608; Figure 2). We provided evidence of a weak phylogenetic signal in the strength of migratory

connectivity ($H^2 = 0.204$; Likelihood Ratio Test with a model not accounting for phylogeny, $\chi^2_1 = 8.077$, $df = 1$, $p = 0.004$).

When including moderator variables, we found a negative effect of migration distance and a positive effect of nonbreeding population spread on the strength of migratory connectivity, with migration distance having more than a threefold stronger effect than nonbreeding population spread (Table 2a; Figure 3). There was a significant quadratic effect of relative population spread on migratory connectivity, though this moderator had a smaller effect size than that of migration distance (Table 2a; Figure 3). Migratory connectivity increased with decreasing nonbreeding latitude (Table 2a; Figure 3), implying that populations wintering farther south showed a lower population mixing. Despite the latter had the weakest effect among geographic predictors, this result did not change when we excluded geographical populations with nonbreeding latitude $< 30^\circ$ N from the analysis (details not shown). Replacing nonbreeding latitude with annual resource surplus available to birds (Appendix S8) provided qualitatively identical results, showing that connectivity increased with better access to resources (Table 2b; Figure 3).

Among biological traits, only habitat breadth was significantly related to migratory connectivity, whereby habitat generalists showed a stronger migratory connectivity (Table 2a; Figure 3). Similarly to the model without moderators, migratory connectivity was generally positive and significantly larger than zero across all populations, suggesting that individuals tend to maintain their reciprocal positions within the geographical populations (Table 2a). We found only moderate evidence of phylogenetic heritability in the strength of migratory connectivity ($H^2 = 0.585$, $\chi^2_1 = 3.920$, $df = 1$, $p = 0.048$). When we excluded weakly migratory populations from the analysis, the effect of relative population spread and habitat breadth disappeared, while other results were unaffected (Appendix S9). Model performance was high ($R^2 = 0.80$; see also Appendices S7 and S9) and model diagnostics indicated the robustness of our analyses (Appendix S10). However, we found significant residual heterogeneity ($Q_E = 4009.408$, $df = 181$, $p < 0.001$; see also Appendix S9), possibly due to unaccounted moderators.

DISCUSSION

Geography predicts migratory connectivity by reflecting migration cost

Finch et al. (2018) showed that low migratory connectivity is widespread in Eurasian-African long-distance migrants, which tend to mix due to reduced land availability as approaching the southern portion of Africa. Our results generalized this conclusion also to short-distance migrants wintering in southern Europe or North Africa, showing that migratory connectivity was stronger in the geographical populations that spread out more extensively

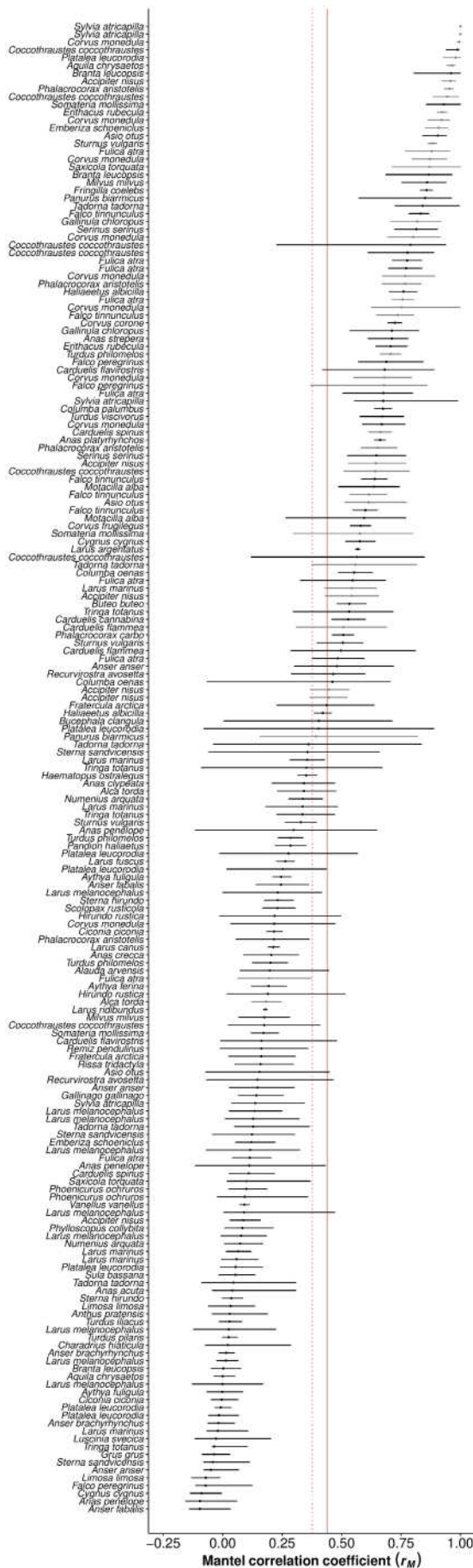


FIGURE 2 Forest plot showing the strength of migratory connectivity (r_M) for 191 geographical populations of 83 bird species, with error bars depicting bootstrapped 95% CIs of r_M values (black: strongly migratory populations; grey: weakly migratory populations). Values are ranked according to connectivity strength. Red lines show the mean connectivity strength estimated by the models without moderators fitted while considering both strongly and weakly migratory populations (solid) or strongly migratory populations only (dotted).

in the nonbreeding period, after controlling for their wintering latitude and migration distance, thus supporting the hypothesis of minimization of migration costs (Somveille et al., 2021). Indeed, since we estimated migratory connectivity at the population level, our measure of nonbreeding population spread reflects the constraints that limit individuals to spread out within clusters, thus corresponding to the total nonbreeding range spread of Finch et al. (2018). Furthermore, when we considered both weakly and strongly migratory populations, connectivity also depended on relative population spread, whereby populations occupying a similar range extent in the breeding and the nonbreeding period tended to be more strongly connected than those with ranges of different sizes.

Previous studies have considered nonbreeding latitude, nonbreeding land availability and migration distance as joint driving forces in shaping migratory connectivity because, in the European-African migration system, land availability typically decreases at southern latitudes and long-distance migrants are those spending the nonbreeding period at the southernmost sites (Finch et al., 2018). By examining populations wintering from Central Europe to Southern Africa, we disentangled the relative contribution of these factors and showed that they reflect different ways for optimizing migration costs. Contrary to our hypothesis, migratory connectivity increased in populations wintering farther south possibly because they benefit of larger annual resource surplus (Somveille et al., 2019), as confirmed by our analysis of migratory land birds that included this variable. This may suggest that migratory connectivity increases in populations wintering farther south, after controlling for migration distance and nonbreeding population spread if migration costs are traded off by better energy acquisition (Somveille et al., 2021).

Similarly, our analysis quantified the negative effect of migration distance on the strength of migratory connectivity, which was the driver with the largest effect size. Birds should minimize energetic costs of seasonal relocation following an optimal redistribution model, implying that individuals tend to migrate along the shortest available path, however, deviations from the optimal path should become relatively less important as far as migration distance increases (Figure S1, Appendix S1; Somveille et al., 2021). Consistently, our data show that the strength of migratory connectivity drops up to a migration distance of 2000–2500 km (approximately corresponding to the length of a direct crossing of the Sahara

TABLE 2 (a) Parameters estimated from the phylogenetic mixed model explaining variation in the strength of migratory connectivity (as Fisher Z-transformation of r_M value) across 191 geographical populations of 83 bird species. (b) Parameters estimated from a similar model fitted to data on land birds only where 'Nonbreeding latitude' was replaced with 'Annual resource surplus (NDVI)' (see Appendix S8, for details).

	Moderator	Coefficient	SE	<i>t</i>	df	<i>p</i>
a.	Intercept	0.558	0.132	4.229	73	<0.001*
	Migration distance (km)	-0.745	0.011	-67.170	181	<0.001*
	Nonbreeding latitude (°)	-0.031	0.010	-3.017	181	0.003*
	Nonbreeding population spread (km)	0.222	0.011	19.708	181	<0.001*
	Relative population spread	0.102	0.017	5.949	181	<0.001*
	Relative population spread ²	-0.068	0.022	-3.081	181	0.002*
	Body mass (kg)	-0.009	0.021	-0.430	73	0.669
	Habitat diversity	0.053	0.026	2.080	73	0.042*
	Diet diversity	0.013	0.023	0.557	73	0.579
	Passerine	-0.033	0.066	-0.501	73	0.618
b.	Intercept	0.621	0.157	3.942	56	<0.001*
	Migration distance (km)	-0.902	0.016	-54.647	135	<0.001*
	Annual resource surplus (NDVI)	0.066	0.011	6.209	135	<0.001*
	Nonbreeding population spread (km)	0.300	0.014	22.075	135	<0.001*
	Relative population spread	0.181	0.021	8.623	135	<0.001*
	Relative population spread ²	-0.138	0.026	-5.271	135	<0.001*
	Body mass (kg)	-0.021	0.026	-0.832	56	0.409
	Habitat diversity	0.054	0.033	1.630	56	0.109
	Diet diversity	0.009	0.033	0.259	56	0.796
	Passerine	-0.051	0.087	-0.588	56	0.559

Note: Transformations were applied to 'Migration distance' and 'Nonbreeding population spread' (Appendices S6 and S7), while the second-order polynomial term of 'Relative population spread' was included to account for quadratic effects (Table 1). All moderators are mean-centred and scaled to 1 SD. An asterisk marks significant ($p < 0.05$) moderators.

desert), then such decrease flattens. Thus, the nonlinear decrease in migratory connectivity with increasing migration distance may arise from the combined costs of avian aerial locomotion and the availability of suitable habitats for refuelling that may determine a nonlinear increase of migration costs with the distance travelled by birds (Hein et al., 2012).

Clearly, unaccounted predictors associated with geography, reflecting migration cost, can explain the residual heterogeneity shown by our model. As suggested by Somveille et al. (2021), avian redistribution patterns in seasonal ranges could be affected by *en route* environmental conditions experienced by migrants such as wind (Kranstauber et al., 2015; Norevik et al., 2020) and, especially in these cases, optimal migration routes may depart substantially from the shortest path connecting seasonal grounds. Indeed, Kranstauber et al. (2015) showed that favourable air currents influence migratory trajectories and suggested that birds could adjust migration routes at the population level by tracking efficiently the wind-optimized route. Additionally, terrestrial birds can surf the so-called 'green wave', following the spatio-temporal gradient of vegetation productivity while migrating or when stationary at their nonbreeding grounds (Kölzsch et al., 2015; Trierweiler et al., 2013). These

opportunities may have contributed to influencing the actual distance travelled by some of the investigated species and populations and, in turn, their migratory connectivity. Likewise, unaccounted predictors may explain why several weakly migratory populations (e.g. razorbill *Alca torda*, Eurasian coot *Fulica atra*, bearded reedling *Panurus biarmicus*; Figure 2) showed a lower migratory connectivity than that expected for the less mobile populations. Some degree of post-breeding dispersal may reduce connectivity, yet including a proxy of dispersal ability (Sheard et al., 2020) in our model ruled out its effect (Appendix S11). Our analytical approach was possibly unable to capture the process underlying connectivity of the three populations above, which remains to be investigated using fine-scale data.

Life-history and evolutionary drivers of migratory connectivity covary with migration propensity

Habitat specialists showed a weaker connectivity than generalists, whereas other life-history traits did not seem to play a role in shaping migratory connectivity. Bird species with a narrower habitat breadth could be

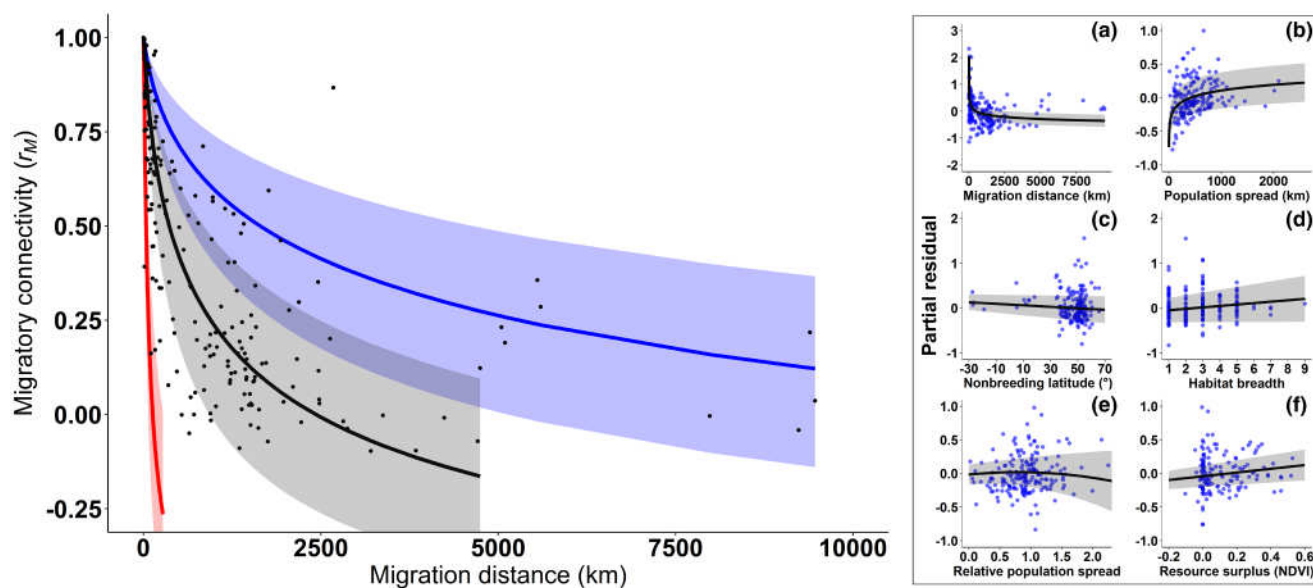


FIGURE 3 The strength of migratory connectivity (r_M) predicted in relation to the covariates with the strongest effect sizes, that is, migration distance and nonbreeding population spread. The latter is categorized into levels corresponding to the minimum (red), median (grey) and maximum (blue) observed values: 16.02 km, 423.99 km and 2100.41 km. The inset (a–f) shows the net effects of covariates significantly influencing the extent of migratory connectivity, as partial residuals. As to (f), the effect is estimated from the model replacing nonbreeding latitude with annual resource surplus (NDVI), which was fitted on land bird species (Appendix S8). For modelling, Fisher Z-transformation was applied to r_M values, and appropriate transformations were applied to migration distance and relative population spread (Appendices S6 and S7), whereas figure and inset depict back-transformed values. Lines and bands: predicted values and 95% confidence intervals. Dots: observed values.

constrained to concentrate in relatively small nonbreeding areas where individuals are more likely to mix. This result may also reflect an adaptive response of habitat specialists to the temporal shifts in the geographical position of suitable nonbreeding habitats that have occurred in the past due to the large variability in climate conditions, particularly rainfalls, that naturally occurred in sub-Saharan Africa. As suggested by Finch et al. (2018), under largely variable conditions, a weaker migratory connectivity may promote the chance of survival of a population because only a part will suffer the negative effects of an unpredicted drought in a part of the nonbreeding ranges.

The analysis conducted on both strongly and weakly migratory populations provided only moderate evidence of a phylogenetic signal in migratory connectivity. Moreover, when we considered only strongly migratory populations, both the life-history traits and the phylogenetic relatedness between species were not supported as influential drivers of migratory connectivity. The lack of phylogenetic heritability thus suggests that migratory connectivity is not a trait shared between common ancestor lineages for migratory populations, implying that the way birds redistribute during migration seems evolutionarily labile. Avian migration patterns such as the time of departure from seasonal sites, route choice and nonbreeding destinations may indeed not only be inherited genetically and under strong selection, but also be highly flexible, mirroring plastic adjustments (Åkesson & Helm, 2020; Winkler et al., 2017).

For example, they may depend on individual-specific learning capacity and social transmission within groups (Foss-Grant et al., 2018; Mueller et al., 2013; Teitelbaum et al., 2016), and change across generations with varying climatic conditions experienced *en route* or in seasonal grounds (Clausen et al., 2018; Dufour et al., 2021; Gu et al., 2021; Jiguet et al., 2019; Lameris et al., 2018; Saino & Ambrosini, 2008). Migratory connectivity is an emergent ecological property ultimately determined by plastic adaptations reflecting changes in the optimization of migration cost, possibly explaining its evolutionary lability. Indeed, episodes of migration loss and returns to resident behaviour have often occurred across avian migratory lineages, following adaptations to novel ecological opportunities (Dufour et al., 2020). In contrast, a slightly more phylogenetically-predictable pattern of migratory connectivity appeared in the analysis that included also weakly migratory populations, suggesting that closely-related species may share traits that have promoted their ‘sedentariness’, thus their stronger migratory connectivity (Winger et al., 2019).

Conservation and management implications

Previous works investigating migratory connectivity in multiple avian species generally found moderate-to-strong connectivity at the species level (Finch et al., 2018: mean $r_M \sim 0.3$, $N=28$ long-distance migrants; Somveille et al., 2021: mean $r_M \sim 0.7$, $N=25$ medium-to-long distance

migrants). Our analysis of both short- and long-distance migrants suggested that individuals tend, on average, to maintain their reciprocal positions even within geographical populations. The large intraspecific variability in avian migration strategies, through which most of our species (58%; $N=83$) geographically split into distinct migratory populations, is likely to underpin this effect. By deepening the concept of migratory connectivity through an analysis able to identify migratory clusters within species, we suggest that conservation and management strategies must consider this large variability occurring between populations. Accurate information on migratory connectivity at the population scale would improve the conservation of mobile species, not only because efforts can be directed toward distinct population-specific nonbreeding areas (Finch et al., 2018; Sarà et al., 2019; Trierweiler et al., 2014), but also because the comprehensive knowledge of the spatial connections between and within populations would allow calibrating efforts by accounting for the dependencies among seasonal ranges (Runge et al., 2014; Taylor & Norris, 2010). Similarly, estimates of migratory connectivity at the population level would be critical to informing the management of migratory species e.g. to prevent the spread of avian-borne diseases (Chen et al., 2005) or bird collisions (Van Doren & Horton, 2018), thus assisting in human health and safety. Knowledge about population-level migratory connectivity would thus allow to managing bird populations more effectively because coordinated and population-specific efforts could be targeted on both seasonal ranges, and by considering biologically relevant spatial units. For example, for bird species that segregate into distinct migratory populations during seasonal migration, improved conservation and management actions could be developed by delineating discrete, including transboundary and potentially overlapping management units (e.g., Bacon et al., 2019; Madsen et al., 2014) emerging from migratory connectivity analyses.

Caveats

There are, inevitably, limitations in any comparative analysis aiming at identifying the eco-evolutionary drivers of complex ecological properties such as migratory connectivity. Improving our model by considering biological traits and phylogeny at the population level could help to assess whether migratory connectivity of closely-related geographical populations is more similar than that expected by chance and whether it is associated with population-specific rather than species-specific characteristics. Unfortunately, avian biological traits and phylogeny at the population level are currently unavailable. Phylogeny size may especially influence the power of phylogenetic analyses (Chamberlain et al., 2012), and the detection of evolutionary processes may also show phylogenetic-scale dependence (Graham

et al., 2018). Therefore, the increasing availability of genomic data may represent a future challenge to build avian phylogenies at the population level for the existing bird species. This opportunity would potentially allow analyses at a finer phylogenetic grain, helping to draw more robust conclusions about the phylogenetic conservatism of migratory connectivity. However, we are unaware of any ecological study implementing phylogenetic trees at the population level, and the phylogenetic tree incorporated into our model clearly represents an advancement compared to previous research on migratory connectivity, which did not consider phylogeny.

Additionally, albeit commonly used in similar studies (e.g., Cohen et al., 2018; Somveille et al., 2021), ring-recovery data are known to be affected by potentially large biases in re-encounter and reporting rates (Thorup et al., 2014). Despite the robust filtering implemented to reduce heterogeneity of ringing encounters, and although our migratory connectivity estimates were not affected by uneven sampling in the nonbreeding range (Appendix S12), we cannot rule out potential biases especially if lower re-encounter probabilities occurred in Sub-Saharan Africa, i.e. for long-distant migrants. An approach chiefly reliant on tracking data may overcome the above issues (Finch et al., 2018; Sarà et al., 2019), provided that these data are available for a large number of geographical populations of different species.

Eventually, our sample size did not allow us to evaluate the interactive effects of migration distance with other predictors. Although the biological basis of our hypotheses assumed the same drivers of migratory connectivity for both long- and short-distance migrants, thus incorporating the migration distance itself as a continuous measure reflecting migration cost (Somveille et al., 2021), future studies based on a larger sample might investigate empirically whether the drivers of migratory connectivity differ throughout the *continuum* of short- and long-distance migrants.

Conclusions and future perspectives

Taking advantage of an exceptionally large dataset spanning a diversified assembly of migratory bird species, our analysis disentangled the drivers of avian migratory connectivity, suggesting that such ecological property is evolutionary labile for strictly migratory species, being conditional on highly variable, population-specific strategies of bird migration. Generally, our findings confirm that connectivity is chiefly explained by geography which, in turn, are proxies for the energetic trade-offs that individuals face when relocating between seasonal ranges, supporting that birds migrate while maximizing energy efficiency (Somveille et al., 2021). For the first time, our approach sheds light on the relative contribution of different geographic factors on migratory connectivity,

improves our knowledge of connectivity by considering both short- and long-distance migrants, and provides empirical evidence that migratory connectivity may depend on relative population spread, which is related to the concept of migratory dispersion (i.e., the extent to which species occupy larger or smaller nonbreeding ranges relative to that occupied in the breeding period; Gilroy et al., 2016). This has critical implications from a practical perspective because, together with other drivers of migratory connectivity, migratory dispersion has been related to population declines in birds. However, whilst the effect of migratory dispersion is clear, with European migrants occupying larger nonbreeding ranges relative to breeding being less likely to decline (Gilroy et al., 2016; Howard et al., 2020), the effects of migration distance and nonbreeding population spread on avian population dynamics are still obscure. Among European breeding species, migrants with a larger nonbreeding population spread appear more likely to show declining populations (Patchett et al., 2018), but confirmation is lacking (Koleček et al., 2018) and an opposite pattern has been shown in the Neotropic migration system (Patchett et al., 2018). Some studies also suggested larger population declines for long-distance European migrants than for short-distance or resident species (Howard et al., 2020; Sanderson et al., 2006; Vickery et al., 2014), but others have shown that migration distance does not influence population trends (Gilroy et al., 2016; Patchett et al., 2018) or have suggested that short migration distances reflect a lower adaptive capacity to environmental changes (La Sorte & Fink, 2017). Thus, our study clearly highlights the necessity that the potential impacts of migratory connectivity on bird population dynamics should be teased apart from those of relative population spread, migration distance and nonbreeding population spread, as each of these players is linked to the others and may have direct, indirect and potentially divergent consequences on population trends. To this end, the complex interplay between drivers of migratory connectivity unravelled here must be taken into account to determine how avian population mixing may change in space and time.

AUTHOR CONTRIBUTIONS

R.A., S.B., F.B. and F.S. conceived and supervised the research project. N.F. and R.A. conceived the paper, analysed data and wrote the first draft. A.C., A.R. and D.R. contributed critical input to the analyses. All authors contributed in writing up the manuscript.

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PEER REVIEW

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DATA AVAILABILITY STATEMENT

The dataset, the phylogenetic consensus trees and the R code used in this study are available in a public repository on Data Dryad at <https://doi.org/10.5061/dryad.gflvhmtq>. Reports including migratory connectivity data and maps for all the species and populations analysed in this study can be also downloaded from <https://migrationatlas.org/research-modules/migratory-connectivity>. The original ringing recoveries data can be requested from the BTO-EURING at <https://euring.org/data-and-codes/obtaining-data>.

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SUPPORTING INFORMATION

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Supporting Information

Eco-evolutionary drivers of avian migratory connectivity

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Appendix S1. *Detailed hypotheses and predictions on the determinants of migratory connectivity*

Our hypotheses on the drivers of migratory connectivity, especially those concerning geographic effects, stem from the optimisation of migration costs, by which individuals are expected to migrate in the most energy efficient way (Somveille et al. 2021). Migration distance is a proxy for the cost of relocating between seasonal grounds, with the shortest available path between breeding and nonbreeding sites corresponding to the maximum travel cost optimisation for birds (Somveille et al. 2021). When migration distance increases, migration cost tends to be more similar between alternative strategies representing either strong or weak connectivity (Fig. S1), in turn more mixing (lower connectivity) is likely to occur in the population (see also Gilroy et al. 2016; Somveille et al. 2021, for more details). In fact, empirical data has suggested low connectivity in long-distance migrants (Finch et al. 2018; Patchett et al. 2018). Thus, we expect that migratory connectivity would decrease with migration distance. The same principle underlies the effect of nonbreeding population spread on the strength of migratory connectivity. When breeding range spread and migration distance are held constant while the nonbreeding population spread decreases, migration cost tends to be more similar between the two opposite situations of strong and weak migratory connectivity (Fig. S1, main text), therefore population mixing (lower connectivity) is promoted. In fact, Finch et al. (2018) have found that species spreading over a larger nonbreeding range showed a stronger migratory connectivity than those spreading over a smaller nonbreeding range. Consequently, we predict that migratory connectivity would increase with increasing nonbreeding population spread. Following this prediction, geographical constraints related to the landmass configuration and availability of suitable land in the European and African continents (the occurrence of peninsulas in Southern Europe and the triangular shape of the African continent) also lead us to hypothesise that migratory populations should be constrained to mix more during the nonbreeding period at southern latitudes due to decreasing land availability (Finch et al. 2018). Therefore, we may expect a decrease in migratory connectivity with decreasing nonbreeding latitude (i.e., more southwards, because we consider latitude as ranging from negative to positive values from the equator). Moreover, the relative population spreads in seasonal grounds might also shape migratory connectivity. In fact, individuals should mix more in the nonbreeding range when nonbreeding population spread is lower than the breeding one, but should also mix more in the breeding range when the breeding population spread is lower than the nonbreeding

one. Indeed, since the measure used to quantify connectivity (r_M) is symmetric, i.e. does not consider breeding and nonbreeding ranges differently, it declines whenever mixing of individuals occurs in either range. This consideration allows predicting that migratory connectivity should peak when the breeding and nonbreeding ranges have the same extent, i.e. when the relative population spread is one, and decrease for lower or larger values. Thus, we expect a quadratic effect of relative population spread on migratory connectivity. The relative population spread is similar to the concept of migratory dispersion, which was defined as the extent to which species occupy larger or smaller nonbreeding ranges relative to that occupied in the breeding period (Gilroy et al. 2016). However, this measure is strongly affected by the absolute size of the breeding and nonbreeding ranges, while the ratio between the range sizes is not, and this allows a better comparison among geographical populations that can largely differ in population size.

Amongst species-specific traits, low niche specialization is common in long-distance migrants, as it can be advantageous due to the variety of habitats and food resources met *en route* and on the nonbreeding grounds (Reif et al. 2016). If so, natural selection could have maintained a stronger migratory connectivity in specialized species in terms of habitat and dietary breadths, whereas a weaker migratory connectivity should be associated to generalists (Cresswell 2014). Body mass is a fundamental predictor of life-history traits and might affect migratory connectivity in several ways. Larger birds generally show longer migration distances due to an optimization of the aerial locomotion costs (Hein et al. 2012), which may exert an indirect effect on migratory connectivity by increasing population mixing. However, larger species also live longer and, in avian migrants, a longer lifespan promotes the social transmission of migratory routes towards nonbreeding sites and helps maintaining migratory knowledge across generations (Teitelbaum et al. 2016; Foss-Grant et al. 2018). Once the effect of migration distance is accounted for, a larger body mass could thus be expected to favour the evolution of a stronger migratory connectivity.

Finally, we tested whether migratory connectivity differed between passerine and non-passerine birds. Our expectation is based on the serial residency hypothesis (Cresswell 2014), which predicts that birds should tend to redistribute stochastically over a wider nonbreeding area if they are unable to forecast and compensate for any favourable or unfavourable events during migration, particularly juveniles at first migration, with consequences on the strength of migratory connectivity. Passerine and non-passerine birds typically show contrasting behaviour upon the first migration. Among passerines, juveniles usually migrate separately from

adults, whilst non-passerine juveniles tend to follow adults upon migration. The serial residency hypothesis indeed predicts that passerines would retain a generally lower migratory connectivity than non-passerines due to larger unpredictability of conditions during migration and in the nonbreeding grounds (Cresswell 2014).

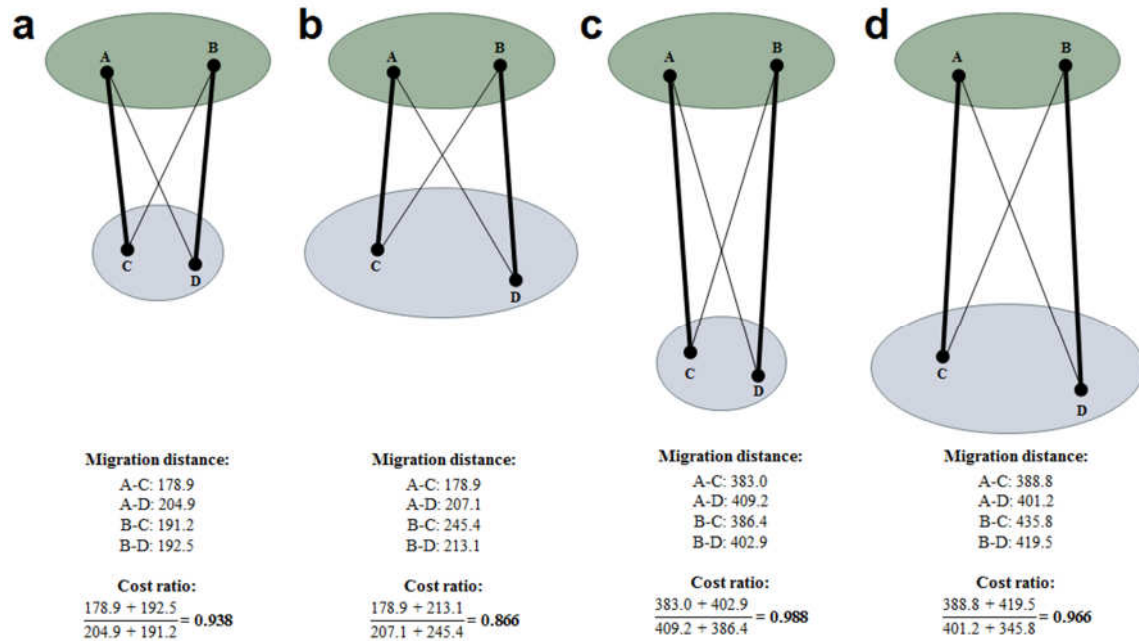


Figure S1. Schematic diagram showing how migration cost potentially underpins the effect of migration distance and nonbreeding population spread on the strength of migratory connectivity, as predicted by theory and assuming that migration cost is linearly related to migration distance (Somveille et al. 2021). Dark and light grey ellipses depict breeding and nonbreeding ranges, respectively. Dots represent two breeding and nonbreeding locations per seasonal range (each is indicated by a different letter). Population spread within a seasonal range can be considered as the linear distance between the two locations in same range (not shown). Migration distance is the line connecting a breeding to a nonbreeding site (either thick line in case of strong migratory connectivity or thin line for weak migratory connectivity). Four scenarios, with populations having the same breeding population spread, are shown: (a) short distance and small nonbreeding spread; (b) short distance and large nonbreeding spread; (c) long distance and small nonbreeding spread; (d) long distance and large nonbreeding spread. For each scenario, migration distances between two breeding and two wintering locations are calculated. Migration distances are summed for each situation (strong migratory connectivity vs weak migratory connectivity), and the ratio (cost ratio) between the two sums is derived. Both when migration distance increases (a vs c, and b vs d) and when nonbreeding population spread decreases (b vs a, and d vs c), the cost ratio increases, meaning that the cost of migration tends to be more similar between the two situations (strong vs weak migratory connectivity), and therefore population mixing (lower migratory connectivity) is more likely to occur.

Appendix S2. *Selection of ringing encounters*

Ringing data are largely heterogeneous as to individual encounter conditions and circumstances, such as those of birds in poor health status or kept in captivity or manipulated for long, or those for which the place/time of recovery was not determined accurately, therefore preliminary data filtering is necessary to reduce heterogeneity (e.g. Paradis et al. 1998). Moreover, ringing encounters collected outside species-specific breeding or nonbreeding grounds and periods, potentially resulting from occasional movements, would introduce spatial and temporal biases in migratory connectivity analysis (e.g. Somveille et al. 2021). Hence, we implemented a robust data selection procedure to discard ringing encounters potentially affecting our estimates of migratory connectivity. Our conservative approach relied on 21 condition-based selection criteria and applied a spatiotemporal masking for each species, discarding on average 97.2% of ringing encounters per species (range: 58% – 99%; see Ambrosini et al. 2016 for a similar data selection procedure).

First, we selected ringing encounter previously validated for their usage in the European-African Bird Migration Atlas (EURING level for ‘use.for.atlas’ = TRUE). Then, following previous studies on ringing encounters obtained from the EURING Data Bank (e.g. Paradis et al. 1998; Ambrosini et al. 2016), we implemented condition-based criteria in order to reduce encounter heterogeneity. In details, we removed:

- 1) birds that were not found freshly dead, or birds that were in poor condition or had an accident when ringed, or birds that were alive and probably healthy but taken into captivity (EURING levels for ‘condition’ = 3, 4, 5, 6);
- 2) birds that were kept for more than 13 hours during ringing, or birds that have been moved or held extensively during ringing, or those hand reared (EURING levels for ‘manipulated’ = C, F, T, M, H);
- 3) birds that were moved unintentionally by man or other agency, or intentionally by man, or moved by water e.g. found on shoreline (EURING levels for ‘moved’ = 2, 4, 6);
- 4) birds for which the dates of ringing and/or recovery were not recorded accurately to the nearest 1 week for both the ringing and the finding date (EURING levels for ‘date accuracy’ = 4, 5, 6, 7, 8);
- 5) birds for which the places of ringing and/or recovery were not recorded accurately to the nearest 100 km for the ringing or finding places (EURING levels for ‘coordinates accuracy’ = 6, 7, 8, 9).

Then, we applied a spatiotemporal masking using species-specific criteria in order to remove encounters in non-stationary periods or found within occasional ranges, thus retaining all individuals at their breeding and nonbreeding grounds. Following the phenology reported by Cramp (1998), we identified for each species an “extended” breeding period, corresponding to that spanning the breeding of the whole species (even though it may still include *en route* birds for some populations), a “focal” breeding stationary period from the end of spring migration for the latest population to the onset of autumn migration of the earliest population, and a nonbreeding stationary period from the end of autumn migration of the latest population until the onset of spring migration of the earliest population (Table S1). We then removed:

6) encounters found out of the focal breeding and the nonbreeding stationary periods. However, we retained encounters of chicks unable to fledge or of individuals found at nest (EURING levels for ‘catching method’ = N and for ‘age by scheme’ = 1) if they occurred during the extended breeding period of each species because we considered that they have occurred in the breeding area of the individual.

Moreover we identified breeding, resident and nonbreeding stationary ranges for each species according to the distribution maps provided by BirdLife International (2019), removing:

7) encounters outside the breeding and resident ranges during the focal or the extended breeding periods;
8) encounters outside the nonbreeding and resident ranges during the nonbreeding stationary period.

After the above step, we checked and manually discarded a few encounters of long-distance migrants remaining outside the European-African migration system, even though within their nonbreeding stationary ranges, to avoid spatial biases in migratory connectivity analysis (e.g. encounters of *Larus* spp. in North America; *Platalea leucorodia* in India; *Sterna hirundo* in Australia and South America). Eventually, we also removed:

9) repeated encounters for the same individual in either the breeding or the nonbreeding range, if any, by retaining the earliest encounter in order to minimise age-bias.

10) individuals that, after the previous steps of data selection, did not have one observation in both the breeding and the nonbreeding ranges.

Table S1. Species-specific phenology reported by Cramp (1998): the first integer represents the month, the second one represents the monthly quarter. An exception was made for *Turdus merula*, for which we considered different periods than those reported by Cramp (1998), which were referred mainly to British populations, particularly for the start of the nonbreeding period (Santos 1982; Olioso 1995; Main 2002; Andreotti et al. 2010).

Species	Extended breeding period		Focal breeding period		Nonbreeding period	
	Start	End	Start	End	Start	End
<i>Anas crecca</i>	3-4	8-2	6-1	7-3	12-3	2-2
<i>Erithacus rubecula</i>	3-1	7-3	5-2	6-4	11-3	2-3
<i>Hirundo rustica</i>	4-4	10-2	6-1	6-4	12-3	1-4
<i>Phalacrocorax carbo</i>	2-4	10-2	4-3	6-2	12-1	1-2
<i>Cygnus olor</i>	4-2	11-2	4-4	8-4	1-1	2-4
<i>Larus argentatus</i>	4-2	8-4	5-3	8-2	12-2	2-2
<i>Larus ridibundus</i>	4-1	9-1	5-2	6-2	12-1	2-2
<i>Turdus merula</i>	2-4	9-2	5-2	7-2	12-1	2-2
<i>Parus caeruleus</i>	4-1	7-3	5-1	6-1	12-1	1-4
<i>Parus major</i>	3-2	7-4	5-1	6-1	12-1	1-4
<i>Anas platyrhynchos</i>	2-1	11-4	5-3	7-4	12-1	1-4
<i>Anser albifrons</i>	6-2	9-1	6-3	8-3	1-1	2-3
<i>Anser anser</i>	3-4	8-1	4-3	7-3	12-3	1-3
<i>Accipiter gentilis</i>	4-1	8-2	6-1	7-3	12-3	2-2
<i>Calidris alpina</i>	5-4	8-4	6-2	7-1	11-3	3-1
<i>Carduelis chloris</i>	4-4	7-2	5-3	7-1	12-1	2-4
<i>Carduelis spinus</i>	4-3	8-3	5-3	7-4	12-3	2-2
<i>Haematopus ostralegus</i>	3-4	9-3	5-1	6-2	11-1	1-4
<i>Larus canus</i>	5-3	9-1	6-1	7-2	12-2	2-2
<i>Larus fuscus</i>	4-4	9-1	6-1	6-4	12-2	1-4
<i>Larus marinus</i>	4-3	8-4	5-2	8-2	12-2	2-2
<i>Ciconia ciconia</i>	4-2	8-2	5-3	7-4	12-1	1-4
<i>Cygnus cygnus</i>	5-3	9-4	6-3	9-2	1-1	2-3
<i>Falco tinnunculus</i>	3-4	8-2	6-1	7-3	12-3	2-2
<i>Haliaeetus albicilla</i>	4-2	9-2	6-1	8-4	12-3	2-1
<i>Larus melanocephalus</i>	5-2	8-3	6-1	6-4	12-1	2-4
<i>Sturnus vulgaris</i>	4-2	6-2	5-1	6-1	12-1	1-4
<i>Turdus philomelos</i>	2-4	9-1	5-3	8-2	11-2	2-3
<i>Phalacrocorax aristotelis</i>	2-3	10-1	4-3	6-4	12-1	1-1
<i>Platalea leucorodia</i>	4-1	9-2	6-1	7-3	11-4	1-4
<i>Riparia riparia</i>	4-4	8-4	6-1	6-4	12-1	2-3
<i>Sylvia atricapilla</i>	4-3	8-1	6-3	7-2	12-1	1-3
<i>Tyto alba</i>	2-4	12-1	4-1	10-4	12-2	2-3
<i>Alauda arvensis</i>	3-3	8-4	5-1	7-4	12-1	1-3
<i>Anas acuta</i>	4-1	8-3	6-1	7-4	12-3	1-4

<i>Anser brachyrhynchus</i>	5-2	8-4	5-3	8-2	12-1	3-2
<i>Anser fabalis</i>	5-3	9-1	6-1	8-3	1-1	2-4
<i>Aythya ferina</i>	4-3	8-2	5-3	7-4	12-2	1-4
<i>Aythya fuligula</i>	5-2	9-3	5-3	8-4	1-1	2-3
<i>Aythya marila</i>	5-3	9-3	5-4	8-2	12-1	2-3
<i>Branta canadensis</i>	3-3	6-4	4-1	7-2	12-1	1-4
<i>Bucephala clangula</i>	5-2	9-2	5-3	8-3	1-1	2-2
<i>Columba oenas</i>	4-3	10-4	5-1	8-2	12-1	1-4
<i>Columba palumbus</i>	2-3	12-1	5-2	8-3	12-2	2-2
<i>Corvus corone</i>	4-1	6-2	5-1	5-4	11-4	2-2
<i>Corvus frugilegus</i>	3-1	5-3	4-2	5-1	12-1	2-2
<i>Corvus monedula</i>	4-2	6-3	4-4	6-1	11-3	2-2
<i>Fulica atra</i>	2-4	10-1	5-3	8-2	12-2	2-3
<i>Gallinago gallinago</i>	3-4	9-3	5-1	7-2	12-1	1-4
<i>Gallinula chloropus</i>	4-3	8-3	5-3	7-2	12-3	2-3
<i>Garrulus glandarius</i>	3-4	7-4	6-2	6-4	11-2	3-3
<i>Limosa limosa</i>	4-1	8-3	5-2	6-3	11-1	2-1
<i>Anas penelope</i>	5-3	8-4	6-1	8-2	12-3	2-4
<i>Anas strepera</i>	4-3	8-3	5-3	7-4	12-3	2-4
<i>Mergus merganser</i>	5-2	9-3	6-2	9-1	1-1	2-4
<i>Netta rufina</i>	4-4	8-3	5-3	8-1	12-2	2-2
<i>Numenius arquata</i>	4-4	9-1	5-2	6-4	11-3	2-3
<i>Pluvialis apricaria</i>	5-2	9-3	6-3	7-4	1-1	2-2
<i>Scolopax rusticola</i>	3-2	10-3	5-2	8-3	12-1	2-3
<i>Somateria mollissima</i>	4-3	9-3	5-2	8-4	12-2	2-3
<i>Anas clypeata</i>	4-4	8-3	5-3	8-1	12-2	2-2
<i>Streptopelia decaocto</i>	5-1	10-1	6-1	7-3	12-1	2-1
<i>Tetrao urogallus</i>	4-3	9-2	5-3	8-2	9-3	4-2
<i>Tringa totanus</i>	3-4	8-3	5-1	6-2	11-1	2-2
<i>Turdus iliacus</i>	4-4	8-1	6-2	7-2	12-2	2-3
<i>Turdus pilaris</i>	4-4	8-4	5-2	8-1	12-2	2-3
<i>Turdus viscivorus</i>	3-3	7-3	4-4	6-4	11-2	2-2
<i>Vanellus vanellus</i>	3-1	9-2	4-2	6-2	12-2	1-3
<i>Carduelis flammea</i>	4-3	7-4	5-3	7-2	12-3	2-3
<i>Accipiter nisus</i>	4-4	8-2	6-1	7-4	12-1	2-3
<i>Acrocephalus melanopogon</i>	4-4	8-2	6-1	7-4	12-2	2-2
<i>Aegithalos caudatus</i>	4-2	9-1	5-2	8-1	11-2	3-3
<i>Aegolius funereus</i>	2-4	8-3	5-1	7-4	12-1	2-4
<i>Aix galericulata</i>	4-3	8-1	4-4	7-3	8-2	4-2
<i>Alca torda</i>	4-3	8-2	5-2	7-2	11-1	2-2
<i>Alcedo atthis</i>	4-3	10-1	6-1	7-2	12-2	2-4
<i>Anthus pratensis</i>	3-4	8-4	5-3	6-2	12-1	1-4
<i>Aquila chrysaetos</i>	3-1	8-2	4-2	7-4	12-2	2-4
<i>Ardea cinerea</i>	3-3	8-2	4-1	6-2	12-1	1-2
<i>Asio otus</i>	2-4	7-4	6-3	7-2	12-3	2-3
<i>Athene noctua</i>	3-3	8-3	4-2	7-4	8-4	3-2
<i>Branta leucopsis</i>	5-4	8-4	6-1	8-2	1-1	3-3

<i>Buteo buteo</i>	3-4	8-1	5-3	7-3	12-1	1-4
<i>Carduelis carduelis</i>	5-1	8-4	5-4	8-1	12-2	2-4
<i>Certhia brachydactyla</i>	3-4	7-4	4-3	7-2	8-1	3-3
<i>Certhia familiaris</i>	3-4	7-4	5-1	6-4	12-1	2-4
<i>Cettia cetti</i>	6-1	8-4	6-3	8-2	11-1	4-4
<i>Charadrius alexandrinus</i>	4-2	9-1	5-4	6-4	11-3	2-2
<i>Charadrius hiaticula</i>	3-4	9-3	5-3	7-4	12-1	1-4
<i>Cinclus cinclus</i>	2-4	9-1	6-1	8-2	11-1	1-4
<i>Coccothraustes coccothraustes</i>	4-2	8-4	4-3	8-3	12-2	1-4
<i>Corvus corax</i>	1-4	8-4	4-1	6-3	9-1	1-3
<i>Luscinia svecica</i>	4-4	8-2	6-2	7-4	12-1	1-4
<i>Dendrocopos major</i>	4-3	7-3	5-1	7-1	11-2	2-4
<i>Emberiza citrinella</i>	4-3	8-2	5-3	7-4	11-3	3-2
<i>Emberiza schoeniclus</i>	5-1	7-3	5-3	7-1	11-3	2-2
<i>Falco peregrinus</i>	5-1	8-3	5-3	7-4	11-3	2-3
<i>Fratercula arctica</i>	4-1	8-4	4-3	7-4	10-2	3-2
<i>Fringilla coelebs</i>	4-3	7-2	5-4	6-3	12-1	2-3
<i>Fringilla montifringilla</i>	5-2	7-3	6-2	7-1	11-3	2-3
<i>Glaucidium passerinum</i>	4-3	7-4	4-4	7-2	12-2	2-4
<i>Grus grus</i>	4-4	9-2	5-3	7-2	12-1	2-2
<i>Sterna caspia</i>	5-1	8-3	5-4	7-2	12-3	2-3
<i>Larus audouinii</i>	4-3	8-3	5-1	7-4	11-3	2-2
<i>Carduelis cannabina</i>	4-3	8-4	5-3	7-4	11-3	2-2
<i>Carduelis flavirostris</i>	4-1	8-3	5-3	7-4	12-3	2-3
<i>Milvus milvus</i>	3-4	7-4	5-3	7-2	12-1	1-4
<i>Sula bassana</i>	4-2	11-1	6-2	7-4	12-1	1-4
<i>Motacilla alba</i>	4-1	8-2	6-1	7-3	11-3	1-4
<i>Nucifraga caryocatactes</i>	2-4	7-4	4-1	5-4	11-1	2-3
<i>Pandion haliaetus</i>	4-3	8-1	6-1	7-4	12-1	2-3
<i>Panurus biarmicus</i>	3-3	7-1	4-3	6-3	12-1	2-3
<i>Passer hispaniolensis</i>	3-1	10-4	5-3	8-3	12-1	2-4
<i>Passer montanus</i>	4-1	9-2	4-4	8-2	11-3	3-3
<i>Parus ater</i>	4-2	7-4	4-4	6-3	10-3	2-4
<i>Phoenicurus ochruros</i>	4-3	7-4	5-1	7-1	11-3	2-3
<i>Phylloscopus collybita</i>	4-4	8-1	6-1	7-2	11-4	2-3
<i>Podiceps cristatus</i>	2-3	9-1	5-3	7-3	12-3	1-4
<i>Parus montanus</i>	4-3	8-1	5-3	6-1	12-1	3-1
<i>Prunella modularis</i>	3-1	9-2	6-1	8-2	11-3	2-3
<i>Pyrrhula pyrrhula</i>	4-3	8-3	5-2	7-3	12-1	2-3
<i>Recurvirostra avosetta</i>	4-2	9-1	5-3	7-2	11-2	2-2
<i>Regulus ignicapillus</i>	4-2	8-2	5-3	7-4	12-3	1-4
<i>Regulus regulus</i>	4-3	8-3	6-1	7-3	12-1	2-3
<i>Remiz pendulinus</i>	4-4	8-3	5-1	6-4	12-3	1-4
<i>Rissa tridactyla</i>	5-2	9-2	6-1	7-4	12-1	1-4
<i>Saxicola torquata</i>	4-2	8-2	6-1	7-4	11-4	1-3
<i>Serinus serinus</i>	4-1	8-2	5-1	7-3	11-2	2-3
<i>Sitta europaea</i>	3-4	7-3	5-1	6-2	11-1	3-3

<i>Sterna hirundo</i>	5-2	9-2	6-3	7-2	11-2	3-1
<i>Sterna albifrons</i>	5-2	9-1	6-1	7-2	11-2	3-1
<i>Strix aluco</i>	2-3	7-1	2-4	6-3	7-4	2-2
<i>Sylvia melanocephala</i>	3-3	7-2	5-1	5-4	1-1	2-2
<i>Tadorna tadorna</i>	4-3	9-1	5-1	8-3	1-1	2-2
<i>Sterna sandvicensis</i>	4-4	7-4	5-2	6-4	12-1	1-2
<i>Troglodytes troglodytes</i>	3-4	8-3	5-4	7-4	12-3	3-2
<i>Uria aalge</i>	4-4	8-1	5-1	7-2	12-1	1-4

Appendix S3. *Discrimination of strongly and weakly migratory geographical populations*

The distinction between migratory and non-migratory species or populations is particularly challenging for birds, where different types of migration are exhibited and large differences exist, even within populations of the same species (Chapman et al. 2011; Gilroy et al. 2016). In this work, we avoided to classify migratory vs resident geographical populations (i.e. clusters identified by the migratory connectivity analysis) on the basis of a cut-off migration distance, because many populations of birds are known to migrate even when migration distance is short, and migration distance is often population-specific as well as is influenced by various ecological and climatic drivers (Visser et al. 2009; Meller et al. 2016; Curley et al. 2020). Indeed, migration distance is only one amongst the many other characteristics that are used to define migration, such as periodicity or directionality of movements (Eyres et al. 2017). Consequently, we implemented a classification based on the inspection of the overall spatial pattern of individual positions observed in the nonbreeding stationary range, (i.e. after migration), relative to that found in the breeding range, i.e. prior to migration. For each geographical population, we first calculated the 95% minimum convex polygon (MCP) of individual locations in the breeding period (breeding MCP) and the 95% MCP of the same individual locations in the nonbreeding stationary period (nonbreeding MCP). We then overlapped the breeding and nonbreeding MCPs and classified those geographical populations for which the overlap was more than 75% of the area of the breeding MCP as weakly migratory. Populations were classified as strongly migratory otherwise. The distribution of the overlap values obtained from our data was clearly bimodal (Fig. S2), indicating that a dichotomous classification into strongly and weakly migratory populations is helpful to distinguish the ‘more’ from the ‘less mobile’ geographical populations. The presence of overlap between breeding and nonbreeding ranges has been already used previously as a criterion to discriminate, at the species level, full migrants, partial migrants and residents (Gilroy et al. 2016). Moreover, MCP has been used to quantify population spread of birds in the breeding and nonbreeding grounds (e.g. Blackburn et al. 2017; Burgess et al. 2020) and we should note that, in our case, the increase in MCP extent with increasing sample size (Burgess et al. 2020) would not affect the classification outcome because both the MCPs include exactly the same number of individuals.

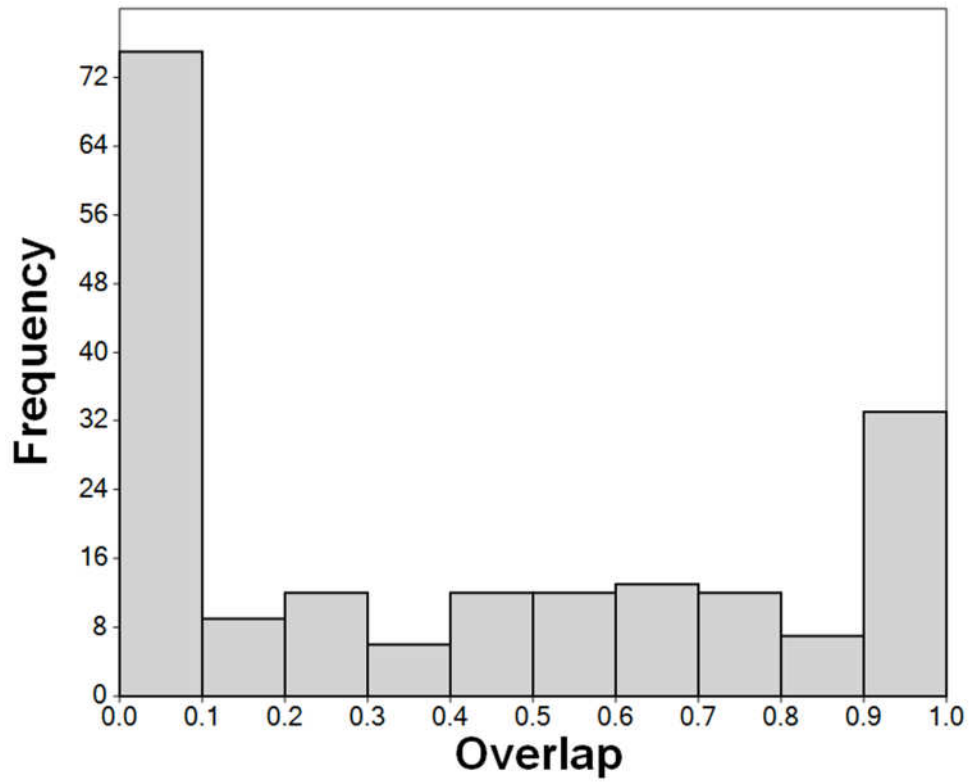


Figure S2. Frequency distribution of values reflecting the overlap between nonbreeding and breeding MCPs in our study populations.

Appendix S4. *Species investigated in phylogenetic comparative analysis*

Table S2. Number of geographical populations (i.e., clusters) and individuals for each bird species included in the phylogenetic comparative analysis of the strength of migratory connectivity.

Family	Species	N geographical populations	N individuals
Accipitridae	<i>Accipiter nisus</i>	6	3481
	<i>Aquila chrysaetos</i>	2	437
	<i>Buteo buteo</i>	1	3509
	<i>Haliaeetus albicilla</i>	2	2629
	<i>Milvus milvus</i>	2	862
Alaudidae	<i>Alauda arvensis</i>	1	53
Alcidae	<i>Alca torda</i>	2	521
	<i>Fratercula arctica</i>	2	193
Anatidae	<i>Anas acuta</i>	1	34
	<i>Anas clypeata</i>	1	278
	<i>Anas crecca</i>	1	305
	<i>Anas penelope</i>	3	147
	<i>Anas platyrhynchos</i>	1	6333
	<i>Anas strepera</i>	1	234
	<i>Anser anser</i>	3	157
	<i>Anser brachyrhynchus</i>	2	1089
	<i>Anser fabalis</i>	2	289
	<i>Aythya ferina</i>	1	552
	<i>Aythya fuligula</i>	2	1095
	<i>Branta leucopsis</i>	3	402
	<i>Bucephala clangula</i>	1	69
	<i>Cygnus cygnus</i>	2	529
	<i>Somateria mollissima</i>	3	1182
<i>Tadorna tadorna</i>	5	310	
Charadriidae	<i>Charadrius hiaticula</i>	1	42
	<i>Vanellus vanellus</i>	1	3646
Ciconiidae	<i>Ciconia ciconia</i>	2	3783
Columbidae	<i>Columba oenas</i>	2	345
	<i>Columba palumbus</i>	1	1522
Corvidae	<i>Corvus corone</i>	1	1264
	<i>Corvus frugilegus</i>	1	517
	<i>Corvus monedula</i>	9	1557
Emberizidae	<i>Emberiza schoeniclus</i>	2	937
Falconidae	<i>Falco peregrinus</i>	3	201

	<i>Falco tinnunculus</i>	5	7662
Fringillidae	<i>Carduelis cannabina</i>	1	491
	<i>Carduelis flammea</i>	2	202
	<i>Carduelis flavirostris</i>	2	66
	<i>Carduelis spinus</i>	2	1370
	<i>Coccothraustes coccothraustes</i>	7	571
	<i>Fringilla coelebs</i>	1	2682
	<i>Serinus serinus</i>	2	672
Gruidae	<i>Grus grus</i>	1	372
Haematopodidae	<i>Haematopus ostralegus</i>	1	2349
Hirundinidae	<i>Hirundo rustica</i>	2	96
Laridae	<i>Larus argentatus</i>	1	19,823
	<i>Larus canus</i>	1	3878
	<i>Larus fuscus</i>	1	5015
	<i>Larus marinus</i>	6	2380
	<i>Larus melanocephalus</i>	9	1099
	<i>Larus ridibundus</i>	1	27,479
	<i>Rissa tridactyla</i>	1	134
Motacillidae	<i>Anthus pratensis</i>	1	51
	<i>Motacilla alba</i>	2	439
Muscicapidae	<i>Erithacus rubecula</i>	2	6621
	<i>Luscinia svecica</i>	1	35
	<i>Phoenicurus ochruros</i>	2	306
	<i>Saxicola torquata</i>	2	65
Pandionidae	<i>Pandion haliaetus</i>	1	369
Paridae	<i>Panurus biarmicus</i>	2	145
Phalacrocoracidae	<i>Phalacrocorax aristotelis</i>	4	916
	<i>Phalacrocorax carbo</i>	1	2033
Phylloscopidae	<i>Phylloscopus collybita</i>	1	133
Rallidae	<i>Fulica atra</i>	9	3806
	<i>Gallinula chloropus</i>	2	528
Recurvirostridae	<i>Recurvirostra avosetta</i>	2	372
Remizidae	<i>Remiz pendulinus</i>	1	53
Scolopacidae	<i>Gallinago gallinago</i>	1	137
	<i>Limosa limosa</i>	2	165
	<i>Numenius arquata</i>	2	485
	<i>Scolopax rusticola</i>	1	518
	<i>Tringa totanus</i>	4	363
Sternidae	<i>Sterna hirundo</i>	2	1089
	<i>Sterna sandvicensis</i>	3	423

Strigidae	<i>Asio otus</i>	3	618
Sturnidae	<i>Sturnus vulgaris</i>	3	7038
Sulidae	<i>Sula bassana</i>	1	350
Sylviidae	<i>Sylvia atricapilla</i>	4	255
Threskiornithidae	<i>Platalea leucorodia</i>	7	817
Turdidae	<i>Turdus iliacus</i>	1	683
	<i>Turdus philomelos</i>	3	5137
	<i>Turdus pilaris</i>	1	1572
	<i>Turdus viscivorus</i>	1	542

Appendix S5. *Compilation of species-specific life-history traits*

Species-specific life-history traits entered as fixed effects in the meta-analysis were: diet breadth, habitat breadth, and body mass. Information on diet breadth was obtained from del Hoyo et al. (2017). We considered the following food sources: 1) terrestrial invertebrates, 2) aquatic invertebrates, 3) terrestrial vertebrates, 4) aquatic vertebrates, 5) fruits and berries, 6) sedges and seeds, 7) other plant materials (e.g. nectar, grass or aquatic plants) and counted the number of them representing a substantial contribution to a species diet to obtain an estimate of the diet breadth. Similarly, habitat breadth was estimated as the total number of primary habitat types exploited by a species, as reported in BirdLife (2020). Body mass was also compiled from BirdLife (2020).

Appendix S6. *Moderators entered in the phylogenetic model*

Table S3. Information on moderators initially considered as fixed effects for the phylogenetic mixed model concerning 191 geographical populations of 83 bird species. For relevant hypotheses, see Table 1 in the main text and Appendix S1.

Moderator (units)	Type	Transformation	Level	Notes
Mean migration distance (km)	Continuous	$\ln(\ln(x) + 1)$	Geographical population	Not collinear with nonbreeding latitude ($r = -0.31$) or nonbreeding population spread ($r = 0.54$)
Mean nonbreeding latitude ($^{\circ}$)	Continuous	–	Geographical population	Not collinear with migration distance ($r = -0.31$) or nonbreeding population spread ($r = -0.35$)
Mean nonbreeding population spread (km)	Continuous	$\ln(x)$	Geographical population	Not collinear with migration distance ($r = 0.54$) or nonbreeding latitude ($r = -0.35$)
Relative population spread (both linear and squared term)	Continuous	–	Geographical population	Not collinear with nonbreeding population spread ($r = -0.21$)
Body mass (grams)	Continuous	–	Species	–
Habitat breadth	Integer	–	Species	Not collinear with diet breadth ($r = 0.14$)
Diet breadth	Integer	–	Species	Not collinear with habitat breadth ($r = 0.14$)
Passerine or non-passerine	Binary	–	Species	–

Appendix S7. Preliminary data exploration and model performance

Visual inspection of the data showed that Z_r values varied non-linearly with mean migration distance. In contrast, the relationship was linear after double-natural logarithm transformation (Fig. S3). The variance of Z_r values also markedly differed between passerine and non-passerines (Fig. S4). Z_r values predicted from the full model showed a good agreement with observed ones indicating a proper fit of the model (Fig. S5).

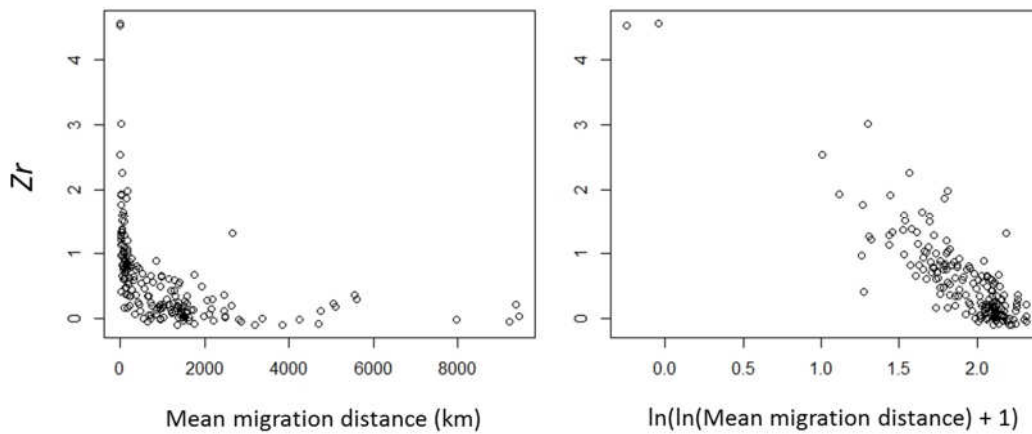


Figure S3. Z_r values according to mean migration distance before and after the double logarithmic transformation. Note that a constant term (1) was added to avoid negative values in the second logarithm.

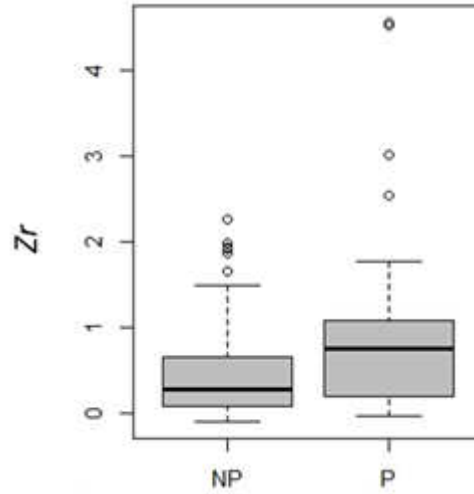


Figure S4. Z_r values of passerine (P) and non-passerine (NP) birds showing variance heterogeneity between groups.

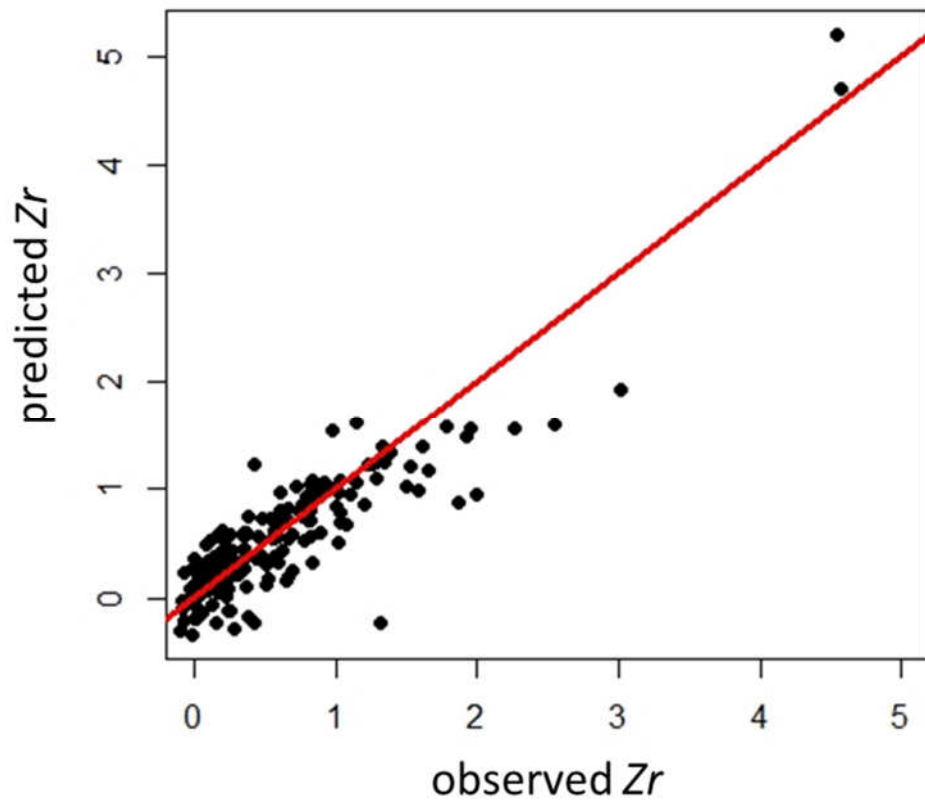


Figure S5. Model performance evaluated through inspection of predicted vs. observed Z_r values ($R^2 = 0.80$). Diagonal line is the 1 to 1 line.

Appendix S8. Modelling the annual surplus in resource availability due to migration strategy

Somveille et al. (2019) have shown that birds travelling farther distances exploit annually better access to resources due to their migration strategy. Hence, stronger connectivity may occur for populations wintering farther south, after controlling for the other geographic predictors in our model, because they can exploit larger amount of resources. To test for this hypothesis, we re-run our analysis to explicitly account for the benefit in resource availability gained annually by birds.

We followed the approach described by Somveille et al. (2019) to assess the annual surplus in resources available to birds. However, we note that these authors first calculated this measure and then took the opposite for their analyses, thus ending up calculating an index of resource scarcity, while we use this measure directly, as we aimed at considering resource availability. We used the normalized difference vegetation index (NDVI), as a general proxy of the resources available to bird species (e.g. food, roosting sites; Somveille et al. 2015; 2019). Similarly to Somveille et al. (2019), we considered only land bird species for this analysis (N = 145 populations of 66 species), because NDVI values would not be representative of resource availability at sea, where marine species spend the nonbreeding period. We obtained NDVI monthly averages, for the period 2000-2019, at a resolution of 0.05° (MOD13C2 Version 6 product; Didan 2015). Only land areas were considered for the calculation of NDVI, through a spatial masking discarding marine areas and water bodies. Previous studies assumed that the resources available to terrestrial birds are related to the surplus in NDVI, i.e. the difference between NDVI in the season when the migrant birds are present and the season when they are absent (Herrera 1978, Hurlbert and Haskell 2003, Dalby et al. 2014, Somveille et al. 2015; 2019). For each geographical population, we calculated the mean NDVI experienced in both the breeding and the nonbreeding ranges when the population is present (i.e., between May-August for breeding and between November-February for nonbreeding periods; Somveille et al. 2015; 2019) and the mean NDVI over the same ranges but when the population is absent (i.e., between November-February for breeding and between May-August for nonbreeding periods; Somveille et al. 2015; 2019). Specifically, we took the average of NDVI values across all pixels over a 'sample population area' corresponding to the range of 75% individuals' locations in each seasonal range (to avoid considering NDVI exploited by individuals located at the borders of each geographical population; mean \pm SE, breeding range: 7837 \pm 1046 pixels/population; nonbreeding range: 6701 \pm 862

pixels/population). For each population and for both seasonal grounds, we then calculated the difference between the mean NDVI value when the species is present and the mean NDVI value when the species is absent, i.e. the resource surpluses each population gained in both seasons (positive values indicate that a resource surplus is available to the population, whereas negative values imply a deficit; Somveille et al. 2019). Finally, we summed the two seasonal surpluses to obtain a measure of the annual surplus in resources available to birds (Somveille et al. 2019). Most of our study populations exploited a resource deficit in the nonbreeding grounds (85% populations) and a resource gain in the breeding grounds (97% populations), leading to an overall annual gain for most (83%) populations, strongly supporting previous findings (cf. Bonnet-Lebrun et al. 2021).

In our models fitted on both weakly and strongly migratory populations and strongly migratory populations only, we replaced the nonbreeding latitude with the annual resource surplus. While other main results were unaffected, the resource surplus gained by each population annually showed a positive effect on the strength of migratory connectivity (Table 2b, main text, and Table S4b, Appendix S9), likely explaining why populations wintering farther south showed stronger migratory connectivity (see Discussion, main text). Model performance (observed vs. predicted values: $R^2 = 0.81$ when considering weakly and strongly migratory populations; $R^2 = 0.68$ when considering strongly migratory populations only) and diagnostics (Appendix S10) were robust and consistent with those of the model including the nonbreeding latitude.

Appendix S9. *Analysis excluding weakly migratory populations*

Our phylogenetic mixed models were re-run by excluding the geographical populations classified as weakly migratory. For the model without moderators, the results confirmed that migratory connectivity was moderate and significantly larger than zero (estimated $Zr = 0.397 \pm 0.083$ SE, $t_{82} = 4.769$, $P < 0.001$, corresponding to $r_M = 0.377$, 95% CI: 0.229 – 0.510; Fig. 2 in main text). In this analysis, we found no significant phylogenetic signal in the strength of migratory connectivity ($H^2 = 0.146$, $\chi^2 = 2.543$, $df = 1$, $P = 0.111$).

For the model with moderators, results confirmed the negative effect of migration distance, the positive effect of population spread and the negative effect of the mean nonbreeding latitude on the strength of migratory connectivity (Table S4a), whereas the quadratic effect of the ratio between the breeding and nonbreeding population spread and the effect of habitat breadth were not significant (Table S4a). Also in this case, there was no phylogenetic signal on migratory connectivity ($H^2 = 0.086$, $\chi^2 = 1.006$, $df = 1$, $P = 0.316$), and there was significant residual heterogeneity ($Q_E = 2490.710$, $df = 140$, $P < 0.001$). Replacing the nonbreeding latitude with the annual resource surplus available to birds (Appendix S8) provided qualitatively identical results, showing that connectivity increased with better access to resources (Table S4b). Both the alternative models showed good performance (observed vs. predicted values: $R^2 = 0.66$ and $R^2 = 0.69$, respectively) and the diagnostics indicated that no relevant deviation from assumptions occurred (Appendix S10).

Table S4. (a) Parameters estimated from the phylogenetic mixed model explaining the strength of migratory connectivity (as Fisher Z-transformation of r_M value) across 150 strongly migratory populations of 83 species. Transformations were applied to ‘Migration distance’ and ‘Nonbreeding population spread’ (Appendix S6-S7), while the second order polynomial term of ‘Relative population spread’ was included to account for quadratic effects (Table 1). All moderators are mean-centered and scaled to 1 SD. An asterisk marks significant ($P < 0.05$) moderators. In (b), ‘Nonbreeding latitude’ is replaced with ‘Annual resource surplus (NDVI)’ and the model was fitted to data on land birds only (Appendix S8, for details).

	Moderator	Coefficient	SE	t	df	P	
a.	Intercept	0.404	0.042	9.522	73	< 0.001	*
	Migration distance (km)	-0.377	0.014	-26.957	140	< 0.001	*
	Nonbreeding latitude (°)	-0.038	0.012	-3.184	140	0.002	*
	Nonbreeding population spread (km)	0.044	0.017	2.494	140	0.014	*
	Relative population spread	-0.006	0.023	-0.265	140	0.791	
	Relative population spread ²	0.003	0.026	0.098	140	0.918	
	Body mass (kg)	-0.012	0.023	-0.515	73	0.608	
	Habitat diversity	0.035	0.024	1.444	73	0.153	
	Diet diversity	0.027	0.021	1.282	73	0.204	
	Passerine	-0.009	0.030	-0.282	73	0.779	
b.	Intercept	0.438	0.036	12.229	56	< 0.001	*
	Migration distance (km)	-0.465	0.023	-20.142	101	< 0.001	*
	Annual resource surplus (NDVI)	0.051	0.015	3.403	101	0.001	*
	Nonbreeding population spread (km)	0.098	0.025	3.903	101	< 0.001	*
	Relative population spread	0.019	0.028	0.671	101	0.503	
	Relative population spread ²	-0.013	0.032	-0.403	101	0.688	
	Body mass (kg)	-0.012	0.029	-0.405	56	0.687	
	Habitat diversity	0.042	0.027	1.520	56	0.134	
	Diet diversity	0.025	0.025	1.011	56	0.316	
	Passerine	-0.017	0.032	-0.542	56	0.590	

Appendix S10. Model diagnostics

Although this study does not represent a meta-analysis, as information were not retrieved from the literature, we applied a set of techniques typical of meta-analyses to assess the robustness of the results. Indeed, these data are, essentially, based on ringing data, and it is well known that some species or some geographical populations may be under or overrepresented due to the large temporal and spatial variability in encounter probability. We argued that these processes may generate biases in the data similar to those deriving from publication bias in typical meta-analyses and we thus assessed the robustness of the results. Rosenthal's fail-safe number was always very large in all the analyses (all geographical populations: 1,787,814; $P < 0.001$; all geographical populations of land bird species: 1,241,064; $P < 0.001$; strongly migratory populations only: 561,658; $P < 0.001$; strongly migratory populations of land bird species only: 338,970; $P < 0.001$). In addition, funnel plots (Fig. S6) were rather symmetric (rank correlation test for funnel plot asymmetry; all geographical populations: Kendall's tau = 0.061, $P = 0.208$; all geographical populations of land bird species: Kendall's tau = 0.046, $P = 0.416$; strongly migratory populations only: Kendall's tau = 0.085, $P = 0.123$; strongly migratory populations of land bird species only: Kendall's tau = 0.030, $P = 0.641$), and the intercepts of Egger's regressions were not significant (all geographical populations: $t_{189} = -0.230$, $P = 0.819$; all geographical populations of land bird species: $t_{143} = 0.531$, $P = 0.596$; strongly migratory populations only: $t_{148} = 0.153$, $P = 0.879$; strongly migratory populations of land bird species only: $t_{109} = 1.205$, $P = 0.231$). These diagnostics thus indicate that the results of our analyses are robust.

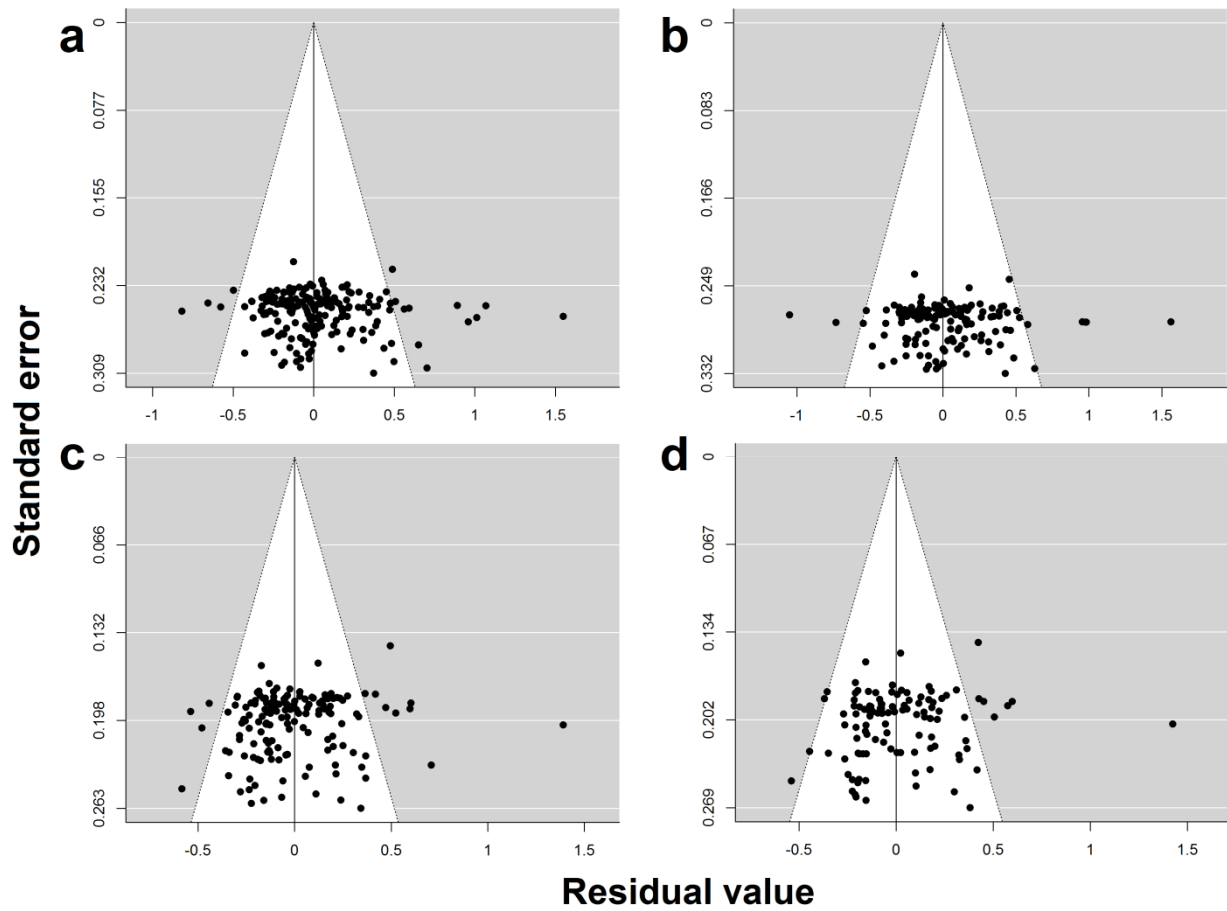


Figure S6. Funnel plots for the analyses conducted on all geographical populations (a), all geographical populations of land bird species (b), strongly migratory populations (c) and strongly migratory populations of land bird species (d).

Appendix S11. *Analysis testing the effect of dispersal ability on migratory connectivity*

Several weakly migratory populations showed a weaker migratory connectivity than that expected for their relatively short mean migration distance (e.g. *Alca torda*, *Fulica atra*, *Panurus biarmicus*; see Fig. 2, main text), implying high seasonal mobility within population range for these species. To evaluate whether this pattern may be due to some degree of dispersal (e.g. post-breeding dispersal), we re-run our model concerning both strongly and weakly migratory populations by including an additional predictor that would be able to account for interspecific differences in dispersal ability. We used the species-specific data on avian dispersal ability available from Sheard et al. (2021). Considering 10,338 bird species, this study has shown that avian dispersal ability was strongly predicted by the hand-wing-index (HWI), as a standardised biometric index of dispersal (see Sheard et al. 2021, and references therein). When we included this predictor, our model's results were unaffected, and HWI did not influence the strength of migratory connectivity (coefficient estimate \pm SE: 0.008 ± 0.034 ; $t_{72} = 0.228$; $P = 0.820$; observed vs predicted values: $R^2 = 0.80$).

Appendix S12. *Sensitivity analysis according to spatially-uneven sampling of ringing-recoveries in the nonbreeding ground and reduced sample size*

For each species, we performed a sensitivity analysis by rarefying the overall sample of individuals according to a spatially stratified subsampling, to assess whether our estimates of migratory connectivity were affected by uneven sampling in the nonbreeding range. Given that reporting rates of ring recoveries (i.e., sampling effort) differ between countries (Korner-Nievergelt et al. 2010), the spatial strata were represented by countries to account for geographic variation in the sampling effort. We performed a subsampling where the sampling probability was inversely proportional to the number of observations recorded in each country in the nonbreeding range. Thus, our stratified sampling assigned a lower sampling probability to countries with a greater number of records in the nonbreeding period, testing the robustness of migratory connectivity analysis in a situation corresponding to a spatially unbalanced sampling in the nonbreeding range. For each species, the sample of records was rarefied at the minimum possible sample size used in our analysis, i.e. 30 individuals. For two species having 30 individuals in the original dataset, the sample was rarefied at 29 individuals. Individuals were drawn from the original dataset without replacement. For each species, we simulated 100 different datasets and performed the migratory connectivity analyses. Simulations were implemented in R 3.6.2. We investigated the potential effects of uneven sampling in the nonbreeding range coupled with reduced sample size on the estimate of migratory connectivity strength, i.e. the Mantel correlation coefficient (r_M), as well as on the power of the Mantel permutation test (i.e., the relative frequency of simulations showing significant connectivity, if the analysis on the whole sample of that species was significant; the latter included $N = 78$ species). For each simulated dataset, we calculated Δr_M as the absolute value of the difference between the r_M value obtained from the simulation and that obtained from the original dataset. For each species, we averaged Δr_M values across the simulated datasets (to obtain mean Δr_M) and calculated the power of the Mantel test as the proportion of simulations showing significant migratory connectivity. Hence, we tested whether mean Δr_M and the power were associated to the mean nonbreeding latitude ($^\circ$). If low sample size coupled with uneven sampling in the nonbreeding range tends to affect migratory connectivity estimates and the power of the analysis, both Δr_M and the power should be correlated to nonbreeding latitude, indicating larger errors and/or lower power at the southernmost or northernmost latitudes.

For most (68%) species having their sample size reduced to the minimum number of individuals with uneven sampling in the nonbreeding range, we found that the mean Δr_M was still below 0.15 (corresponding to an error of about 15% on the possible range of r_M). Also, for 70% species among those showing significant connectivity, the power of the Mantel test was still above the 80% threshold. Most importantly, neither the mean Δr_M nor the power were correlated to nonbreeding latitude (Δr_M : $r = 0.15$, $P = 0.160$; power: $r = 0.15$, $P = 0.202$), suggesting that biases were unlikely to occur more often for species wintering farther south or farther north.

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