

1 **Response to microplastic exposure: an exploration into the sea urchin**
2 **immune cell proteome**

3 Carola Murano^{1§}, Simona Nonnis^{2,3§}, Francesca Grassi Scalvini², Elisa Maffioli², Ilaria Corsi⁴,
4 Gabriella Tedeschi^{2,3}, Anna Palumbo^{5*}

5

6 ¹Department of Integrative Marine Ecology, Stazione Zoologica Anton Dohrn, Naples, Italy

7 ²Department of Veterinary Medicine and Animal Science (DIVAS), Università degli Studi di Milano,
8 Milano, Italy

9 ³CRC "Innovation for well-being and environment" (I-WE), Università degli Studi di Milano,
10 Milano, Italy

11 ⁴Department of Physical, Earth and Environmental Sciences, University of Siena, Siena, Italy

12 ⁵Department of Biology and Evolution of Marine Organisms, Stazione Zoologica Anton Dohrn,
13 Naples, Italy

14

15 § Coauthors- Carola Murano and Simona Nonnis contributed equally

16

17 *Corresponding author

18 Anna Palumbo, Department of Biology and Evolution of Marine Organisms, Stazione Zoologica
19 Anton Dohrn, Naples, Italy. Email: anna.palumbo@szn.it

20

21

22

23

24

25

26

27

28

29

30

31

32

33

34 **Abstract**

35 It is now known that the Mediterranean Sea currently is one of the major hotspot for microplastics
36 (MPs; < 5mm) pollution and that the risks will be even more pronounced in the coming years. Thus,
37 the in-depth study of the mechanisms underlying the MPs toxicity in key Mediterranean organisms,
38 subjected to high anthropic pressures, has become a categorical imperative to pursue. Here, we
39 explore for the first time the sea urchins immune cells profile combined to their proteome upon *in*
40 *vivo* exposure (72h) to different concentrations of polystyrene-microbeads (micro-PS) starting from
41 relevant environmental concentrations (10, 50, 10³, 10⁴ MPs/L). Every 24h, immunological
42 parameters were monitored. After 72h, the abundance of MPs was examined in various organs and
43 coelomocytes were collected for proteomic analysis based on a shotgun label free proteomic
44 approach. While sea urchins treated with the lowest concentration tested (10 and 50 micro-PS/L) did
45 not show the presence of micro-PS in any tissue, in the specimens exposed to the highest
46 concentration (10³ and 10⁴ micro-PS) there was an internalisation of 9.75 ± 2.75 and 113.75 ± 34.5
47 MPs/g, respectively. Proteomic analyses revealed that MPs exposure altered coelomocytes protein
48 profile not only compared to the control group but also among the different micro-PS concentrations
49 and these variations are micro-PS concentration dependent. The proteins exclusively expressed in the
50 coelomocytes of specimens exposed to MPs are mainly metabolite interconversion enzymes, involved
51 in cellular processes, indicating a severe alteration of the cellular metabolic pathways. Overall, these
52 findings provide new insights on the mode of action of MPs in the sea urchin immune cells both at
53 the molecular and cellular level.

54

55

56

57

58

59 **Keywords:** coelomocytes; microplastics; oxidative stress; nitrosative stress; proteomics; sea urchin

60

61

62

63

64

65

66

67 **1. Introduction**

68 It is commonly believed that the Mediterranean Sea became over the years the sixth gyre with
69 accumulation of plastics debris due to the semi-enclosed geographical configuration of the Atlantic
70 Ocean with limited outlet flow (Còzar et al., 2015; Suaria et al., 2016; Macias et al., 2019; Everaert
71 et al., 2020). In fact, approximately 229 thousand tons of plastics items are expected to end up in the
72 Mediterranean every year (Boucher & Bilard, 2020). More deeply troubling are the recent estimates
73 which have identified the presence of approximately 3.2×10^{12} - 28.2×10^{12} plastic particles floating
74 on the surface of Mediterranean Sea, known as microplastics (MPs, < 5mm) (van Sebille et al., 2015;
75 Suaria et al., 2016). During their permanence in sea water, MPs undergo different transformations
76 involving their surface properties such as shape, roughness, and charge. In addition, MPs interact
77 with the surrounding environment, bacteria and chemical contaminants that contribute in determining
78 their ecological impacts driving the rapid changes in seawater (Galloway et al., 2017). Actually, most
79 of these MPs seems to be accumulated especially on the coastlines and on the sea bottom, this latter
80 probably representing the long term- sinks for MPs (Fries et al., 2013; Còzar et al., 2014; Nuelle et
81 al., 2014; Courtene-Jones et al., 2017). Thus, in the benthic environment of coastal areas, the
82 concentrations of MPs are much higher than in the rest of the basin due also to the close proximity to
83 the potential sources (Soto-Navarro et al., 2021).

84 In light of the extraordinary richness of the Mediterranean Sea as a biodiversity hotspot, such levels
85 of MPs highlight the potential risk for marine species up to marine coastal areas and benthic
86 environments. Moreover, model-based studies evidenced that species with smaller home ranges are
87 more likely to be exposed to plastic particles compared to species with larger home ranges (Compa
88 et al., 2019). Until now, the presence of MPs has been recorded in different Mediterranean benthic
89 species such as sea squirts, sea cucumbers, sea urchins, clams, oysters and mussels up to reaching
90 concentrations of 23 MPs/individual, mostly fibres and fragments mainly of polyacrylamide and
91 polyethylene (Vered et al., 2019; Bulleri et al., 2021; Expòsito et al., 2022; Murano et al., 2022).

92 Over the years, growing attention was paid on the potential effects of MPs on marine organisms.
93 Different laboratory studies have been performed both *in vivo* and *in vitro* evaluating different
94 endpoints for neurotoxicity, immunotoxicity, embryonic development, cytotoxicity, ingestion and
95 egestion (Barboza et al., 2018; Gambardella et al., 2018; Tang et al., 2020; Capolupo et al., 2021). In
96 most of these studies, polystyrene was chosen as proxy for MPs due to the fact that it's one of the
97 most largely used non-biodegradable plastic worldwide and unlike other polymers, it shows a greater
98 stability in sea water suspension with low styrene release (Cohen et al., 2002; Messinetti et al., 2018).
99 Overall, the lethal and sub-lethal effects of MPs on various marine organisms, belonging to different

100 trophic levels, have highlighted the involvement of oxidative stress and pathways activation,
101 including inflammatory responses (Hu & Palic, 2020).

102 However, in spite of the growing concern about MPs pollution, there are still several scientific issues
103 to be addressed, especially related to the molecular mechanisms and the cellular processes that are
104 activated in response to MPs exposure. The development of advanced "omics" technologies has
105 allowed to explore the effects of numerous contaminants or natural toxins on the proteome of the
106 different organisms examined through both in field and in laboratory experiments (Su et al., 2019;
107 Balbi et al., 2021; Sánchez-Marin et al., 2021). As a matter of fact, proteomic approaches lead to a
108 deeper comprehension of the response mechanisms against environmental stressors providing
109 important information on the abundance of proteins, key players of the biological processes (Gouveia
110 et al., 2019; Liang et al., 2020). To the best of our knowledge, very few studies have analysed the
111 effects of MPs on benthic organisms through proteomics techniques. For instance, Teng and co-
112 workers (2021) demonstrated that the oyster gland proteome was significantly affected by MPs
113 exposure, PE and PET (10 and 1000 $\mu\text{g L}^{-1}$), especially in terms of cytoskeleton organisation,
114 metabolic processes, signal transduction and protein synthesis. Green et al. (2019) discovered that
115 both the conventional plastic HDPE, as well as the biodegradable alternative, PLA (both at 2.5 $\mu\text{g L}^{-1}$
116 and 25 $\mu\text{g L}^{-1}$), greatly affected the proteome of mussels haemolymph, the main effector of the innate
117 immune response, after *in vivo* exposure.

118 Cell-mediated immune response can be considered as the first target of contaminant-exerted toxicity
119 in aquatic organisms. Among the benthic invertebrates of the Mediterranean Sea, the sea urchin
120 *Paracentrotus lividus* stands out for its intrinsic and unique immune system (Smith et al., 2018). This
121 peculiar immune system exhibits very similar features of the non-adaptive system or the innate system
122 of vertebrates and can be considered the baseline of the original deuterostome ancestor (Smith &
123 Davidson, 1992; Pancer et al., 1999). The innate immunity system of sea urchin is a very complex
124 network based on cellular and humoral factors which together are able to counteract pathogens,
125 foreign substances and other kinds of environmental challenges (Smith, 2010). In the coelomic cavity,
126 the immune cells, named as coelomocytes, circulating freely in the coelomic fluid also reaching
127 tissues and organs, control the cellular response (Buckley & Rast, 2019). In recent years, these cells
128 generated great interest as prominent biosensors for environmental monitoring not only for their
129 features but also for the sea urchin's strategic phylogenetic position. Indeed, thanks to the availability
130 of the full sea urchin genome sequence (Sea Urchin Genome Sequencing Consortium, 2006), an
131 extraordinary and also unexpected relationship to humans was disclosed. In particular, these strong
132 similarities between sea urchin and humans mainly involve the immune system in terms of alternative

133 adaptive and anticipatory immune functions (Hibino et al., 2006; Rast et al., 2006). Being recognised
134 as sensitive tools for investigating sea urchin health status, coelomocytes have been used for studying
135 the effects of environmental conditions, such as ocean acidification, marine pollution including
136 emerging contaminants as MPs and nanoplastics (Falugi et al., 2012; Pinsino et al., 2015; Marques-
137 Santos et al., 2018; Migliaccio et al., 2019; Alijagic et al., 2020; Milito et al., 2020; Murano et al.,
138 2020; Murano et al., 2021a). As evidence of this, field observations have shown that *P. lividus* is
139 currently a species subject to contamination by MPs in coastal marine areas and in particular by
140 microfibers, fragments and films and also plasticizers (Murano et al., 2022; Raguso et al. 2022). In
141 this context and considering the knowledge-gap on immunological responses upon MPs exposure,
142 this study aims to explore for the first time the sea urchin's immune cells profile combined to their
143 proteome upon *in vivo* exposure using different concentrations of polystyrene-microplastics starting
144 from relevant environmental concentrations.

145 **2. Methods**

146 *2.1 Sea urchin's collection and handling*

147 Adult specimens of *P. lividus* (Lamarck, 1816) (diameter 5.32 ± 0.51 cm) were collected from a coastal
148 site (40°42.335' N; 13°57.351' E) in the Gulf of Naples by the scuba diving staff of the Stazione
149 Zoologica Anton Dohrn of Naples. This site is not privately-owned nor protected in any way,
150 according to the authorisation of Marina Mercantile (DPR 1639/68, 09/19/1980, confirmed on
151 01/10/2000). Once in laboratory, sea urchins were acclimated for one weeks in glass tanks filled with
152 circulating natural seawater (NSW) (temperature 17.7 ± 1.4 °C, salinity 40 ± 1 , dissolved O₂ 7 mg/L,
153 pH 8.2; all the parameters remained constant during the experiment) and fed *ad libitum* with *Ulva*
154 *lactuca*.

155 *2.2 Polystyrene microbeads*

156 Fluorescent-labelled polystyrene microbeads (45 µm micro-PS) (441excitation/485 emission) were
157 purchased from Polysciences (Warrington, PA, U.S.A.). According to the supplier, the particles were
158 packaged as 2.5% aqueous suspension (5×10^5 micro-PS/mL) without biocides or stabilisers. The
159 beads size is in agreement with the data sheet provided by the manufacturer and the polystyrene
160 composition of the particles was previously confirmed by FTIR analysis (please refer to Murano et
161 al., 2021b). Micro-PS working solutions (10^4 particles mL⁻¹ and 10^2 particles mL⁻¹) were prepared in
162 deionised water and then added to filtered natural sea water (NSW, 0.22 µm) to reach the final
163 concentration of 10, 50, 10^3 and 10^4 particles L⁻¹. Stock and working solutions were vortexed for 3
164 min prior to use.

165 2.3 *Sea urchin's in vivo exposure*

166 Specimens of adult sea urchin were placed in 5L glass tanks (ratio of 1 specimen per liter) supplied
167 with filtered NSW (0.22 μm) from the coastal site in the Gulf of Naples in a closed flow-through
168 system constantly aerated. Micro-PS exposed treatments (10, 50, 10^3 and 10^4 particles L^{-1}) were set
169 up by directly adding the micro-PS into the tanks from the previously prepared and vortexed stock
170 solutions. The sea urchins were exposed to NSW only (control) and to micro-PS for 72h, during
171 which the urchins were not fed.

172 2.4 *Micro-PS internalisation*

173 The micro-PS extraction from sea urchin's tissues/organs was set according to our previous studies
174 Murano et al. (2020, 2021b). Briefly, 1M KOH was added to the tissues/organs (digestive system,
175 water vascular system and gonads) under examination (1:20, w:v) in glass flasks and the samples
176 were then slowly stirred for 48h at room temperature. After 48h, the digested solution was filtered
177 through a vacuum system on cellulose acetate membrane filters (0.45 μm) and then analysed under
178 optical microscopy to quantify micro-PS.

179 2.5 *Immune Cell Response*

180 The impact of micro-PS exposure on the sea urchin immune system was evaluated by measuring
181 different immunological parameters. In brief, after 24, 48, 72 h from incubation, the coelomic fluid
182 was collected through the coelomic cavity using a sterile syringe (5 mL, needle 26 gauge) pre-loaded
183 with the anticoagulant solution CCM 2X (NaCl 1M, MgCl_2 10 mM, EGTA 2 mM, Hepes 40 mM,
184 pH 7.2) at a ratio of 1:1 (anticoagulant: coelomic fluid) as reported in Murano et al. (2021b).
185 Heterogeneous coelomocytes were counted using a Neubauer counting chamber (Bright-Line
186 Hemacytometer) under light microscope (ZEISS Apotome.2). At different times, aliquots of the
187 coelomic fluid were washed twice in CCM 1X and the pellet was stored at -80°C until levels of
188 Reactive Oxygen Species (ROS), Reactive Nitrogen Species (RNS) and Total Antioxidant Capacity
189 (TAC) were assessed.

190 Using two probes, DCFH-DA (2,7-dichloro dihydro-fluorescein-diacetate) and DAF-DA (4-amino-
191 5-methylamino-2,7-dichlorofluorescein-diacetate), we specifically measured the intracellular levels
192 of ROS and RNS respectively. From the aliquots, about $1.5 \cdot 10^6$ coelomocytes were isolated and
193 subsequently exposed for 1h in the dark to a concentration of 20 μM DCFH-DA/ DAF-DA in
194 anticoagulant cells/mL. After exposure and different washing, the cells were sonicated in Tris-HCl
195 (40 mM- pH 7.4) and centrifuged for 10 min at 8000 rcf at $+4^\circ\text{C}$. The supernatant was harvested,
196 and the fluorescence was measured using the spectrofluorometer (Tecan) at ex 488/em 525 nm for

197 DCFH-DA and at ex 495/em 515 nm for DAF-DA. Fluorescence values were normalised by
198 subtracting the autofluorescence of unlabelled extracts (DMSO) and results are expressed as
199 fluorescence intensity referred to $1.5 \cdot 10^6$ coelomocytes. The remaining aliquots were used to analyse
200 the antioxidant capacity by exploiting the ability of hydrogen-donating antioxidants of the cells to
201 induce a de-coloration of the pre-formed radical cation of 2,2-azinobis-3-ethylbenzothiazoline-6-
202 sulfonic acid (ABTS●+) during the reaction between ABTS and H₂O₂ in the presence of peroxidase.
203 The TAC is quantified by measuring the absorbance at 730 nm using as a reference the standard curve
204 of ascorbic acid (1-15 μM) and then the values were normalised versus total protein content. Total
205 proteins were measured according to Bradford (1976) at 595 nm.

206 *2.6 Proteomic analysis by a shotgun label free approach*

207 The coelomocytes, obtained by centrifugation of the coelomic fluid in CCM 1X as reported above,
208 were collected from animals exposed to different concentrations of micro-PS and from animals kept
209 as control and analysed by a shotgun label free proteomic approach for the identification and
210 quantification of expressed proteins. The samples were resuspended in urea 8 M /Hepes 20 mM pH
211 8.0 containing protease inhibitors cocktail (Roche), sonicated using an ultrasonic probe in bursts of
212 20–30 s and centrifuged at 16560 x g for 15 min at 16 °C to pellet the tissue debris as previously
213 reported (Mortarino et al., 1998). The protein content was determined by the Bradford assay with
214 bovine serum albumin as standard. Prior to proteolysis, proteins were reduced with 13 mM
215 dithioerythriol (DTE; 15 min at 50 °C) and alkylated with 26 mM iodoacetamide (IAA; 30 min at
216 room temperature, in the dark). Protein digestion was performed using sequence-grade trypsin
217 (Promega) for 16 h at 37 °C using a protein: enzyme ratio of 20:1. The collected peptides were
218 desalted using Zip-Tip C18 before mass spectrometric (MS) analysis as reported in Eberini et al.
219 (2002). NanoHPLC coupled to MS/MS analysis was performed on Dionex UltiMate 3000 directly
220 connected to an Orbitrap Fusion Tribrid mass spectrometer (Thermo Fisher Scientific, Waltham, MA,
221 USA) by a nanoelectrospray ion source. Peptide mixtures were enriched on 75μm ID × 150 mm
222 EASY- Spray PepMap RSLC C18 column (Thermo Fisher Scientific) and separated using the LC
223 gradient: 4% ACN in 0.1% formic acid for 3 min, 4–28% ACN in 0.1% formic acid for 100 min, 28–
224 40% ACN in 0.1% formic acid for 10 min, 40–95% ACN in 0.1% formic acid for 1 min and 95–4%
225 ACN in 0.1% formic acid for 3 min at a flow rate of 0.3 μL/min. Orbitrap-MS spectra of eluting
226 peptides were collected over an m/z range of 375–1500 at resolution of 120000, operating in a data-
227 dependent mode with a cycle time of 3 s between master scans. HCD MS/MS spectra were acquired
228 in Orbitrap at resolution of 15000 using a normalised collision energy of 35%, and an isolation

229 window of 1.6 m/z. Dynamic exclusion was set to 60 s. Rejection of +1 and unassigned charge states
230 were enabled.

231 The mass spectrometry proteomics data have been deposited to the ProteomeXchange Consortium
232 via the PRIDE (Vizcaino et al., 2016) partner repository, with the dataset identifier PXD033665.

233 A database search was conducted against the Uniprot *Strongylocentrotus purpuratus* database (34417
234 entries) (<https://www.uniprot.org/proteomes>, release 11/02/22) with MaxQuant (version 1.6.1.0)
235 software. The initial maximum allowed mass deviation was set to 10 ppm for monoisotopic precursor
236 ions and 0.5 Da for MS/MS peaks. Enzyme specificity was set to trypsin, defined as C-terminal to
237 Arg and Lys excluding Pro, and a maximum of two missed cleavages were allowed.
238 Carbamidomethylcysteine was set as a fixed modification, while Met oxidation and Asn/Gln
239 deamidation were set as variable modifications. Quantification in MaxQuant was performed using
240 the built-in label free quantification algorithms (LFQ) based on extracted ion intensity of precursor
241 ions. False protein identifications (1%) were estimated by searching MS/MS spectra against the
242 corresponding reversed-sequence (decoy) database. Statistical analysis was performed using the
243 Perseus software (version 1.5.5.3) (Nicastrò et al., 2015). Only the proteins present and quantified in
244 at least 75% of the repeats were positively identified in a sample and used for statistical analysis.
245 PCA was carried out by grouping quantitative data related to proteins in the coelomocytes from
246 control group (CTR) and from animals exposed to different micro-PS exposed treatments: 10 micro-
247 PS/L, 50 micro-PS/L, 10³ micro-PS/L and 10⁴ micro-PS/L, respectively. Focusing on specific
248 comparisons, proteins were considered differentially expressed if they were present only in one
249 condition or showed significant t-test difference (Welch's test $P \leq 0.05$). Bioinformatic analyses were
250 carried out by Panther software (release 17.0) (Mi et al., 2021) to classify the proteins in the various
251 data sets and to cluster enriched annotation groups of Biological Processes, Molecular Function,
252 Pathways, and Networks within the set of identified proteins. Functional grouping was based on
253 Fischer's exact test $P \leq 0.05$. Interaction networks were visualised using the "Search Tool for
254 Recurring Instances of Neighbouring Genes" (STRING) (Szklarczyk et al., 2021) setting the
255 minimum required interaction score at 0.7 and hiding disconnected nodes.

256 2.7 Statistical analysis

257 The data on quantitative and qualitative analysis of coelomocytes were analysed by two-way analysis
258 of variance ANOVA followed by Bonferroni's multiple comparisons test. Intracellular levels of
259 ROS/RNS and TAC were analysed by two-way analysis of variance (ANOVA) ($P < 0.05$) followed
260 by Tukey's multiple comparison test. Data are presented as mean \pm SD and statistics was performed
261 using GraphPad Prism version 7.00 for Windows.

262

263 **3. Results**

264 *3.1 Micro-PS uptake*

265 Micro-PS content analysis in the different organs of sea urchins showed a concentration dependent
266 uptake after 72 h exposure. While sea urchins treated with the lowest concentration tested (10 and 50
267 micro-PS/L) did not report the presence of micro-PS in any tissue/organ, the specimens exposed to
268 the highest concentration tested (10^3 and 10^4 micro-PS) revealed micro-PS internalisation of $9.75 \pm$
269 2.75 and 113.75 ± 34.5 particles normalised by the weight of fresh organs, respectively (Figure 1).
270 As expected, the digestive system includes the organs most affected by internalisation which exceeds
271 more than 70% of the total particles internalised in both cases, followed by gonads and esophagus.
272 Interestingly, only at the highest concentration used (10^4 micro-PS), micro-PS were found at the level
273 of the aquifer system (ring canal and ampullae) even if in a very small amount (total quantity below
274 the 6% of the total particles).

275 *3.2 Quantitative and qualitative analysis of coelomocytes*

276 Starting from 24h exposure, a statistically significant increase of the total coelomocytes count was
277 detected in specimens treated with 10^4 micro-PS/L compared to the control group ($4.6 \cdot 10^6 \pm 3.3 \cdot 10^5$
278 vs $3.2 \cdot 10^6 \pm 6.8 \cdot 10^5$ cells/mL, respectively) (Figure 2). This increase in specimens treated with 10^4
279 micro-PS/L occurred up to 72h. Conversely, after 24h of exposure, specimens treated with 10^3 micro-
280 PS/L displayed a significant decrease in total number of coelomocytes compared to the control
281 ($2.0 \cdot 10^6 \pm 3.0 \cdot 10^5$ vs $3.2 \cdot 10^6 \pm 6.8 \cdot 10^5$ cells/mL, respectively). But after 24h, the cell concentration
282 returned to be similar to the control. In the case of both low concentrations tested (10 and 50 micro-
283 PS/L), no alterations in total cell number was observed compared to control values.

284 Figure 3 shows the morphological profile of the sea urchin's immune cells along different
285 experimental groups. No variations between the different cell types was detected. However, the ratio
286 between red and white amoebocytes resulted affected by exposure to micro-PS. In fact, a ratio
287 between 0.5 and 1 values, indicative of an initial stressful condition, was found at different times of
288 exposure to micro-PS (10, 50, 10^3 , 10^4). A more harmful situation with a value greater than 1 was
289 detected after 48h of exposure to micro-PS 10^4 (Table 1).

290 *3.3 Oxidative stress status in coelomocytes*

291 By exploring the oxidative stress status of coelomocytes, increases in intracellular levels of both ROS
292 and RNS as well as variations in antioxidant capacity were recorded upon exposure to micro-PS. In

293 details, at 24h coelomocytes of specimens treated with 10^4 micro-PS/L showed a significant increase
294 of ROS levels compared to the control group ($2.5 \cdot 10^5 \pm 2.2 \cdot 10^3$ vs $1.5 \cdot 10^5 \pm 4.6 \cdot 10^3$ a.u.) which is still
295 evident at 48h (Figure 4A). At this exposure time, a significant increase of ROS levels was also
296 detected after treatment with 10 and 10^3 micro-PS/L ($3.1 \cdot 10^5 \pm 2.3$ and $2.1 \cdot 10^5 \pm 1.0$, respectively).
297 Then, at 72h, the ROS levels were comparable to control. In the case of intracellular RNS levels,
298 starting from 24h up to 72h, the coelomocytes of organisms treated with 10^4 micro-PS/L showed a
299 statistically significant increase compared to the control (72h: $3.9 \cdot 10^5 \pm 4.2 \cdot 10^3$ vs $1.2 \cdot 10^5 \pm 5.7 \cdot 10^3$)
300 (Figure 4B). On the other hand, for the other concentrations of microplastics, a significant increase
301 in RNS at 72h was observed in the case of treatment with 10 and 10^3 micro-PS/L ($2.5 \cdot 10^5 \pm 3.2 \cdot 10^3$
302 and $2.2 \cdot 10^5 \pm 6.2 \cdot 10^3$ a.u., respectively). At the highest concentration tested (10^4 micro-PS/L) no
303 alterations in the total antioxidant capacity of the coelomocytes was detected. At the lowest
304 concentration (10 micro-PS/L) some imbalance was revealed. In fact, at 24h a statistically significant
305 increase was first observed (1.20 ± 0.4 vs 0.70 ± 0.23 $\mu\text{mol eq. ascorbate}/\mu\text{g protein}$) followed by a
306 significant decrease at 48h compared to the control (0.3 ± 0.02 vs 0.96 ± 0.09 $\mu\text{mol eq. ascorbate}/\mu\text{g}$
307 protein, respectively) (Figure 4C). In addition, a decrease was detected at 48h and 72h after treatment
308 with 10^3 micro-PS/L (0.44 ± 0.22 and 0.27 ± 0.09 $\mu\text{mol eq. ascorbate}/\mu\text{g protein}$, respectively).

309 3.4 Impact of micro-PS on the proteome of coelomocytes

310 The coelomocytes were collected from specimens exposed to NSW (controls) and to those exposed
311 to micro-PS/L and analysed by a label free shotgun proteomic approach for the identification and
312 quantification of all the proteins expressed as previously described by Inguglia et al. (2020). In the
313 present study this approach allowed to identify 1323, 1305, 1152, 1275 and 1152 proteins in CTR,
314 10 micro-PS/L, 50 micro-PS/L, 10^3 micro-PS/L and 10^4 micro-PS/L, respectively, as reported in the
315 corresponding Supplementary Tables S1-S5. The proteomic analysis identified proteins common to
316 all samples as well as proteins differentially expressed in the different conditions as reported in the
317 Venn diagram of all the data sets (Figure 5A). The PCA analysis, reported in Figure 5B, clearly
318 showed a marked effect on the proteome of coelomocytes from animals exposed to MPs in
319 comparison to the control but also among different MPs concentrations. The result prompted us to
320 compare by Perseus the proteins expressed in the control (1323 entries) and all the proteins expressed
321 in the samples exposed to micro-PS/L (1341 entries) (Figure 6A). The comparison allowed to identify
322 1218 common proteins, 105 proteins expressed only in the CTR and 123 proteins exclusively
323 expressed in samples exposed to different concentrations of micro-PS/L (Table S6). These latter were
324 further analysed by Panther for protein classification, biological function (GOBP) and molecular
325 function (GOMF) as reported in Figure 6B. Based on the classification results, many proteins (34.2%)

326 are metabolite interconversion enzymes suggesting that micro-PS exposure heavily alter the
327 metabolism response of sea urchin. More than 11% are proteins involved in transport and trafficking
328 and almost 8% (7.6 %) are cytoskeletal proteins (Figure 6B). In keeping, the GOBP classification
329 shows that 12.1% are proteins involved in localisation suggesting a cytoskeleton reorganisation and
330 increased intracellular membrane trafficking (Figure 6B). In accordance, most of the coelomocytes
331 are of the large phagocytes class (up to 80% in *P. lividus*), characterised and described by an important
332 and complex cytoskeletal organisation.

333 The results are further confirmed by the Panther enrichment test analysis (Table S7, all micro-PS/L
334 vs CTR) that underlines in the 123 proteins analysed a statistically significant enrichment of proteins
335 involved in endosome transport via multivesicular body sorting pathway (35-fold enrichment) and in
336 establishment of protein localisation (6.7-fold enrichment). Proteins involved in catabolic processes
337 are also enriched in the analysis (5.3-fold enrichment).

338 To disclose the contribution of the micro-PS exposure concentration on the immune cells proteome,
339 specific analyses were carried out by one to one comparison: 10 micro-PS/L vs CTR, 50 micro-PS/L
340 vs CTR, 10^3 micro-PS/L vs CTR and 10^4 micro-PS/L vs CTR. According to the results, no proteins
341 increased or decreased were identified with a statistical significance. On the contrary, the analysis
342 allowed to identify either common proteins or proteins exclusively expressed in each condition as
343 shown in the Venn diagrams reported in Figure 7A.

344 The bioinformatic analysis by Panther (Table S7) highlights that, starting from 50 micro-PS/L
345 concentration on, there is an enrichment of terms related to protein transport, vacuole for ubiquitin-
346 dependent catabolic processes as well as proteins involved in the cell aerobic respiration (Figure 7B).
347 Upon increasing the micro-PS concentration at 10^3 and 10^4 micro-PS/L, a significant effect on the
348 cytoskeleton became evident with enrichment of proteins involved in cytoskeletal regulation,
349 cadherin signalling and integrin pathway. Some of these proteins are also involved in the
350 inflammation pathway mediated by chemokine and cytokine and some are components of the
351 endosomal sorting complexes required for transport complex assembly (ESCRT) and disassembly as
352 classified by STRING (Figure 7C). The ESCRT complex enables a membrane remodelling, resulting
353 in membranes bending/budding away from the cytoplasm. This machinery plays a vital role in a
354 number of cellular processes including multivesicular body (MVB) biogenesis, cellular abscission,
355 and is essential for cells to destroy misfolded and damaged proteins (Schmidt & Teis, 2012).

356 **4. Discussion**

357 Considering their dual chemical-physical nature facet, MPs represent a real challenge for marine
358 organisms, especially for the immune system which is the first line of defence. Thus, a proper immune

359 response plays a vital role for preventing immunological disorders possibly caused by MPs. In light
360 of this, the aim of this work was to study for the first time how exposure to different concentrations
361 of micro-PS could alter the immunological response in adult sea urchins both from a functional point
362 of view and more deeply at the level of the proteome.

363 *4.1 Microplastics exposure induces functional alterations in immune system*

364 Our data indicate that the *in vivo* waterborne exposure of sea urchins to different concentrations of
365 micro-PS reshapes the immunological profiles of the immune cells, coelomocytes. As expected, these
366 changes concern especially the highest concentrations tested of 10^4 micro-PS/L although at the lowest
367 concentration of 10 micro-PS/L the cellular homeostasis is altered. In detail, the total number of
368 coelomocytes was nearly similar across all treatments except at the highest concentration in which a
369 significant increase starting from 24h of exposure was observed, confirming our previous findings
370 (Murano et al. 2020). The increase of coelomocytes, and thus the proliferative response, is a
371 phenomenon that occurs in sea urchins usually in response to immunological challenges (Brockton
372 et al., 2008). This process could derive from an increased production of coelomocytes in coelomic
373 fluid or from an increase of their migration from surrounding tissues into the coelomic fluid
374 (Golconda et al., 2019). However, the absence of micro-PS in the coelomic cavity suggests that the
375 proliferation occurs in other organs such as the digestive system, whose internal walls are filled with
376 coelomocytes (Holland, 2020). In fact, the highest amount of micro-PS was recorded in the digestive
377 system at the two highest exposure concentrations both at 10^4 and 10^3 micro-PS/L. Interestingly, at
378 this last concentration, the number of coelomocytes in the coelomic cavity at 24h is lower than that
379 of the corresponding controls and then reaches the control levels in the following hours. In this case
380 it is possible to hypothesise that the momentary decrease is due to a migration from the coelomic
381 cavity. The cellular morphological profile confirms that in terms of composition, phagocytes
382 represent the majority of the immune cells, exceeding about 80% in all treatments and at all times.
383 Alterations mainly concern minor morphological cellular types, the red cells whose number increases
384 as reflected by the increased ratio between red and white amoebocytes, already detected at the lowest
385 concentration. This ratio reflects the fitness of the sea urchin's populations. Indeed, an increase in red
386 cells compared to white cells indicates that sea urchins are in a condition of injury (Matranga et al.,
387 2005; Pinsino & Matranga, 2015). For example, it was demonstrated that in pollution or hypoxia
388 conditions, after injuries or after MPs exposure, the homeostasis between red/white amoebocytes was
389 visibly affected (Matranga et al., 2002; Matranga et al., 2005; Pinsino et al., 2007; Suh et al., 2014;
390 Murano et al., 2020; Murano et al., 2021b). The common defence mechanisms used by most
391 invertebrates to face injuries include phagocytosis, encapsulation and production of ROS and nitrogen

392 radicals, RNS (Canesi & Procházková, 2014). Our results indicate that the highest concentration
393 tested caused a significant increase in both reactive species at all times compared to controls.
394 Nevertheless, the cells try to counteract low concentration MPs exposure already at 24h, as revealed
395 by the increase in the antioxidant system at 24h followed by a decrease at 48h and a concomitant
396 increase in ROS and RNS levels at 48h and 72h, respectively. Our finding of the modulation of redox
397 homeostasis starting from low concentrations certainly strengthens what it is commonly believed that
398 oxidative stress represents the universal common factor of toxicity caused by exposure to MPs (Hu
399 & Palic, 2020).

400 *4.2 Microplastics exposure affects the immune cell proteome*

401 An important outcome of this study concerns the use of the proteomic approach that allowed us to
402 demonstrate, for the first time, that the MPs *in vivo* exposure affects the immune cell proteome in sea
403 urchin. As also suggested by the PCA analysis, these variations are not only evident compared to the
404 control group but also among the different micro-PS exposure concentrations. While the highest
405 concentrations (10^3 and 10^4 micro-PS/L) cause similar changes, the lowest (10 and 50 micro-PS/L)
406 differ from each other and from the highest, indicating a concentration dependent proteomic profile.
407 The set of proteins exclusively expressed in the coelomocytes of specimens exposed to MPs are
408 mostly classified as metabolite interconversion enzymes, mainly involved in cellular processes,
409 indicating a severe alteration of the cellular metabolic pathways.

410 This most likely depends on the fact that the immune defence action against external factors such as
411 MPs includes the involvement of cellular detoxification processes which have a significant metabolic
412 cost (Guderley & Pörtner, 2010; Gardon et al., 2020).

413 In fact, it is well known that MPs have the ability to cause metabolic disorders in various organisms
414 belonging to different trophic levels (Paul-Pont et al., 2016; Kim et al., 2019; Magni et al., 2019;
415 Green et al., 2019; Qiao et al., 2019; Duan et al., 2021). For instance, Duan et al. (2021) demonstrated
416 that the haemolymph metabolic functions of the shrimp *Litopenaeus vannamei* were altered upon
417 long-term MPs exposure. Similarly, Green et al. (2019) showed that the MPs exposure (HDPE and
418 PLA) altered the abundance of metabolic as well as detoxification proteins in haemolymph of *M.*
419 *galloprovincialis*. In accordance, mitochondrial proteins involved in the energetic metabolism are
420 found among the differentially expressed proteins in coelomocytes of specimens exposed to micro-
421 PS (Table S6). Under normal environmental conditions, energy allocations in organisms remain
422 optimal (Sokolova, 2013), but in moderate environmental stress, such as upon MPs exposure, marine
423 organisms potentially compensate the increased energy requirements by increasing energy generation
424 and integration, as well as through metabolic regulation to meet the elevated ATP demands (Lannig

425 et al., 2010). This compensatory process is evident from our finding showing that proteins associated
426 with energetic metabolic activities are among the protein expressed upon micro-PS exposure and not
427 in CTR, like the mitochondrial ATP synthase d (Table S3), NADH dehydrogenase iron-sulfur protein
428 6 (Table S3, S4, S5, S6), cytochrome b-c1 complex subunit 8 (Table S3), cytochrome c domain-
429 containing protein (Table S4), cytochrome c oxidase and cytochrome b5 heme-binding domain-
430 containing protein (Table S5) and other proteins involved in redox activities, such as glutamate
431 dehydrogenase, oxoglutarate dehydrogenase, 3-hydroxyisobutyrate dehydrogenase and D-3-
432 phosphoglycerate dehydrogenase (Table S3, S4, S5, S6).

433 The increase in mitochondrial activity is related to the production of ROS that, under normal
434 conditions, is associated with homeostatic regulation (Manduzio et al., 2005) but, when
435 environmental conditions become unfavorable, tethers precariously towards an imbalance between
436 the generation and removal of ROS by antioxidants, leading to oxidative stress. This can either be
437 due to excessive ROS production or the impairment of antioxidant machinery, or both (Sheehan &
438 McDonagh, 2008). Concerning this aspect, proteomic data identified superoxide dismutase and
439 catalase among the proteins expressed upon exposure to micro-PS suggesting that the increase of
440 intracellular ROS in specimens treated for 48h (Figure 4A) may be counteracted by antioxidant
441 enzymes resulting in a decrease of ROS production upon 72h exposure (Figure 4A, Table S3, S4, S5,
442 S6).

443 Notoriously, the effects of oxidative stress and excess ROS are associated with cytoskeletal damage
444 with repercussions for processes such as protein transport and intracellular signaling (Raftos et al.,
445 2016). Interestingly, the enrichment analysis revealed a significant variation of the “Rho GTPase
446 pathway” only at the highest concentrations (10^3 - 10^4 micro-PS/L). As well known, Rho GTPases are
447 involved in different cellular functions mainly related to cytoskeletal alterations, intracellular
448 trafficking, cell migration, gene expression and cell proliferation (Li et al., 2015). Precisely, they
449 control intermediate filaments, the cortical actin cytoskeleton, myosin filaments and microtubules to
450 regulate immune cell migration, activation, proliferation and phagocytosis (Ridley, 2001). These
451 processes are supported by the action of the Wiskott-Aldrich syndrome protein family members,
452 which controls cell polarity, required to generate and maintain migration, synapse formation and
453 polarised cytokine secretion (El Masri & Delon, 2021). Notably, in this study these proteins were
454 found exclusively expressed in the high concentrations tested. The most representative cell population
455 of coelomocytes are phagocytes (>80 %) which occur in two morphotypes: petaloid and filopodial
456 (Pinsino & Matranga, 2015). The phagocytes appear to be the core of immunity also thanks to their
457 cytoskeleton that rearranges itself rapidly in association with immune response (Smith et al., 2006).
458 Under stressful conditions or in response to different stimuli, a petaloid-filopodial transition occurs

459 which involves a significant change in the cytoskeletal morphology switching from bladder-like
460 shape to filopodial (Smith et al., 2019). This phenomenon also occurs in other echinoderms such as
461 sea stars in which proteomic studies have shown that the transition is regulated mainly by two major
462 pathways: integrin signalling and the Rho GTPase (Franco et al., 2011; Andrade et al., 2021). These
463 observations clearly suggest that exposure to these high concentrations influences the cytoskeletal
464 rearrangement of immune cells inducing morphological changes predicting a condition of heavy
465 stress. In fact, the morphological transition was observed in sea urchins upon UV-B radiation and
466 post exposure to zinc and silver nanoparticles (Matranga et al., 2006; Pagliara & Stabili, 2012;
467 Magesky et al., 2016) or in sea stars as post-traumatic response (Pinsino et al., 2007). Moreover, in
468 the 10⁴ micro-PS/L treatment there is the activation of the cadherin and integrin pathways that are
469 closely linked to the Rho-GTPase in the cell-substratum and cell-cell adhesion processes,
470 respectively, suggesting that in this scenario the alteration is in a more advanced step with more severe
471 features (Fukata et al., 1999). At highest concentrations there is also the enrichment of “Huntington
472 disease (6 proteins cytoskeleton)”, a key regulator of several important phenomena, including
473 immune synapse organisation, viral infection, cell migration, and the transformation and degradation
474 of misfolded proteins (Valenzuela-Fernandez et al., 2008).

475 Interestingly, the enrichment of cadherin and integrin pathways was also described in the early life
476 stages of *Oryzias melastigma* exposed to the endocrine disruptors 17 α -ethinylestradiol and bisphenol
477 A (Bhandhari et al., 2020), suggesting a similar mechanism of action between endocrine disruptors
478 and MPs but more specifically with styrene oligomers (Choi et al., 2005).

479 The bioinformatic analysis by Panther of proteins differentially expressed in samples exposed to
480 different concentrations of micro-PS, highlights that more than 11% are involved in transport and
481 trafficking and 12.1% are involved in localisation. In addition, the analysis shows a significant
482 enrichment of pathways, such as “endosome transport via multivesicular body sorting” and “late
483 endosome to vacuole transport”, related to the formation of multivesicular endosomes, known as
484 compartments for receptor downregulation and intermediates in the formation of secretory lysosomes
485 (Raiborg et al., 2003). To date, this type of mechanism was reported in coelomocytes of *S. purpuratus*,
486 in response to LPS in which the upregulation in protein sorting and trafficking of transport vesicles
487 was highlighted through microarray analysis of coelomocyte gene expression (Nair et al., 2005).
488 More recently, the same mechanism was evidenced by a high throughput iTRAQ-based proteomics
489 methodology in haemocytes of the South African abalone *Haliotis midae*, as molecular response to
490 the exposure at acidic waters for prolonged periods, in which “Intracellular trafficking, secretion and
491 vesicular transport” was among the most enriched functional classes (Carroll & Coyne, 2021).

492

494 **5. Conclusions**

495 In a scenario of continuously pronounced global changes, the priority of understanding how MPs
 496 compromise marine organisms becomes more evident, particularly by elucidating the toxicity
 497 mechanisms of action in species highly exposed to several anthropogenic pressures in coastal areas
 498 such as *P. lividus*.

499 For the first time with this study, it was shown that exposure to micro-PS causes an immunological
 500 challenge such as to remodel the structural and functional components of the coelomocytes
 501 concurrently to the activation of particular cellular compartments involved in trafficking and protein
 502 sorting. Changes in the morphological profile of immune cells have been recorded during exposure
 503 to MPs as well as a promotion of the ROS and RNS formation starting already from the lowest
 504 concentrations.

505 **References**

- 506 Alijagic, A., Gaglio, D., Napodano, E., Russo, R., Costa, C., Benada, O., Kofroňová, O., Pinsino, A.,
 507 2020. Titanium dioxide nanoparticles temporarily influence the sea urchin immunological
 508 state suppressing inflammatory-related gene transcription and boosting antioxidant metabolic
 509 activity. *Journal of Hazardous Materials* 384, 121389.
 510 <https://doi.org/10.1016/j.jhazmat.2019.121389>
- 511 Andrade, C., Oliveira, B., Guatelli, S., Martinez, P., Simões, B., Bispo, C., Ferrario, C., Bonasoro,
 512 F., Rino, J., Sugni, M., Gardner, R., Zilhão, R., Coelho, A.V., 2021. Characterization of
 513 Coelomic Fluid Cell Types in the Starfish *Marthasterias glacialis* Using a Flow
 514 Cytometry/Imaging Combined Approach. *Frontiers in Immunology* 12.
 515 <https://doi.org/10.3389/fimmu.2021.641664>
- 516 Balbi, T., Auguste, M., Ciacci, C., Canesi, L., 2021. Immunological Responses of Marine Bivalves
 517 to Contaminant Exposure: Contribution of the -Omics Approach. *Frontiers in Immunology*
 518 12, 618726. <https://doi.org/10.3389/fimmu.2021.618726>
- 519 Barboza, L.G.A., Vieira, L.R., Guilhermino, L., 2018. Single and combined effects of microplastics
 520 and mercury on juveniles of the European seabass (*Dicentrarchus labrax*): Changes in
 521 behavioural responses and reduction of swimming velocity and resistance time.
 522 *Environmental Pollution* 236, 1014–1019. <https://doi.org/10.1016/j.envpol.2017.12.082>
- 523 Bhandari, R.K., Wang, X., Saal, F.S.V., Tillitt, D.E., 2020. Transcriptome analysis of testis reveals
 524 the effects of developmental exposure to bisphenol a or 17 α -ethinylestradiol in medaka
 525 (*Oryzias latipes*). *Aquatic Toxicology* 225, 105553.

526 <https://doi.org/10.1016/j.aquatox.2020.105553>

527 Boucher, J., Billard, G., 2020. The Mediterranean: Mare plasticum. Technical Report IUCN, Global
528 Marine and Polar Programme. Gland, Switzerland: IUCN, 62pp.
529 www.iucn.org/resources/publications

530 Bradford, M.M., 1976. A rapid and sensitive method for the quantitation of microgram quantities of
531 protein utilizing the principle of protein-dye binding. *Anal. Biochem.* 72, 248e254.
532 <https://doi.org/10.1006/abio.1976.9999>

533 Brockton, V., Henson, J.H., Raftos, D.A., Majeske, A.J., Kim, Y.-O., Smith, L.C., 2008. Localization
534 and diversity of 185/333 proteins from the purple sea urchin – unexpected protein-size range
535 and protein expression in a new coelomocyte type. *Journal of Cell Science* 121, 339–348.
536 <https://doi.org/10.1242/jcs.012096>

537 Buckley, K.M., Rast, J.P., 2019. Immune activity at the gut epithelium in the larval sea urchin. *Cell*
538 *Tissue Research* 377, 469–474. <https://doi.org/10.1007/s00441-019-03095-7>

539 Bulleri, F., Ravaglioli, C., Anselmi, S., Renzi, M., 2021. The sea cucumber *Holothuria tubulosa* does
540 not reduce the size of microplastics but enhances their resuspension in the water column.
541 *Science of The Total Environment* 781, 146650.
542 <https://doi.org/10.1016/j.scitotenv.2021.146650>

543 Canesi, L., Procházková, P., 2014. Chapter 7 - The Invertebrate Immune System as a Model for
544 Investigating the Environmental Impact of Nanoparticles, in: Boraschi, D., Duschl, A. (Eds.),
545 Nanoparticles and the Immune System. Academic Press, San Diego, pp. 91–112.
546 <https://doi.org/10.1016/B978-0-12-408085-0.00007-8>

547 Capolupo, M., Gunaalan, K., Booth, A.M., Sørensen, L., Valbonesi, P., Fabbri, E., 2021. The sub-
548 lethal impact of plastic and tire rubber leachates on the Mediterranean mussel *Mytilus*
549 *galloprovincialis*. *Environmental Pollution* 283, 117081.
550 <https://doi.org/10.1016/j.envpol.2021.117081>

551 Carroll, S.L., Coyne, V.E., 2021. A proteomic analysis of the effect of ocean acidification on the
552 haemocyte proteome of the South African abalone *Haliotis midae*. *Fish & Shellfish*
553 *Immunology* 117, 274–290. <https://doi.org/10.1016/j.fsi.2021.08.008>

554 Choi, J.O., Jitsunari, F., Asakawa, F., sun Lee, D., 2005. Migration of styrene monomer, dimers and
555 trimers from polystyrene to food simulants. *Food Additives and Contaminants* 22, 693–699.
556 <https://doi.org/10.1080/02652030500160050>

557 Cohen, J.T., Carlson, G., Charnley, G., Coggon, D., Delzell, E., Graham, J.D., Greim, H., Krewski,
558 D., Medinsky, M., Monson, R., Paustenbach, D., Petersen, B., Rappaport, S., Rhomberg, L.,
559 Ryan, P.B., Thompson, K., 2002. A comprehensive evaluation of the potential health risks

560 associated with occupational and environmental exposure to styrene. *Journal of Toxicology*
561 *and Environmental Health, Part B* 5, 1–263. <https://doi.org/10.1080/10937400252972162>

562 Compa, M., Alomar, C., Wilcox, C., van Sebille, E., Lebreton, L., Hardesty, B.D., Deudero, S., 2019.
563 Risk assessment of plastic pollution on marine diversity in the Mediterranean Sea. *Science of*
564 *The Total Environment* 678, 188–196. <https://doi.org/10.1016/j.scitotenv.2019.04.355>

565 Courtene-Jones, W., Quinn, B., Gary, S.F., Mogg, A.O.M., Narayanaswamy, B.E., 2017.
566 Microplastic pollution identified in deep-sea water and ingested by benthic invertebrates in
567 the Rockall Trough, North Atlantic Ocean. *Environmental Pollution* 231, 271–280.
568 <https://doi.org/10.1016/j.envpol.2017.08.026>

569 Cózar, A., Echevarría, F., González-Gordillo, J.I., Irigoien, X., Úbeda, B., Hernández-León, S.,
570 Palma, Á.T., Navarro, S., García-de-Lomas, J., Ruiz, A., Fernández-de-Puelles, M.L., Duarte,
571 C.M., 2014. Plastic debris in the open ocean. *Proceedings of the National Academy of*
572 *Sciences U.S.A.* 111, 10239–10244. <https://doi.org/10.1073/pnas.1314705111>

573 Cózar, A., Sanz-Martín, M., Martí, E., González-Gordillo, J.I., Ubeda, B., Gálvez, J.Á., Irigoien, X.,
574 Duarte, C.M., 2015. Plastic Accumulation in the Mediterranean Sea. *PLoS ONE* 10,
575 e0121762. <https://doi.org/10.1371/journal.pone.0121762>

576 Duan, Y., Xiong, D., Wang, Y., Zhang, Z., Li, H., Dong, H., Zhang, J., 2021. Toxicological effects
577 of microplastics in *Litopenaeus vannamei* as indicated by an integrated microbiome,
578 proteomic and metabolomic approach. *Science of The Total Environment* 761, 143311.
579 <https://doi.org/10.1016/j.scitotenv.2020.143311>

580 Eberini, I., Calabresi, L., Wait, R., Tedeschi, G., Pirillo, A., Puglisi, I., L., Sirtori, C.R., Gianazza, E.,
581 2002. Macrophage metalloproteinases degrade high-density-lipoprotein-associated
582 apolipoprotein A-I at both the N- and C-termini. *Biochemical Journal* 362, 627–634.
583 <https://doi.org/10.1042/bj3620627>

584 El Masri, R., Delon, J., 2021. RHO GTPases: From new partners to complex immune syndromes.
585 *Nature Reviews Immunology* 21, 499–513. <https://doi.org/10.1038/s41577-021-00500-7>

586 Everaert, G., De Rijcke, M., Lonneville, B., Janssen, C.R., Backhaus, T., Mees, J., van Sebille, E.,
587 Koelmans, A.A., Catarino, A.I., Vandegehuchte, M.B., 2020. Risks of floating microplastic
588 in the global ocean. *Environmental Pollution* 267, 115499.
589 <https://doi.org/10.1016/j.envpol.2020.115499>

590 Expósito, N., Rovira, J., Sierra, J., Gimenez, G., Domingo, J.L., Schuhmacher, M., 2022. Levels of
591 microplastics and their characteristics in molluscs from North-West Mediterranean Sea:
592 Human intake. *Marine Pollution Bulletin* 181, 113843.
593 <https://doi.org/10.1016/j.marpolbul.2022.113843>

594 Falugi, C., Aluigi, M.G., Chiantore, M.C., Privitera, D., Ramoino, P., Gatti, M.A., Fabrizi, A.,
595 Pinsino, A., Matranga, V., 2012. Toxicity of metal oxide nanoparticles in immune cells of the
596 sea urchin. *Marine Environmental Research* 76, 114–121.
597 <https://doi.org/10.1016/j.marenvres.2011.10.003>

598 Franco, C.F., Santos, R., Coelho, A.V., 2011. Proteome characterization of sea star coelomocytes -
599 The innate immune effector cells of echinoderms. *Proteomics* 11, 3587–3592.
600 <https://doi.org/10.1002/pmic.201000745>

601 Fries, E., Dekiff, J.H., Willmeyer, J., Nuelle, M.-T., Ebert, M., Remy, D., 2013. Identification of
602 polymer types and additives in marine microplastic particles using pyrolysis-GC/MS and
603 scanning electron microscopy. *Environmental Science: Processes & Impacts* 15, 1949.
604 <https://doi.org/10.1039/c3em00214d>

605 Fukata, Y., Oshiro, N., Kinoshita, N., Kawano, Y., Matsuoka, Y., Bennett, V., Matsuura, Y.,
606 Kaibuchi, K., 1999. Phosphorylation of Adducin by Rho-Kinase Plays a Crucial Role in Cell
607 Motility. *Journal of Cell Biology* 145, 347–361. <https://doi.org/10.1083/jcb.145.2.347>

608 Galloway, T.S., Cole, M., Lewis, C., 2017. Interactions of microplastic debris throughout the marine
609 ecosystem. *Nature Ecology & Evolution* 1, 0116. <https://doi.org/10.1038/s41559-017-0116>

610 Gambardella, C., Morgana, S., Bramini, M., Rotini, A., Manfra, L., Migliore, L., Piazza, V.,
611 Garaventa, F., Faimali, M., 2018. Ecotoxicological effects of polystyrene microbeads in a
612 battery of marine organisms belonging to different trophic levels. *Marine Environmental*
613 *Research* 141, 313–321. <https://doi.org/10.1016/j.marenvres.2018.09.023>

614 Gardon, T., Morvan, L., Huvet, A., Quillien, V., Soyeux, C., Le Moullac, G., Le Luyer, J., 2020.
615 Microplastics induce dose-specific transcriptomic disruptions in energy metabolism and
616 immunity of the pearl oyster *Pinctada margaritifera*. *Environmental Pollution* 266, 115180.
617 <https://doi.org/10.1016/j.envpol.2020.115180>

618 Golconda, P., Buckley, K.M., Reynolds, C.R., Romanello, J.P., Smith, L.C., 2019. The Axial Organ
619 and the Pharynx Are Sites of Hematopoiesis in the Sea Urchin. *Frontiers in Immunology* 10.
620 <https://doi.org/10.3389/fimmu.2019.00870>

621 Gouveia, D., Almunia, C., Cogne, Y., Pible, O., Degli-Esposti, D., Salvador, A., Cristobal, S.,
622 Sheehan, D., Chaumot, A., Geffard, O., Armengaud, J., 2019. Ecotoxicoproteomics: A decade
623 of progress in our understanding of anthropogenic impact on the environment. *Journal of*
624 *Proteomics, 10 Year Anniversary of Proteomics* 198, 66–77.
625 <https://doi.org/10.1016/j.jprot.2018.12.001>

626 Green, D.S., Colgan, T.J., Thompson, R.C., Carolan, J.C., 2019. Exposure to microplastics reduces
627 attachment strength and alters the haemolymph proteome of blue mussels (*Mytilus edulis*).

628 Environmental Pollution 246, 423–434. <https://doi.org/10.1016/j.envpol.2018.12.017>

629 Guderley, H., Pörtner, H.O., 2010. Metabolic power budgeting and adaptive strategies in zoology:
630 examples from scallops and fish. *Canadian Journal of Zool.* 88, 753–763.
631 <https://doi.org/10.1139/Z10-039>

632 Hibino, T., Loza-Coll, M., Messier, C., Majeske, A.J., Cohen, A.H., Terwilliger, D.P., Buckley, K.M.,
633 Brockton, V., Nair, S.V., Berney, K., Fugmann, S.D., Anderson, M.K., Pancer, Z., Cameron,
634 R.A., Smith, L.C., Rast, J.P., 2006. The immune gene repertoire encoded in the purple sea
635 urchin genome. *Developmental Biology* 300, 349–365.
636 <https://doi.org/10.1016/j.ydbio.2006.08.065>

637 Holland, N.D., 2020. Digestive system in regular sea urchins. In *Sea Urchins: Biology and Ecology*.
638 4th ed.; Lawrence, J.M., Ed.; Elsevier: Vol. 43; pp 147–163. [https://doi.org/10.1016/B978-0-](https://doi.org/10.1016/B978-0-12-819570-3.00008-1)
639 [12-819570-3.00008-1](https://doi.org/10.1016/B978-0-12-819570-3.00008-1)

640 Hu, M., Palić, D., 2020. Micro- and nano-plastics activation of oxidative and inflammatory adverse
641 outcome pathways. *Redox Biology* 37, 101620. <https://doi.org/10.1016/j.redox.2020.101620>

642 Inguglia, L., Chiamonte, M., Arizza, V., Turiák, L., Vékey, K., Drahos, L., Pitonzo, R., Avellone,
643 G., Di Stefano, V., 2020. Changes in the proteome of sea urchin *Paracentrotus lividus*
644 coelomocytes in response to LPS injection into the body cavity. *PLoS ONE* 15, e0228893.
645 <https://doi.org/10.1371/journal.pone.0228893>

646 Kim, H.M., Lee, D.-K., Long, N.P., Kwon, S.W., Park, J.H., 2019. Uptake of nanopolystyrene
647 particles induces distinct metabolic profiles and toxic effects in *Caenorhabditis elegans*.
648 *Environmental Pollution* 246, 578–586. <https://doi.org/10.1016/j.envpol.2018.12.043>

649 Lanning, N.J., Looyenga, B.D., Kauffman, A.L., Niemi, N.M., Sudderth, J., DeBerardinis, R.J.,
650 MacKeigan, J.P., 2014. A Mitochondrial RNAi Screen Defines Cellular Bioenergetic
651 Determinants and Identifies an Adenylate Kinase as a Key Regulator of ATP Levels. *Cell*
652 *Reports* 7, 907–917. <https://doi.org/10.1016/j.celrep.2014.03.065>

653 Li, X., Wang, R., Xun, X., Jiao, W., Zhang, M., Wang, Shuyue, Wang, Shi, Zhang, L., Huang, X.,
654 Hu, X., Bao, Z., 2015. The Rho GTPase Family Genes in Bivalvia Genomes: Sequence,
655 Evolution and Expression Analysis. *PLOS ONE* 10, e0143932.
656 <https://doi.org/10.1371/journal.pone.0143932>

657 Liang, X., Martyniuk, C.J., Simmons, D.B.D., 2020. Are we forgetting the “proteomics” in multi-
658 omics ecotoxicology? *Comparative Biochemistry and Physiology Part D: Genomics and*
659 *Proteomics* 36, 100751. <https://doi.org/10.1016/j.cbd.2020.100751>

660 Macias, D., Cózar, A., Garcia-Gorrioz, E., González-Fernández, D., Stips, A., 2019. Surface water
661 circulation develops seasonally changing patterns of floating litter accumulation in the

662 Mediterranean Sea. A modelling approach. *Marine Pollution Bulletin* 149, 110619.
663 <https://doi.org/10.1016/j.marpolbul.2019.110619>

664 Magesky, A., Ribeiro, C.A.O., Pelletier, É., 2016. Physiological effects and cellular responses of
665 metamorphic larvae and juveniles of sea urchin exposed to ionic and nanoparticulate silver.
666 *Aquatic Toxicology* 174, 208–227. <https://doi.org/10.1016/j.aquatox.2016.02.018>

667 Magni, S., Della Torre, C., Garrone, G., D'amato, A., Parenti, C., Binelli, A., 2019. First evidence of
668 protein modulation by polystyrene microplastics in a freshwater biological model.
669 *Environmental Pollution* 250, 407-415. <https://doi.org/10.1016/j.envpol.2019.04.088>

670 Manduzio, H., Rocher, B., Durand, F., Galap, C., Leboulenger, F., 2005. The point about oxidative
671 stress in molluscs. *Invertebrate Survival Journal* 2, 91–104.

672 Marques-Santos, L.F., Grassi, G., Bergami, E., Faleri, C., Balbi, T., Salis, A., Damonte, G., Canesi,
673 L., Corsi, I., 2018. Cationic polystyrene nanoparticle and the sea urchin immune system:
674 biocorona formation, cell toxicity, and multixenobiotic resistance phenotype. *Nanotoxicology*
675 12, 847–867. <https://doi.org/10.1080/17435390.2018.1482378>

676 Matranga, V., Bonaventura, R., Di Bella, G., 2002. Hsp70 as a stress marker of sea urchin
677 coelomocytes in short term cultures. *Cellular and Molecular Biology (Noisy-le-grand)* 48,
678 345–349. PMID: 12064441

679 Matranga, V., Pinsino, A., Celi, M., Natoli, A., Bonaventura, R., Schröder, H.C., Müller, W.E.G.,
680 2005. Monitoring Chemical and Physical Stress Using Sea Urchin Immune Cells, in:
681 Matranga, V. (Ed.), *Echinodermata, Progress in Molecular and Subcellular Biology*. Springer-
682 Verlag, Berlin/Heidelberg, pp. 85–110. https://doi.org/10.1007/3-540-27683-1_5

683 Matranga, V., Pinsino, A., Celi, M., Bella, G.D., Natoli, A., 2006. Impacts of UV-B radiation on
684 short-term cultures of sea urchin coelomocytes. *Marine Biology* 149, 25–34.
685 <https://doi.org/10.1007/s00227-005-0212-1>

686 Messinetti, S., Mercurio, S., Parolini, M., Sugni, M., Pennati, R., 2018. Effects of polystyrene
687 microplastics on early stages of two marine invertebrates with different feeding strategies.
688 *Environmental Pollution* 237, 1080–1087. <https://doi.org/10.1016/j.envpol.2017.11.030>

689 Mi, H., Ebert, D., Muruganujan, A., Mills, C., Albu, L.-P., Mushayamaha, T., Thomas, P.D., 2021.
690 PANTHER version 16: a revised family classification, tree-based classification tool, enhancer
691 regions and extensive API. *Nucleic Acids Research* 49, D394–D403.
692 <https://doi.org/10.1093/nar/gkaa1106>

693 Migliaccio, O., Pinsino, A., Maffioli, E., Smith, A.M., Agnisola, C., Matranga, V., Nonnis, S.,
694 Tedeschi, G., Byrne, M., Gambi, M.C., Palumbo, A., 2019. Living in future ocean
695 acidification, physiological adaptive responses of the immune system of sea urchins resident

696 at a CO₂ vent system. *Science of The Total Environment* 672, 938–950.
697 <https://doi.org/10.1016/j.scitotenv.2019.04.005>

698 Milito, A., Murano, C., Castellano, I., Romano, G., Palumbo, A., 2020. Antioxidant and immune
699 response of the sea urchin *Paracentrotus lividus* to different re-suspension patterns of highly
700 polluted marine sediments. *Marine Environmental Research* 160, 104978.
701 <https://doi.org/10.1016/j.marenvres.2020.104978>

702 Mortarino, M., Tedeschi, G., Negri, A., Ceciliani, F., Gottardi, L., Maffeo, G., Ronchi, S., 1998. Two-
703 dimensional polyacrylamide gel electrophoresis map of bull seminal plasma proteins.
704 *Electrophoresis* 19, 797–801. <https://doi.org/10.1002/elps.1150190532>

705 Murano, C., Agnisola, C., Caramiello, D., Castellano, I., Casotti, R., Corsi, I., Palumbo, A., 2020.
706 How sea urchins face microplastics: Uptake, tissue distribution and immune system response.
707 *Environmental Pollution* 264, 114685. <https://doi.org/10.1016/j.envpol.2020.114685>

708 Murano, C., Bergami, E., Liberatori, G., Palumbo, A., Corsi, I., 2021a. Interplay Between
709 Nanoplastics and the Immune System of the Mediterranean Sea Urchin *Paracentrotus lividus*.
710 *Frontiers in Marine Science* 8, 647394. <https://doi.org/10.3389/fmars.2021.647394>

711 Murano, C., Donnarumma, V., Corsi, I., Casotti, R., Palumbo, A., 2021b. Impact of Microbial
712 Colonization of Polystyrene Microbeads on the Toxicological Responses in the Sea Urchin
713 *Paracentrotus lividus*. *Environmental Science. Technology* 55, 7990–8000.
714 <https://doi.org/10.1021/acs.est.1c00618>

715 Murano, C., Vaccari, L., Casotti, R., Corsi, I., Palumbo, A., 2022. Occurrence of microfibers in wild
716 specimens of adult sea urchin *Paracentrotus lividus* (Lamarck, 1816) from a coastal area of the
717 central Mediterranean Sea. *Marine Pollution Bulletin* 176, 113448.
718 <https://doi.org/10.1016/j.marpolbul.2022.113448>

719 Nair, S.V., Del Valle, H., Gross, P.S., Terwilliger, D.P., Smith, L.C., 2005. Macroarray analysis of
720 coelomocyte gene expression in response to LPS in the sea urchin. Identification of
721 unexpected immune diversity in an invertebrate. *Physiological Genomics* 22, 33–47.
722 <https://doi.org/10.1152/physiolgenomics.00052.2005>

723 Nicastro, R., Tripodi, F., Gaggini, M., Castoldi, A., Reghellin, V., Nonnis, S., Tedeschi, G., Coccetti,
724 P., 2015. Snf1 Phosphorylates Adenylate Cyclase and Negatively Regulates Protein Kinase
725 A-dependent Transcription in *Saccharomyces cerevisiae*. *Journal of Biological Chemistry*
726 290, 24715–24726. <https://doi.org/10.1074/jbc.M115.658005>

727 Nuelle, M.-T., Dekiff, J.H., Remy, D., Fries, E., 2014. A new analytical approach for monitoring
728 microplastics in marine sediments. *Environmental Pollution* 184, 161–169.
729 <https://doi.org/10.1016/j.envpol.2013.07.027>

730 Pagliara, P., Stabili, L., 2012. Zinc effect on the sea urchin *Paracentrotus lividus* immunological
731 competence. *Chemosphere* 89, 563–568. <https://doi.org/10.1016/j.chemosphere.2012.05.052>

732 Pancer, Z., Rast, J.P., Davidson, E.H., 1999. Origins of immunity: transcription factors and
733 homologues of effector genes of the vertebrate immune system expressed in sea urchin
734 coelomocytes. *Immunogenetics* 49, 773–786. <https://doi.org/10.1007/s002510050551>

735 Paul-Pont, I., Lacroix, C., González Fernández, C., Hégaret, H., Lambert, C., Le Goïc, N., Frère, L.,
736 Cassone, A.-L., Sussarellu, R., Fabioux, C., Guyomarch, J., Albentosa, M., Huvet, A.,
737 Soudant, P., 2016. Exposure of marine mussels *Mytilus* spp. to polystyrene microplastics:
738 Toxicity and influence on fluoranthene bioaccumulation. *Environmental Pollution* 216, 724–
739 737. <https://doi.org/10.1016/j.envpol.2016.06.039>

740 Pinsino, A., Matranga, V., 2015. Sea urchin immune cells as sentinels of environmental stress.
741 *Developmental & Comparative Immunology* 49, 198–205.
742 <https://doi.org/10.1016/j.dci.2014.11.013>

743 Pinsino, A., Thorndyke, M.C., Matranga, V., 2007. Coelomocytes and post-traumatic response in the
744 common sea star *Asterias rubens*. *Cell Stress Chaperones* 12, 331–341.
745 <https://doi.org/10.1379/CSC-288.1>

746 Pinsino, A., Russo, R., Bonaventura, R., Brunelli, A., Marcomini, A., Matranga, V., 2015. Titanium
747 dioxide nanoparticles stimulate sea urchin immune cell phagocytic activity involving
748 TLR/p38 MAPK-mediated signalling pathway. *Scientific Reports* 5, 14492.
749 <https://doi.org/10.1038/srep14492>

750 Qiao, R., Sheng, C., Lu, Y., Zhang, Y., Ren, H., Lemos, B., 2019. Microplastics induce intestinal
751 inflammation, oxidative stress, and disorders of metabolome and microbiome in zebrafish.
752 *Science of The Total Environment* 662, 246–253.
753 <https://doi.org/10.1016/j.scitotenv.2019.01.245>

754 Raftos, D.A., Melwani, A.R., Haynes, P.A., Muralidharan, S., Birch, G.F., Amaral, V., Thompson,
755 E.L., Taylor, D.A., 2016. The biology of environmental stress: molecular biomarkers in
756 Sydney rock oysters (*Saccostrea glomerata*). *Environmental Science: Processes & Impacts*
757 18, 1129–1139. <https://doi.org/10.1039/c6em00322b>

758 Raguso, C., Grech, D., Becchi, A., Ubaldi, P.G., Lasagni, M., Guala, I., Saliu, F., 2022. Detection of
759 microplastics and phthalic acid esters in sea urchins from Sardinia (Western Mediterranean
760 Sea). *Marine Pollution Bulletin* 185, 114328.
761 <https://doi.org/10.1016/j.marpolbul.2022.114328>

762 Raiborg, C., Rusten, T.E., Stenmark, H., 2003. Protein sorting into multivesicular endosomes. *Current*
763 *Opinion in Cell Biology* 15, 446–455. [https://doi.org/10.1016/S0955-0674\(03\)00080-2](https://doi.org/10.1016/S0955-0674(03)00080-2)

764 Rast, J.P., Smith, L.C., Loza-Coll, M., Hibino, T., Litman, G.W., 2006. Genomic Insights into the
765 Immune System of the Sea Urchin. *Science* 314, 952–956.
766 <https://doi.org/10.1126/science.1134301>

767 Ridley, A.J., 2001. Rho GTPases and cell migration. *Journal of Cell Science* 114, 2713–2722.
768 <https://doi.org/10.1242/jcs.114.15.2713>

769 Sánchez-Marín, P., Vidal-Liñán, L., Fernández-González, L.E., Montes, R., Rodil, R., Quintana, J.B.,
770 Carrera, M., Mateos, J., Diz, A.P., Beiras, R., 2021. Proteomic analysis and biochemical
771 alterations in marine mussel gills after exposure to the organophosphate flame retardant
772 TDCPP. *Aquatic Toxicology* 230, 105688. <https://doi.org/10.1016/j.aquatox.2020.105688>

773 Schmidt, O., Teis, D., 2012. The ESCRT machinery. *Current Biology* 22, R116–120. doi:
774 10.1016/j.cub.2012.01.028.

775 Sea Urchin Genome Sequencing Consortium, et al., 2006. The genome of the Sea Urchin
776 *Strongylocentrotus purpuratus*. *Science* 314, 941. <https://doi.org/10.1126/science.1133609>

777 Sheehan, D., McDonagh, B., 2008. Oxidative stress and bivalves: a proteomic approach. *Invertebrate*
778 *Survival Journal* 5, 110–123.

779 Smith, L.C., 2010. Diversification of innate immune genes: lessons from the purple sea urchin.
780 *Disease Models & Mechanisms* 3, 274–279. <https://doi.org/10.1242/dmm.004697>

781 Smith, L.C., Davidson, E.H., 1992. The echinoid immune system and the phylogenetic occurrence of
782 immune mechanisms in deuterostomes. *Immunology Today* 13, 356–362.
783 [https://doi.org/10.1016/0167-5699\(92\)90172-4](https://doi.org/10.1016/0167-5699(92)90172-4)

784 Smith, L.C., Rast, J.P., Brockton, V., Terwilliger, D.P., Nair, S.V., Buckley, K.M., Majeske, A.J.,
785 2006. The sea urchin immune system. *Invertebrate Survival Journal* 3, 25–39.

786 Smith, L.C., Arizza, V., Barela Hudgell, M.A., Barone, G., Bodnar, A.G., Buckley, K.M., Cunsolo,
787 V., Dheilly, N.M., Franchi, N., Fugmann, S.D., Furukawa, R., Garcia-Ararras, J., Henson,
788 J.H., Hibino, T., Irons, Z.H., Li, C., Lun, C.M., Majeske, A.J., Oren, M., Pagliara, P., Pinsino,
789 A., Raftos, D.A., Rast, J.P., Samasa, B., Schillaci, D., Schrankel, C.S., Stabili, L., Stensväg,
790 K., Sutton, E., 2018. Echinodermata: The Complex Immune System in Echinoderms, in:
791 Cooper, E.L. (Ed.), *Advances in Comparative Immunology*. Springer International
792 Publishing, Cham, pp. 409–501. https://doi.org/10.1007/978-3-319-76768-0_13

793 Smith, L.C., Hawley, T.S., Henson, J.H., Majeske, A.J., Oren, M., Rosental, B., 2019. Methods for
794 collection, handling, and analysis of sea urchin coelomocytes. *Methods in Cell Biology* 150,
795 357–389. <https://doi.org/10.1016/bs.mcb.2018.11.009>

796 Sokolova, I.M., 2013. Energy-Limited Tolerance to Stress as a Conceptual Framework to Integrate
797 the Effects of Multiple Stressors. *Integrative and Comparative Biology* 53, 597–608.

798 <https://doi.org/10.1093/icb/ict028>

799 Soto-Navarro, J., Jordá, G., Compa, M., Alomar, C., Fossi, M.C., Deudero, S., 2021. Impact of the
800 marine litter pollution on the Mediterranean biodiversity: A risk assessment study with focus
801 on the marine protected areas. *Marine Pollution Bulletin* 165, 112169.
802 <https://doi.org/10.1016/j.marpolbul.2021.112169>

803 Su, C., Jiang, Y., Yang, Y., Zhang, W., Xu, Q., 2019. Responses of duckweed (*Lemna minor* L.) to
804 aluminum stress: Physiological and proteomics analyses. *Ecotoxicology and Environmental*
805 *Safety* 170, 127–140. <https://doi.org/10.1016/j.ecoenv.2018.11.113>

806 Suaria, G., Avio, C.G., Mineo, A., Lattin, G.L., Magaldi, M.G., Belmonte, G., Moore, C.J., Regoli,
807 F., Aliani, S., 2016. The Mediterranean Plastic Soup: synthetic polymers in Mediterranean
808 surface waters. *Scientific Reports* 6, 37551. <https://doi.org/10.1038/srep37551>

809 Suh, S.-S., Hwang, J., Park, M., Park, S.Y., Ryu, T.K., Lee, S., Lee, T.-K., 2014. Hypoxia-modulated
810 gene expression profiling in sea urchin (*Strongylocentrotus nudus*) immune cells.
811 *Ecotoxicology and Environmental Safety* 109, 63–69.
812 <https://doi.org/10.1016/j.ecoenv.2014.08.011>

813 Szklarczyk, D., Gable, A.L., Nastou, K.C., Lyon, D., Kirsch, R., Pyysalo, S., Doncheva, N.T.,
814 Legeay, M., Fang, T., Bork, P., Jensen, L.J., von Mering, C., 2021. The STRING database in
815 2021: customizable protein–protein networks, and functional characterization of user-
816 uploaded gene/measurement sets. *Nucleic Acids Research* 49, D605–D612.
817 <https://doi.org/10.1093/nar/gkaa1074>

818 Tang, Y., Rong, J., Guan, X., Zha, S., Shi, W., Han, Y., Du, X., Wu, F., Huang, W., Liu, G., 2020.
819 Immunotoxicity of microplastics and two persistent organic pollutants alone or in
820 combination to a bivalve species. *Environmental Pollution* 258, 113845.
821 <https://doi.org/10.1016/j.envpol.2019.113845>

822 Teng, J., Zhao, J., Zhu, X., Shan, E., Wang, Q., 2021. Oxidative stress biomarkers, physiological
823 responses and proteomic profiling in oyster (*Crassostrea gigas*) exposed to microplastics with
824 irregular-shaped PE and PET microplastic. *Science of The Total Environment* 786, 147425.
825 <https://doi.org/10.1016/j.scitotenv.2021.147425>

826 Valenzuela-Fernández, A., Cabrero, J.R., Serrador, J.M., Sánchez-Madrid, F., 2008. HDAC6: a key
827 regulator of cytoskeleton, cell migration and cell–cell interactions. *Trends in Cell Biology* 18,
828 291–297. <https://doi.org/10.1016/j.tcb.2008.04.003>

829 Van Sebille, E., Wilcox, C., Lebreton, L., Maximenko, N., Hardesty, B.D., Franeker, J.A. van,
830 Eriksen, M., Siegel, D., Galgani, F., Law, K.L., 2015. A global inventory of small floating
831 plastic debris. *Environmental Research Letters* 10, 124006. [25](https://doi.org/10.1088/1748-</p></div><div data-bbox=)

832 [9326/10/12/124006](https://doi.org/10.1016/j.marpolbul.2018.12.013)

833 Vered, G., Kaplan, A., Avisar, D., Shenkar, N., 2019. Using solitary ascidians to assess microplastic
834 and phthalate plasticizers pollution among marine biota: A case study of the Eastern
835 Mediterranean and Red Sea. *Marine Pollution Bulletin* 138, 618–625.
836 <https://doi.org/10.1016/j.marpolbul.2018.12.013>

837 Vizcaíno, J.A., Csordas, A., del-Toro, N., Dianas, J.A., Griss, J., Lavidas, I., Mayer, G., Perez-
838 Riverol, Y., Reisinger, F., Ternent, T., Xu, Q.-W., Wang, R., Hermjakob, H., 2016. 2016
839 update of the PRIDE database and its related tools. *Nucleic Acids Research* 44, D447–D456.
840 <https://doi.org/10.1093/nar/gkv1145>

841 **Figure Legends**

842 **Figure 1.** Abundance of micro-PS in different organs of sea urchins after 72h of exposure, expressed
843 as percentage of micro-PS found normalised by the weight of fresh organs.

844 **Table 1.** Ratio between red (%) and white amoebocytes (%) of sea urchin at 24,48,72h after exposure
845 to different concentrations of micro-PS (10, 50, 10³, 10⁴). Green background indicated a normal status
846 (<0.5); orange background indicated a beginning stress status (between 0.5-1); red background
847 indicated a harmful condition (>1).

Ratio R/W	24h	48h	72h
Control	0.22	0.30	0.49
10 micro-PS	0.43	0.50	0.63
50 micro-PS	0.54	0.62	0.74
10 ³ micro-PS	0.62	0.35	0.40
10 ⁴ micro-PS	0.44	1.28	0.87

848

849 **Figure 2.** Total immune cells count of sea urchin at 24, 48, 72h after exposure to different
850 concentrations of micro-PS (10, 50, 10³, 10⁴). All data were analysed by Two-way ANOVA followed
851 by Bonferroni post-test compared with the control. Bars represent mean ± SD. Asterisks indicate
852 values that are significantly different from the control, *P < 0.05; **P < 0.01.

853 **Figure 3.** Immune cells morphological profile of sea urchin at 24, 48, 72h after exposure to different
854 concentrations of micro-PS (10, 50, 10³, 10⁴). Bars represent mean ± SD.

855 **Figure 4.** Oxidative stress status of sea urchin immune cells at 24, 48, 72h after exposure to different
856 concentrations of micro-PS (10, 50, 10³, 10⁴). A) intracellular levels of ROS; B) intracellular levels
857 of RNS; C) total antioxidant capacity. All data were analysed by Two-way ANOVA followed by

858 Bonferroni post-test compared with the control. Bars represent mean \pm SD. Asterisks indicate values
859 that are significantly different from the control, *P < 0.05; **P < 0.01; ***P < 0.001.

860 **Figure 5.** Proteomic analysis of immune cells of sea urchins exposed to different concentrations of
861 micro-PS/L. The coelomocytes were collected from control animals and from sea urchins exposed to
862 different concentrations of micro-PS/L and analysed by a label-free shotgun proteomic approach. (A)
863 Venn diagram of all the data sets. The proteomic analysis allowed to identify proteins common to all
864 samples as well as proteins differentially expressed in the presence of 10 micro-PS/L, 50 micro-PS/L,
865 10^3 micro-PS/L and 10^4 micro-PS/L. (B) PCA analysis.

866 **Figure 6.** Venn diagram and bioinformatic analysis by Panther of the proteins exclusively expressed
867 in the immune cells of sea urchins exposed to different concentration of microplastics in comparison
868 to control samples (CTR). A) Venn diagram of all micro-PS/L vs CTR. (B) A total of 123 proteins
869 exclusively expressed in the coelomocytes of sea urchins exposed to different concentration of
870 microplastics in comparison to control samples were classified by Panther in terms of Functional
871 Protein Classification, GO-Biological Process classification (GOBP) and GO-Molecular Function
872 classification (GOMF).

873 **Figure 7.** Effect of the microplastics concentration on the immune cells proteome. A) Venn diagrams
874 of the comparisons: 10 micro-PS/L vs CTR, 50 micro-PS/L vs CTR, 10^3 micro-PS/L vs CTR and 10^4
875 micro-PS/L vs CTR. B) Network analysis by String of the proteins exclusively expressed in the
876 comparison 50 micro-PS/L vs CTR and involved in aerobic respiration, ATP synthesis and TCA
877 Cycle. In the figure the orthologs human gene are reported A0A7M7NI38: UQCRQ, A0A7M7N4E1:
878 NDUFS6, A0A7M7RBT7, SUCLG. C) Network analysis by String of the proteins exclusively
879 expressed in the comparison of all micro-PS/L vs CTR and involved in ESCRT complex assembly
880 (Blue) and disassembly (Red). In the figure the orthologs human gene are reported
881 (A0A7M7PCE8:SNF8, A0A7M7N202: VPS25, A0A7M7SU07:CHMP1A).

882 **Supplementary Materials**

883 **Table S1. List of the proteins identified in the CTR samples.** A database search was conducted
884 against the Uniprot *Strongylocentrotus purpuratus* database with MaxQuant software. Statistical
885 analysis was performed using the Perseus software. Proteins were considered unequivocally
886 identified if present in 75% of the replicates.

887 **Table S2. List of the proteins identified in the 10 micro-PS/L samples.** A database search was
888 conducted against the Uniprot *Strongylocentrotus purpuratus* database with MaxQuant software.
889 Statistical analysis was performed using the Perseus software. Proteins were considered
890 unequivocally identified if present in 75% of the replicates.

891 **Table S3. List of the proteins identified in the 50 micro-PS/L samples.** A database search was
892 conducted against the Uniprot *Strongylocentrotus purpuratus* database with MaxQuant software.
893 Statistical analysis was performed using the Perseus software. Proteins were considered
894 unequivocally identified if present in 75% of the replicates.

895 **Table S4. List of the proteins identified in the 10³ micro-PS/L samples.** A database search was
896 conducted against the Uniprot *Strongylocentrotus purpuratus* database with MaxQuant software.
897 Statistical analysis was performed using the Perseus software. Proteins were considered
898 unequivocally identified if present in 75% of the replicates.

899 **Table S5. List of the proteins identified in the 10⁴ micro-PS/L samples.** A database search was
900 conducted against the Uniprot *Strongylocentrotus purpuratus* database with MaxQuant software.
901 Statistical analysis was performed using the Perseus software. Proteins were considered
902 unequivocally identified if present in 75% of the replicates.

903 **Table S6. List of the proteins exclusively expressed in all the samples exposed to microplastics
904 in comparison to control.** A database search was conducted against the Uniprot *Strongylocentrotus
905 purpuratus* database with MaxQuant software. Statistical analysis was performed using the Perseus
906 software. Proteins were considered differentially expressed if they were present only in one condition
907 or showed statistically significant difference (FDR \leq 0.05, Welch's t-test). Proteins were analysed
908 using Perseus.

909 **Table S7. Enrichment analysis by Panther of the proteins exclusively expressed in the samples
910 exposed to microplastics in comparison to control.** The statistical overrepresentation analysis for
911 GO-Slim Biological Process, GO-Slim Molecular Function and Pathways was carried out by Panther
912 on the proteins exclusively expressed in the samples exposed to micro-PS/L in the following
913 comparison: all micro-PS/L vs CTR, 10 micro-PS/L vs CTR, 50 micro-PS/L vs CTR, 10³ micro-PS/L
914 vs CTR and 10⁴ micro-PS/L vs CTR. Functional grouping was based on Fisher's exact test $P \leq 0.05$.

915 **Acknowledgements**

916 We thank the OMIC facility of the University of Milano for mass spectrometry data acquisition. C.M.
917 was supported by a SZN fellowship.