

# Are “liquid plastics” a new environmental threat?

## The case of polyvinyl alcohol

Lara Nigro<sup>1</sup>, Stefano Magni<sup>1\*</sup>, Marco Aldo Ortenzi<sup>2</sup>, Stefano Gazzotti<sup>2</sup>,  
Camilla Della Torre<sup>1</sup>, Andrea Binelli<sup>1</sup>

<sup>1</sup>Department of Biosciences, University of Milan, Via Celoria 26, 20133 Milan, Italy

<sup>2</sup>Department of Chemistry, University of Milan, Via Golgi 19, 20133 Milan, Italy

\*Corresponding author: stefano.magni@unimi.it; (+39) 0250314729

### Abstract

Despite the pollution induced by plastics become a well-known and documented problem, bringing many countries to adopt restrictions about their production, commercialization and use, the impact of another emerging category of synthetic polymers, represented by the Water-Soluble Polymers (WSPs), also known as “liquid plastics”, is overlooked by scientific community. WSPs are produced in large quantities and used in a wide plethora of applications such as food packaging, pharmaceuticals and personal care products, cosmetics and detergents, with a consequent continuous release in the environment. The aim of this study was the investigation of the possible toxicity induced by polyvinyl alcohol (PVA), one of the main produced and used WSPs, on two freshwater model organisms, the crustacean *Daphnia magna* and the teleost *Danio rerio* (zebrafish). We evaluated the effects of solubilized standard PVA powder and PVA-based commercial bags for carp-fishing, at 3 different concentrations (1 µg/L, 0.5 mg/L and 1 mg/L), through the exposures for 14 days of *D. magna* (daphnids; age < 24 h) and for 5 days of zebrafish embryos (up to 120 hours post fertilization - hpf). As acute effects we evaluated the immobilization/mortality of specimens, while for chronic toxicity we selected several endpoints with a high ecological relevance, as the behavioural alteration on swimming performance, in real-time readout, and the activity of monoamine oxidase (MAO), a

27 neuro-enzyme with a potential implication in the organism movement. The results showed the lack  
28 of significant effects induced by the selected substances, at all tested concentrations and in both model  
29 organisms. However, considering the wide plethora of available WSPs, other investigations are  
30 needed to provide the initial knowledge of risk assessment of these compounds contained in some  
31 consumer products.

32  
33 Keywords: *water-soluble polymers; freshwaters, behaviour, neurotoxicity*

34  
35 1 INTRODUCTION

36 Plastics represent one of the main inventions of the 20<sup>th</sup> Century due to their low cost, mechanical  
37 properties, light weight, stability, and durability (Raddadi and Fava, 2019). In addition, these  
38 materials are suitable for a wide plethora of uses, and their production reached the 367 million tons  
39 in 2020 (PlasticsEurope, 2021). Consequently, the pollution induced by plastics represents an  
40 emerging global concern, well documented in scientific literature (Magni et al., 2019, 2021; Binelli  
41 et al., 2020, 2022; Talbot and Chang, 2022 and citations therein). However, another emerging form  
42 of plastic pollution, represented by the Water-Soluble Polymers (WSPs), also called “liquid plastics”,  
43 is overlooked by scientific community. WSPs are substances that can be water-soluble under specific  
44 conditions of pH or temperature but may become insoluble if such conditions change. Consequently,  
45 they can affect the viscosity of the aqueous solution and can be modified, through water dispersion  
46 or dissolution, in gelled, stabilized, concentrated, and emulsified formulations (Ammar et al., 2019).  
47 Being the conventional plastics characterized by solid state and insolubility in water (Hartmann et al.,  
48 2019), WSPs escape from the current legislations to limit the plastic pollution (Lam et al., 2018). In  
49 addition, WSPs are not registered under the Regulation, Evaluation, Authorization and Restriction of  
50 Chemicals (REACH) of the European Union (EU) and, consequently, there are not concrete evidence  
51 on their production volume, but also there are many gaps about their presence and effects in the

52 environment (Arp and Knutsen, 2020; Huppertsberg et al., 2020). Furthermore, this scenario is  
53 complicated by the heterogeneous variety of WSPs available for many different industrial  
54 applications and in some consumer products. Indeed, the presence of some of which is already  
55 detected in wastewaters (Antić et al., 2011; Mairinger et al., 2021) because polyacrylamide (PAM)  
56 and its co-polymers are used as flocculants in Wastewater Treatment Plants (WWTPs), while  
57 polypropylene oxide (PPO) and polyethylene glycol (PEG; also known as polyethylene oxide - PEO)  
58 are added in paints and fertilizers as dispersing agents. In addition, PEG and polyvinylpyrrolidone  
59 (PVP) are used in pharmaceuticals and personal care products (PPCPs), while polyacrylic acid (PAA)  
60 as excipient in cosmetics (Patil and Ferritto, 2013; Penlidis et al., 2018; Rivas et al., 2018;  
61 Huppertsberg et al., 2020; Rozman and Kalčíková, 2021). Another well-known WSP, the polyvinyl  
62 alcohol (PVA or PVOH), is widely used in the production of textile and industrial fibers, adhesives,  
63 binders, water-soluble films for packaging materials, and in detergent pods (DeMerlis et al., 2003;  
64 Gaaz et al., 2015). This WSP is produced by polyvinyl acetate (PVAc; FAO, 2004) and its wide use  
65 is associated to its theoretical biodegradability, chemical and thermal stability, resistance to organic  
66 solvents and high-water solubility (Julinová et al., 2018). Other specific properties, such as  
67 biocompatibility, made the PVA useful in the biomedical and pharmaceutical fields, as the production  
68 of contact lenses, synthetic tear eye-drops, surgical sponges, and drug delivery (DeMerlis et al., 2003;  
69 Muppalaneni and Omidian 2013; Gaaz et al., 2015).

70 PVA-based products represent the largest volume of WSP produced in this century, reaching 650,000  
71 tons/year worldwide (Xu et al., 2018). However, despite the massive application of PVA, previous  
72 studies reported its degradation as a slow process that can occur only under specific environmental  
73 conditions (Chiellini et al., 2003; Rolsky and Kelkar, 2021). Indeed, the physical properties of PVA  
74 as density, crystallinity and solubility are related to hydrolysis degree, crystal precipitation, molecular  
75 mass, and moisture (Saunders et al., 2012; Gaaz et al., 2015). To support this evidence, Suaria and  
76 co-authors (2016) showed that the percentage of PVA in the Mediterranean Sea was about 1.2 % of  
77 the total floating particles > 700 µm, pointing out the environmental persistence of this polymer.

78 Based on this very complicated scenario, new evidence regarding the possible ecotoxicological  
79 impact of this polymer is then required. The aim of the present study was the evaluation of the effects  
80 of 3 different concentrations (1 µg/L, 0.5 mg/L and 1 mg/L) of solubilized standard PVA powder and  
81 PVA-based commercial bags for carp-fishing, on two model organisms well representative of the  
82 aquatic ecosystem: the crustacean *Daphnia magna* (daphnids; age < 24 h, exposure of 14 days) and  
83 the teleost *Danio rerio* (zebrafish embryos; exposure from 0 to 120 hours post fertilization - hpf). We  
84 assessed the acute toxicity as immobilization/mortality, and the chronic toxicity evaluating the  
85 behavioural alteration on swimming performance, in real-time readout, as well as the activity of the  
86 neuro-enzyme monoamine oxidase (MAO), as neurotoxicity endpoint just linked to movement. The  
87 monitoring of behavioural alterations is a sensitive biomarker to evaluate the xenobiotic impact,  
88 considering that some chemicals, as pesticides or microorganism products, and nanoparticles are able  
89 to alter the locomotor-based behaviour (Bownik, 2017; Simão et al., 2019 and citations therein).  
90 Therefore, based on the ability of PVA to affect viscosity of the aqueous media, the measurement of  
91 behavioural parameters as horizontal and vertical movement and positive/negative phototaxis ratio  
92 could highlight eventual modulation of key ecological functions as predator/prey relationship and the  
93 capacity to get food (Bownik et al., 2017; Horzmann et al., 2018). The evaluation of MAO activity  
94 fits coherently with this aspect, considering the role of this enzyme family in the degradation of  
95 monoamines (e.g., catecholamines and indolamines), in turn involved in important physiological  
96 mechanisms, as movement and reproduction in both vertebrate and invertebrates (Pearson, 1993;  
97 Campos et al., 2012, 2013; McCooles et al., 2012; Bellot et al., 2021).

98

## 99 2. MATERIALS AND METHODS

### 100 2.1 PVA powder and PVA-based bag characterization

101 Standard PVA powder was purchased from Sigma Aldrich. The producer declared a molecular weight  
102 (Mw) of 89,000 - 98,000 Da, a viscosity in water (4 % at 20 °C) of 11.6 - 15.4 cps and a molar degree  
103 of hydrolysis of 99.0-99.8 %.

104 Due to the absence of technical information about the PVA-based bags (TKING; size of 100 x 140  
105 mm), a commercial product commonly used as bait container in fishing activity, we deeply  
106 investigated its chemical/physical characteristics through an integrated analytical approach. In  
107 particular, we used the Proton Nuclear Magnetic Resonance ( $^1\text{H-NMR}$ ), the Fourier-Transform  
108 Infrared Spectroscopy (FT-IR) and the Gas Chromatography-Mass Spectrometry (GC-MS) to  
109 identify the eventual presence of additives as well as the hydrolysis degree of PVA-based bags. For  
110 their characterization we used as reference standard the PVA powder Mowiol<sup>®</sup> 4-98 by Kuraray  
111 Europe GmbH, purchased from Sigma Aldrich. The producer declared a Mw of about 27,000 Da, a  
112 viscosity in water (DIN 53015) of 4.0-5.0 mPa x s, a molar degree of hydrolysis on 98.0 - 98.8% and  
113  $97.5 \pm 2.5$  % *non*-volatile components (water and organic solvents). About the  $^1\text{H-NMR}$ , the 400  
114 MHz spectra were recorded on a Bruker Ultrashield 400 spectrometer at 298 K. Samples were  
115 prepared dissolving about 10 mg of PVA in heavy water ( $\text{D}_2\text{O}$ ). The FT-IR spectra were obtained  
116 through a Spectrum 100 spectrophotometer (Perkin Elmer) in attenuated total reflection (ATR) mode  
117 using a resolution of 4.0 and 256 scans, in a range of wavenumber between 4,000 and  $400\text{ cm}^{-1}$ , using  
118 air at standard temperature and environmental moisture (23 °C and 50 % RH) as background. Lastly,  
119 the GC-MS analysis was performed using an ISQ<sup>™</sup> QD single quadrupole GC-MS (Thermo Fisher)  
120 and an Agilent technology VF-5ms (30 m x 0.25 mm i.d. x 0.25  $\mu\text{m}$ ) GC column. Parameters used in  
121 the GC oven were as follows: 60 °C held for 2 min, 50-300 °C at 10 °C/min, and 300 °C held for 5  
122 min. Carrier gas helium (He; purity  $\geq 99.999$  %) with a flow rate of 1.2 mL/min, injection temperature  
123 250 °C, injection volume of 1  $\mu\text{L}$ , and a split flow of 6.0 mL/min. MS apparatus transfer line and ion  
124 source temperatures were set at 270 °C with a delay time of 5 min. The m/z range was set between  
125 45 and 1,000. In total, 0.5 mg of the samples were dispersed in 1 mL of dichloromethane ( $\text{CH}_2\text{Cl}_2$ ;  
126 HPLC purity) and stirred at room temperature for 24 h. The solution was filtered and then analysed.

127

128 *2.2 Preparation of PVA solutions and exposure tests*

129 For the PVA powder and PVA-based bag exposures we used both zebrafish (embryos; up to 120 hpf)  
130 and *D. magna* specimens (daphnids; age < 24 h). The PVA powder and PVA-based bag solutions,  
131 used in the zebrafish exposures, were prepared in deionized water with 0.1 g/L Instant Ocean<sup>®</sup>, 0.1  
132 g/L sodium bicarbonate (NaHCO<sub>3</sub>), 0.2 g/L calcium sulphate (CaSO<sub>4</sub>), 0.1 % methylene blue and  
133 aerated for 15 min before the use. To obtain a complete solubilization of compounds, the solutions  
134 were heated up to 100 °C. On the other hand, the solutions for the *D. magna* exposures were prepared  
135 using commercial mineral water with conductivity 415 µS/cm at 25 °C, pH 7.7, 57.1 mg/L hydrogen  
136 carbonate (HCO<sub>3</sub><sup>-</sup>), 21 mg/L Calcium (Ca<sup>2+</sup>), 1.7 mg/L Magnesium (Mg<sup>2+</sup>), 1.9 mg/L Sodium (Na<sup>+</sup>),  
137 1.8 mg/L Potassium (K<sup>+</sup>), 16.9 mg/L sulphate (SO<sub>4</sub><sup>-</sup>), 1.6 mg/L nitrate (NO<sub>3</sub><sup>-</sup>), 0.2 mg/L Fluoride (F<sup>-</sup>)  
138 ), 5.9 mg/L silicon dioxide (SiO<sub>2</sub>) and aerated for 15 min.

139

#### 140 2.2.1 Zebrafish exposures

141 Regarding zebrafish, the fertilized eggs were provided by the facility of the Department of Earth and  
142 Environmental Sciences of the University of Milan Bicocca, according to the Italian laws, rules and  
143 regulations (Legislative Decree no. 116/92; authorization n. 0020984 - 12/02/2018). Considering the  
144 lack of information about the presence of WSPs in the aquatic environment, we performed a  
145 preliminary range-finding test on zebrafish embryos (and not on *D. magna*) to 1 µg/L, 0.5 mg/L, 1  
146 mg/L, 0.5 g/L and 1 g/L of PVA-based bag to select the exposure concentrations. We detected an  
147 acute effect (100 % mortality within the 120 hpf) starting to 0.5 g/L and for this reason we performed  
148 the subsequent exposures for chronic toxicity evaluation to 1 µg/L, 0.5 mg/L and 1 mg/L of PVA  
149 powder and PVA-based bag, carried out in static conditions and at 28 °C. Zebrafish embryos were  
150 exposed, in triplicate, from 0 to 120 hpf within 50 mL Petri dish with 20 organisms for each treatment  
151 group. Viability and mortality were daily reported. At the end of the exposure, we performed the  
152 analysis of behavioural alteration and subsequently the specimens were frozen at -80 °C for the  
153 measurement of MAO activity.

154

### 155 2.2.2 *D. magna* exposures

156 Regarding the *D. magna* exposures, we used specimens (daphnids, age < 24 h) derived from  
157 Daphtoxkit F ephippia (MicroBio Tests). For the ephippia hatching, 2 L of standard freshwater was  
158 prepared (UNI EN ISO 6341, 2013) with sodium bicarbonate (NaHCO<sub>3</sub>; 129.5 mg), calcium chloride  
159 dihydrate (CaCl<sub>2</sub> x 2H<sub>2</sub>O; 588 mg), magnesium sulphate heptahydrate (MgSO<sub>4</sub> x 7H<sub>2</sub>O; 264.5 mg),  
160 potassium chloride (KCl; 11.5 mg) and aerated for 15 min before the use. Subsequently, the ephippia  
161 were placed in a micro-sieve, washed using tap water to eliminate the storage medium, and transferred  
162 in a Petri dish with 15 mL of pre-aerated standard water. The ephippia incubation was carried out for  
163 72 h at 20 °C under continuous illumination at 6000 lx. Before the test the daphnids were fed for 2 h  
164 with a suspension of the blue-green alga *Spirulina* spp. The exposure was conducted for 14 days in  
165 semi-static conditions at 20.0 °C with a photoperiod of 16 h light (1500 lx) and 8 h dark, according  
166 to the *D. magna* Reproduction Test of Organisation for Economic Co-operation and Development  
167 (OECD) guideline 211 (2012). During the exposure, organisms were fed daily with the *Spirulina* spp.  
168 (3 µg/µL) and the yeast *Saccharomyces cerevisiae* (1 µg/µL), renewing the media solutions 3 times  
169 *per* week. The exposure was conducted in triplicate and for each treatment group we used 4 different  
170 beakers (50 mL) containing 5 specimens each one. Viability and immobilisation were daily reported.  
171 At the end of the exposure, we performed the analysis of behavioural alteration and subsequently the  
172 specimens were frozen at -80 °C for the measurement of MAO activity.

173 2.3

### 174 2.3 Behavioural alteration on zebrafish and *D. magna*

175 In the last decades, many ecotoxicological tests were developed to evaluate the behavioural  
176 alterations, as coiling, touch-induced escape response, optomotor and optokinetic response and light-  
177 dark challenge test (Ahmad et al., 2012 and citations therein). In this study, we conducted several  
178 experiments to evaluate both the horizontal and vertical movements and positive/negative phototaxis  
179 ratio as behavioural parameters, using the light intensity as stimulus. Potential horizontal movement  
180 alterations, induced by PVA and PVA-based bag, on the swimming activity of both zebrafish embryos

181 and *D. magna* specimens, were evaluated using the DanioVision™ video tracking system (Noldus  
182 IT, Wageningen, Netherlands). For each treatment group we used 18 specimens (54 total embryos or  
183 daphnids for each treatment). Each specimen was put in a single well of a 24-multiplate, in 3 mL of  
184 water (without tested contaminants to avoid the eventual direct effects induced by PVA on the  
185 viscosity of aqueous media) and submitted to 2 cycles of alternating dark period/low intensity light  
186 period, 2 cycles of dark period/high intensity light period and 2 cycles of dark period/highest intensity  
187 light periods (Table 1). In detail, 10 min of adaptation were followed by 2 cycles of 5 min of dark  
188 and 5 min of low intensity light at 300 lx, 2 cycles of 5 min of dark and 5 min of light at 2200 lx and  
189 2 cycles of 5 min of dark and 5 min of highest intensity light at 4400 lx (100% DanioVision™  
190 illumination), recording the swimming activity at 30 frames *per* second (Table 1). These light  
191 intensities were chosen on the basis of lx detected in an oligotrophic lake in which the genus *Daphnia*  
192 lives (300-2000 lx; Tilzer et al., 1995), as well as on previous studies aimed to evaluate the  
193 behavioural alteration on *D. magna* induced by other emerging contaminants at different light stimuli  
194 (Spulber et al., 2014; González et al., 2018; Nikitin et al., 2019; Simão et al., 2019; Bedroissant et al.,  
195 2020; Fuertes and Barata, 2020; Zheng et al., 2021). The total duration of each analysis for each  
196 multiwell was 1.10 h (10 min for each cycle). During the entire test the temperature was maintained  
197 at 28 °C for zebrafish, by the DanioVision™ temperature control unit, and at 20 °C for *D. magna*  
198 though a room temperature. The data were acquired every 30 s for 60 min and analysed using the  
199 software EthoVision XT (Noldus IT, Wageningen, Netherlands) by measuring the total distance  
200 moved (mm).

201 Considering that *D. magna* specimens in the environment migrate also vertically along the water  
202 column based on photoperiod, we evaluated their vertical migration and positive/negative phototaxis  
203 ratio. An experimental chamber was designed aligning 9 cylindrical glass cuvettes (5 x 1; h x  
204 diameter), containing an individual each one. For each group, 9 *D. magna* specimens of the same  
205 experimental treatment were distributed among the 9 vials filled with 3 mL of water (without tested  
206 contaminants to avoid the eventual direct effects induced by PVA on the viscosity of aqueous media).



207 In total, 18 specimens of the same treatment group of each experimental replicate were analysed. A  
208 visible light LED lamp (4000 K) of 25 cm, mounted on the top of the cuvettes, was used to provide  
209 the light stimuli at 300 lx, 2200 lx and 4400 lx (Table 1). The dark condition (80 lx) was obtained  
210 positioning the lamp at 2 m from the chamber. Animals were acclimated in dark conditions for 10  
211 min before the video recording. In detail, 5 min of dark period (80 lx) were followed by 15 min of  
212 increasing light intensities: 5 min at 300 lx, 5 min at 2200 lx and 5 min at 4400 lx (measured by  
213 HoldPeak 881d Digit lx meter on the top of the water column), with a total duration of 20 min for  
214 each experiment (Table 1). Video-tracking was recorded by a Basler acA1300-60gm GigE camera  
215 with an optical 8 mm HR 2.2" F1.4 lens and a resolution of 1,280 × 1,024 pixels, positioned squarely  
216 32 cm from the rack containing the experimental chambers. The GigE camera was connected through  
217 a Power PoE single injector (Ace series) to the EthoVision XT 11.5 software (Noldus IT,  
218 Wageningen, Netherlands) and the chamber was surrounded by a black paper sheet to avoid the  
219 entrance of interfering light in the system. After videorecording at 25 frames *per* second (fps),  
220 EthoVision XT 12 video tracking software was used for analysing the movement of each animal. The  
221 individual tracks, acquired every 30 s for 20 min, were analysed using the software EthoVision XT  
222 (Noldus IT, Wageningen, Netherlands) determining the total distance moved (mm) for each animal.  
223 Lastly, considering that some chemicals can alter phototaxis (Rivetti et al., 2016), we also evaluated  
224 the *D. magna* positive/negative phototaxis ratio in response to different light intensities. The analysis  
225 was performed by splitting the experimental chamber (the same used for the vertical migration  
226 assessment) in two different zones (upper, zone 1; lower, zone 2) through the EthoVision XT 11.5  
227 software. The movement of each animal in both zones was measured as the mean distance moved  
228 (mm) every 5 min (5 min of dark period, 5 min at 300 lx, 5 min at 2200 lx and 5 min at 4400 lx; Table  
229 1). Subsequently, the ratio between the distance moved in the two zones was calculated. Through this  
230 ratio, it was possible to define in which zone the *D. magna* specimens conducted the major movement  
231 under the different light conditions, defining the positive (toward light, evaluated in zone 1) or  
232 negative phototaxis (far to light, evaluated in zone 2).

233

#### 234 *2.4 Neurotoxicity biomarker*

235 Contextually to the evaluation of behavioural alteration, we also assessed the potential neurotoxicity  
236 of selected substances, on 3 pools of 20 specimens for each group, following the procedure described  
237 by Gagné (2014) and Magni et al. (2018, 2021). The homogenates were obtained pottering the  
238 specimens in 25 mM 4-(2-hydroxyethyl)-1-piperazineethanesulfonic acid) - sodium hydroxide  
239 (HEPES-NaOH) buffer (pH = 7.4), in a ratio 1:10 W/V, with 100 mM sodium chloride (NaCl), 0.1  
240 mM dithiothreitol (DTT) and protease inhibitor. Subsequently, the homogenates were centrifuged at  
241 1,000 g for 20 min at 4°C. We quantified the proteins in the S1 fraction using the Bradford method  
242 (Bradford, 1976), to normalize the MAO activity. The kinetics of MAO was measured in the S1  
243 fraction using 1 mM tyramine as substrate, 10 µM dichlorofluorescein diacetate in a 140 mM NaCl,  
244 10 mM HEPES-NaOH buffer, pH = 7.4, 1 mg/mL peroxidase and 10 mM of 3-amino-1,2,4-triazole  
245 (catalase inhibitor). We measured the fluorescence for 3 min at 485 nm (excitation) and 530 nm  
246 (emission) at the EnSight™ multimode plate reader (PerkinElmer).

247

#### 248 *2.5 Statistical analysis*

249 STATISTICA 7.0 software was used to perform Statistical analyses on biomarker and acute effect  
250 data. The significant differences between treated and control were assessed by two-way analysis of  
251 variance (two-way ANOVA) for the horizontal and vertical movement (treatment and time as  
252 variables) and by one-way ANOVA for phototaxis, MAO activity (treatment as variable) and acute  
253 effects, followed by Bonferroni Correction test ( $p < 0.05$  as significant cut-off).

254

### 255 3 RESULTS AND DISCUSSION

#### 256 *3.1 Material characterization*

257 The  $^1\text{H}$  NMR spectra were recorded to determine the actual hydrolysis degree of Mowiol<sup>®</sup> 4-98 in the  
258 form of powder and the PVA-based bag. The calculation was made according to previous approaches  
259 (Budhlall et al. 2000): spectra of Mowiol<sup>®</sup> and PVA-based bag (Figure 1A and B, respectively) are  
260 characterized by a peak centred at 1.96 ppm, which can be attributed to the methyl group (Figure 1A  
261 and B, in red) of the acetyl group of the *non*-hydrolysed repeating units. Given the higher hydrolysis  
262 degree, the peak is only slightly detectable in the spectrum of the Mowiol<sup>®</sup> (Figure 1A), while it is  
263 well visible in the PVA-based bag (Figure 1B). The integration of the peak, compared to the  
264 integration of the signals of the -CH and -CH<sub>2</sub> groups of the chain, allowed us the calculation of the  
265 hydrolysis degree of the samples assessed to be 98 % for the Mowiol<sup>®</sup> and 85 % for the PVA-based  
266 bag. In this context, superimposed FT-IR spectra for PVA-based bag and Mowiol<sup>®</sup> are reported in  
267 Figure 1C. The main difference between the two spectra is located in the strong band at 1,736 cm<sup>-1</sup>  
268 which was attributed to the carbonyl stretching of the acetyl moieties on the partially hydrolysed  
269 PVA-based bag. The analysis did not allow the detection of additives which are likely embedded in  
270 the material itself and therefore not detectable as such through FT-IR. However, as shown in Figure  
271 1A and B, the spectrum of the PVA-based bag highlighted the presence of signal that cannot be  
272 attributed to Mowiol<sup>®</sup> and is therefore likely due to additives (peaks at 4.57 and 3.43 ppm) which are  
273 not present in the Mowiol<sup>®</sup>. To characterize the additives, both the Mowiol<sup>®</sup> and the PVA-based bag  
274 were subjected to solvent extraction with dichloromethane and GC-MS chromatograms were  
275 recorded on the extracts. As shown in Figure 2A, no peaks are detectable in the Mowiol<sup>®</sup> extract,  
276 while three main peaks are detectable in the chromatogram of the PVA-based bag (Figure 2B) extract  
277 at 10.22, 13.78, and 17.00 min, which were assigned by NIST2017 database to triethylene glycol  
278 (TEG), tetraethylene glycol (TetraEG) and pentaethylene glycol, respectively. The presence of these  
279 three species also finds confirmation in the peaks at 4.57 and 3.43 ppm detected in the  $^1\text{H}$ NMR  
280 spectrum reported in Figure 1B. These additives, derived from ethylene glycol (EG), were probably  
281 used as plasticizer in the PVA-based bag production, considering the EG use in the polyester industry,  
282 *non*-volatile antifreeze, and plasticizer (Guo et al., 2007; Yin et al., 2019). In addition, TEG/PVA

283 blend, due to its high hygroscopic property, is used in dehumidification applications (Bui et al., 2017).  
284 For this reason, the addition of TEG in the PVA product could reduce the moisture absorption by  
285 PVA, increasing both product conservation and quality. Based on these evidence, the identification  
286 of TEG, TetraEG and pentaethylene glycol in the PVA-based bag could be interesting in the context  
287 of ecotoxicological effects of considered materials.

288

### 289 3.2 Toxicity evaluation

290 Regarding the acute effects induced by PVA powder and PVA-based bag on both *D. magna* and  
291 zebrafish specimens, coherently with the results obtained in the range-finding test, we did not observe  
292 significant differences in the viability parameter. In detail, in *D. magna* specimens all treatment  
293 groups showed a viability  $\geq$  to 77 % (78 % for 1  $\mu\text{g/L}$ , 77 % for 0.5 mg/L and 82 % for 1 mg/L PVA  
294 powder and 85 % for 1  $\mu\text{g/L}$  and 83 % for both 0.5 and 1 mg/L of PVA-based bag), perfectly  
295 comparable with a value of 88 % in the control. In the same manner, in zebrafish we obtained a  
296 viability  $\geq$  to 86 % in all treatments (92 % for 1  $\mu\text{g/L}$ , 93 % for 0.5 mg/L and 95 % for 1 mg/L PVA  
297 powder and 92 % for 1  $\mu\text{g/L}$ , 86 % for 0.5 mg/L and 92 % for 1 mg/L of PVA-based bag), with 98 %  
298 of viability in the control.

299 After the confirmation of the absence of acute effects, we evaluated the potential chronic toxicity on  
300 the selected biological models. Concerning the behavioural alteration, after 14 days of exposure to  
301 PVA powder and PVA-based bag, the *D. magna* horizontal swimming showed a similar behavioural  
302 response among the treatments without significant differences compared to control (Figures 3A and  
303 B). Only the *D. magna* specimens exposed to 0.5 mg/L of PVA powder showed a higher, but not  
304 significant, swimming performance during the entire light/dark 60 min cycle (Figure 3A). The  
305 increase of distance moved under the different lights could be related to the phototaxis, to find darker  
306 zones to make themselves less visible to visual-oriented predators, whose activity increases at high  
307 light intensities (Tałanda et al., 2018; Simão et al., 2019). However, no differences, compared to

308 control, are induced by the two tested materials on positive/negative phototaxis ratio (Figure 4). Since  
309 phototaxis is also related to *D. magna* vertical migration to limit the predation during the day, we also  
310 checked this parameter, but no differences were observed between treated and controls (Figures 5A  
311 and B).

312 Moving to the results of the tests on zebrafish, which represents the potential zooplankton predator,  
313 the behavioural analysis did not show again any differences in the horizontal swimming, compared  
314 to control (Figure 3C and D). In general, differently to *D. magna*, the total distance moved was higher  
315 under dark compared to light conditions. In line with previous studies (Llanos et al., 2018; Basnet et  
316 al., 2019; Hussain et al., 2020), the reduction of locomotion in response to light stimuli test was  
317 observed under all light intensities. This phenomenon is known as “freezing”, a common anxiety  
318 index (Champagne et al., 2010) together with erratic movements, thigmotaxis and scototaxis,  
319 characterized by the absence of movement, apart from gills and eyes (Ahmad et al., 2012). In general,  
320 adult specimens present freezing behaviour if exposed to anxiogenics, as illumination conditions and  
321 environmental characteristics (inner/outer and opaque/transparent zones of the tank; Egan et al., 2009;  
322 Champagne et al., 2010), while young larvae remain in freezing condition for more time (Thirumalai  
323 and Cline, 2008; Colwill and Creton, 2011). In this context, the light conditions are a pivotal  
324 parameter to zebrafish embryos, which can decrease or increase their swimming activity depending  
325 on light intensity (Padilla et al., 2011).

326 Moving to the results obtained after our exposure assays, the freezing behaviour was more visible in  
327 the control groups than in other exposure groups, although no significant differences were observed  
328 (Figures 3 C and D). Interestingly, zebrafish specimens did not show the natural freezing behaviour  
329 in the treated to 1 µg/L and 0.5 mg/L PVA powder when exposed for the first time to 4400 lx, while  
330 the groups exposed to 1 mg/L PVA powder and the control showed a rapid fall of movement  
331 performance when exposed for the first time at the same light condition (Figure 3C). This strange  
332 behavior, the cause of which should be investigated further in the future, is also confirmed in the  
333 second exposure at 4400 lx, albeit to a lesser extent than the first.

334 The lack of significant alteration was obtained, once again, in zebrafish exposed to PVA-based bag  
335 in all the treatment groups, although a *non*-significant decrease in the distance moved can be observed  
336 for all three treatments for the two initial dark phases and for the exposure to 300 lx (Figure 3D).  
337 Furthermore, a general decrease of the movement under 2200 lx and 4400 lx conditions was detected  
338 for all treatment groups, while the specimens exposed to 0.5 mg/L PVA-based bag showed a lower  
339 movement during all the analysis, compared to control (Figure 3C and D). To deeply investigate the  
340 effects of selected materials on the movement of exposed organisms, we also evaluated the MAO  
341 activity. After the exposure, *non*-significant increase in the biological trend of MAO activity was  
342 observed in both *D. magna* and zebrafish embryos (Figure 6). In detail, *D. magna* showed an increase  
343 of MAO activity at 1 µg/L and 1 mg/L of PVA powder (Figure 6A), while zebrafish at 1 µg/L of  
344 PVA powder (Figure 6C) and at all tested concentrations of PVA-based bag (Figure 6D).  
345 As final consideration, the lack of significant differences for the MAO activities of the exposed  
346 organisms with the relative controls confirms that the two tested materials, and therefore the PVA  
347 which represents their major component, does not seem to have any negative effect on the two  
348 selected biological models, at least for the tested concentrations and for the measured end points.  
349 Indeed, we must remember that the preliminary range-finding had highlighted an extensive mortality  
350 in zebrafish embryos exposed to higher PVA concentrations. The lack of effects observed in this  
351 study suggests the absence of ecotoxicological differences between the standard PVA powder and  
352 PVA-based bag, highlighting how the additives detected in the commercial product did not have a  
353 key role in the toxicity. Lastly, we would highlight the need to improve the knowledge regarding the  
354 impact of these emerging contaminants on the aquatic environment, making possible the eventual  
355 toxicity comparison between experiments on WSP toxicity, actually not feasible due to the absence  
356 of data.

357

358 4 CONCLUSIONS

359 This study is surely the first to investigate the potential ecotoxicological effect of one of the most  
360 widely used WSPs using two different biological models. The presented results showed a lack of  
361 toxicity induced by standard PVA powder and a commercial PVA-based bag on selected freshwater  
362 species investigating both apical parameters, such as any effects on behaviour, and biochemicals such  
363 as the measurement of MAO activity. However, other investigations are necessary in this field for the  
364 following reasons: 1) could be interesting to evaluate also the impact of these contaminants at  
365 molecular and cellular levels, using biomarkers of cellular stress, oxidative damage and cyto-  
366 genotoxicity and “omics” techniques, to propose a possible WSP mechanism of action; 2) the  
367 presented results are referred to PVA, but considering the wide plethora of produced and used WSPs,  
368 as PEG, PVP and PAA, the future studies could also consider the impact of these substances; 3) a  
369 pivotal aspect in the ecotoxicology of WSP could be their monitoring in the aquatic environment, to  
370 certify their presence, as well as to provide environmental relevant concentrations useful for the  
371 laboratory exposures.

372

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375 Stefano Magni by Department of Biosciences of the University of Milan.

376

## 377 6 AUTHOR STATEMENT

378 **Binelli Andrea:** Conceptualization; Writing - Review & Editing; Supervision - **Della Torre Camilla:**  
379 Writing - Review & Editing - **Gazzotti Stefano:** Methodology; Formal analysis - **Magni Stefano:**  
380 Conceptualization; Data Curation; Writing - Original Draft; Funding acquisition; Project  
381 administration - **Nigro Lara:** Methodology; Formal analysis; Validation; Data Curation; Writing -  
382 Original Draft - **Ortenzi Marco Aldo:** Methodology; Formal analysis

383

384 7 REFERENCES

- 385 Ahmad, F., Noldus, L.P.J.J., Tegelenbosch, R.A.J., Richardson, M.K., 2012. Zebrafish embryos and  
386 larvae in behavioural assays. *Behaviour*, 149, 1241-1281.
- 387 Ammar, S., Ma, I.A.W, Ramesh, K., Ramesh, S., 2019. Chapter 2 - Polymers-based nanocomposite  
388 coatings. *Nanomaterials-Based Coatings*, 9-39.
- 389 Antić, V.V., Antić, M.P., Kronimus, A., Oing, K., Schwarzbauer, J., 2011. Quantitative determination  
390 of poly(vinylpyrrolidone) by continuous-flow off-line pyrolysis-GC/MS. *J. Anal. Appl.*  
391 *Pyrolysis*, 90, 93-99.
- 392 Arp, H.P.H., Knutsen, H., 2020. Could We Spare a Moment of the Spotlight for Persistent, Water-  
393 Soluble Polymers? *Environ. Sci. Technol.*, 7, 3-5.
- 394 Basnet, R.M., Zizoli, D., Taweedet, S., Finazzi, D., Memo, M., 2019. Zebrafish Larvae as a  
395 Behavioral Model in Neuropharmacology. *Biomed.*, 7, 23.
- 396 Bedrossiantz, J., Jerònimo, M.F., Bellot, M., Raldua, D., Canela, G.C., Barata, C., 2020. A high-  
397 throughput assay for screening environmental pollutants and drugs impairing predator  
398 avoidance in *Daphnia magna*. *Sci. Total Environ.*, 740, 140045.
- 399 Bellot, M., Faria, M., Gómez-Canela, C., Raldúa, D., Barata, C., 2021. Pharmacological Modulation  
400 of Behaviour, Serotonin and Dopamine Levels in *Daphnia magna* Exposed to the Monoamine  
401 Oxidase Inhibitor Deprenyl. *Toxics*, 9, 187.
- 402 Binelli, A., Della Torre, C., Nigro, L., Riccardi, N., Magni, S., 2022. A realistic approach for the  
403 assessment of plastic contamination and its ecotoxicological consequences: A case study in  
404 the metropolitan city of Milan (N. Italy). *Sci. Total Environ.*, 806, 150574.
- 405 Binelli, A., Pietrelli, L., Di Vito, S., Coscia, L., Sighicelli, M., Della Torre, C., Parenti, C.C., Magni,  
406 S., 2020. Hazard evaluation of plastic mixtures from four Italian subalpine great lakes on the  
407 basis of laboratory exposures of zebra mussels. *Sci. Total Environ.*, 699, 134366.
- 408 Bownik, A., 2017. *Daphnia* swimming behaviour as a biomarker in toxicity assessment: A review.  
409 *Sci. Total Environ.*, 601-602, 194-205.
- 410 Bradford, M.M., 1976. A rapid and sensitive method for the quantitation of microgram quantities of  
411 protein utilizing the principle of protein-dye binding. *Anal. Biochem.*, 72, 248-254.



- 412 Budhlall, B.; Landfester, K.; Nagy, D.; Sudol, E. D.; Dimonie, V.; Sagl, D.; El-Aasser, M. S., 2000.  
413 Characterization of partially hydrolyzed poly(vinyl alcohol) I : Sequence distribution of  
414 poly(vinyl alcohol) via <sup>13</sup>C and <sup>1</sup>H-NMR and a reversed-phased gradient elution HPLC  
415 technique. *Macromol. Symp.*, 155, 63-84.
- 416 Bui, T.D., Wong, Y., Thu, K., Oh, S.J., Ja, M.K., Ng, K.C., Raisul, I., Chua, K.J., 2017. Effect of  
417 hygroscopic materials on water vapor permeation and dehumidification performance of  
418 polyvinyl alcohol membranes. *J. Appl. Polym. Sci.*, 134.
- 419 Campos, B., Garcia-Reyero, N., Rivetti, C., Escalon, L., Habib, T., Tauler, R., Tsakovski, S., Pina,  
420 B., Barata, C., 2013. Identification of metabolic pathways in *Daphnia magna* explaining  
421 hormetic effects of selective serotonin reuptake inhibitors and 4-nonylphenol using  
422 transcriptomic and phenotypic responses. *Environ. Sci. Technol.*, 47, 9434-9443.
- 423 Campos, B., Pina, B., Barata, C., 2012. Mechanisms of action of selective serotonin reuptake  
424 inhibitors in *Daphnia magna*. *Environ. Sci. Technol.*, 46, 2943–2950.
- 425 Champagne, D.L., Hoefnagels, C.C., de Kloet, R.E., Richardson, M.K., 2010. Translating rodent  
426 behavioral repertoire to zebrafish (*Danio rerio*): relevance for stress research. *Behav. Brain*  
427 *Res.*, 214, 332-342.
- 428 Chiellini, E., Corti, A., D’Antone, S., Solaro, R., 2003. Biodegradation of poly (vinyl alcohol) based  
429 materials. *Prog. Polym. Sci.*, 28, 963-1014.
- 430 Colwill, R.M., Creton, R., 2011. Locomotor behaviors in zebrafish (*Danio rerio*) larvae. *Behav. Proc.*,  
431 86, 222-229.
- 432 DeMerlis, C.C., Schoneker, D.R., 2003. Review of the oral toxicity of polyvinyl alcohol (PVA).  
433 *Food Chem. Toxicol.*, 41, 319-326.
- 434 Egan, R.J., Bergner, C.L., Hart, P.C., Cachat, J.M., Canavello, P.R., Elegante, M.F., Elkhayat, S.I.,  
435 Bartels, B.K., Tien, A.K., Tien, D.H., Mohnot, S., Beeson, E., Glasgow, E., Amri, H.,  
436 Zukowska, Z. & Kalueff, A.V., 2009. Understanding behavioral and physiological phenotypes  
437 of stress and anxiety in zebrafish. *Behav. Brain Res.*, 205, 38-44.
- 438 FAO (Food and Agriculture Organization), 2004. Chemical and Technical Assessment (CTA), United  
439 Nations. Polyvinyl Alcohol (PVA), First Draft Prepared by S.K. Saxena.

- 440 Fuertes, I., Baraata, C., 2020. Characterization of neurotransmitters and related metabolites in  
441 *Daphnia magna* juveniles deficient in serotonin and exposed to neuroactive chemicals that  
442 affect its behavior: A targeted LC-MS/MS method. *Chemosphere*, 127814.
- 443 Gaaz, T.S., Sulong, A.B., Akhtar, M.N., Kadhum, A.A., Mohamad, A.B., Al-Amiery, A.A., 2015.  
444 Properties and applications of polyvinyl alcohol, halloysite nanotubes and their  
445 nanocomposites. *Molecules*, 20, 22833-22847.
- 446 Gagné, F., 2014. *Biochemical Ecotoxicology*. Elsevier.
- 447 González, E.A., Carty, D.R., Tran, F.D., Cole, A.M., Lein, P.J., 2018. Developmental exposure to  
448 silver nanoparticles at environmentally relevant concentrations alters swimming behavior in  
449 zebrafish (*Danio rerio*). *Environ. Toxicol. Chem.*, 37, 3018-3024.
- 450 Guo, R., Ma, X., Hu, C., Jiang, Z., 2007. Novel PVA–silica nanocomposite membrane for  
451 pervaporative dehydration of ethylene glycol aqueous solution. *Polymer*, 48, 2939-2945.
- 452 Hartmann, N.B., Hüffer, T., Thompson, R.C., Hassellöv, M., Verschoor, A., Daugaard, A.E., Rist,  
453 S., Karlsson, T., Brennholt, N., Cole, M., Herrling, M.P., Hess, M.C., Ivleva, N.P., Lusher,  
454 A.L., Wagner, M., 2019. Are We Speaking the Same Language? Recommendations for a  
455 Definition and Categorization Framework for Plastic Debris. *Environ. Sci. Technol.*, 53,  
456 1039-1047.
- 457 Horzmann, K.A., Freeman, J.L., 2018. Making waves: New developments in Toxicology with the  
458 Zebrafish. *Toxicol. Sci.*, 163, 5-12.
- 459 Huppertsberg, S., Zahn, D., Pauelsen, F., Reemtsma, T., Knepper, T.P., 2020. Making waves: Water-  
460 soluble polymers in the aquatic environment: an overlooked class of synthetic polymers?  
461 *Water Res.*, 181, 115931.
- 462 Hussain, A., Audira, G., Malhotra, N., Uapipatanakul, B., Chen, J. R., Lai, Y. H., Huang, J.C., Chen,  
463 K.H.C., Lai, H.T., Hsiao, C.D., 2020. Multiple screening of pesticides toxicity in zebrafish  
464 and daphnia based on locomotor activity alterations. *Biomolecules*, 10, 1224.
- 465 Julinová, M., Vaňharová, L., Jurča, M., 2018. Water-soluble polymeric xenobiotics – Polyvinyl  
466 alcohol and polyvinylpyrrolidon – And potential solutions to environmental issues: A brief  
467 review. *J. Environ. Manag.*, 228, 213-222.
- 468 Lam, C.S., Ramanathan, S., Carbery, M., Gray, K., Vanka, K.S., Maurin, C., Bush, R., Palanisami,  
469 T., 2018. *A Comprehensive Analysis of Plastics and Microplastic Legislation Worldwide*.

470 Water Air Soil Pollut., 229, 345. Llanos, P.J., Andrijauskaite, K., Rubinstein, M.P., Chan,  
471 S.S.L., 2018. Investigation of Zebrafish Larvae Behavior as Precursor for Suborbital Flights:  
472 Feasibility Study. Gravit. Space Res., 6.

473 Magni, S., Binelli, A., Pittura, L., Avio, C.G., Della Torre, C., Parenti, C.C., Gorbi, S., Regoli, F.,  
474 2019. The fate of microplastics in an Italian Wastewater Treatment Plant. Sci. Total Environ.,  
475 652, 602-610.

476 Magni, S., Gagné, F., André, C., Della Torre, C., Auclair, J., Hanana, Houda, Parenti, C.C., Bonasoro,  
477 F., Binelli, A., 2018. Evaluation of uptake and chronic toxicity of virgin polystyrene  
478 microbeads in freshwater zebra mussel *Dreissena polymorpha* (Mollusca: Bivalvia). Sci.  
479 Total Environ., 631-632, 778-788.

480 Magni, S., Nigro, L., Della Torre, C., Binelli, A., 2021. Characterization of plastics and their  
481 ecotoxicological effects in the Lambro River (N. Italy). J. Hazard. Mater., 412, 125204.

482 Mairinger, T., Loos, M., Hollender, J., 2021. Characterization of water-soluble synthetic polymeric  
483 substances in wastewater using LC-HRMS/MS. Water Res., 190, 116745.

484 McCoole, M.D., Atkinson, N.J., Graham, D.I., Grasser, E.B., Joselow, A.L., McCall, N.M., Welker,  
485 A.M., Wilsterman J, E.J., Baer, K.N., Tilden, A.R., Christie, A.E., 2012. Genomic analyses  
486 of aminergic signaling systems (dopamine, octopamine and serotonin) in *Daphnia pulex*.  
487 Comp. Biochem. Physiol. – Part D: Genomics Proteomics, 7, 35-58.

488 Muppalaneni, S., Omidian, H., 2013. Polyvinyl alcohol in medicine and pharmacy: a perspective. J.  
489 Dev. Drugs, 2, 1000112.

490 Nikitin, O., Nasyrova, E., Kalinina, A., Latypova, V., 2019. Effect of various temperature and light  
491 intensity regimes on *Daphnia magna* swimming behaviour. Int. Multidiscip. Sci.  
492 GeoConference- SGEM.

493 OECD, 2012. Test No. 211: *Daphnia magna* Reproduction Test.

494 Padilla, S., Hunter, D.L., Padnos, B., Frady, S., MacPhail, R.C., 2011. Assessing locomotor activity  
495 in larval zebrafish: influence of extrinsic and intrinsic variables. Neurotoxicol. Teratol., 33,  
496 624-630.

497 Patil, A., Ferritto, M.S., 2013. Polymers for Personal Care and Cosmetics. Am. Chem. Soc., 1148.

498 Pearson, K.G., 1993. Common principles of motor control in vertebrates and invertebrates. Annu.  
499 Rev. Neurosci., 16, 265-297.

500 Penlidis, A., 2018. Water Soluble Polymers. Processes, MDPI.

501 Plastics Europe, 2021, Plastics- the Facts 2021.

502 Radaddi, N., Fava, F., 2019. Biodegradation of oil-based plastics in the environment: Existing  
503 knowledge and needs of research and innovation. Sci. Total Environ., 679, 148-158.

504 Rivas, B.L., Urbano, B.F., Sánchez, J., 2018. Water-Soluble and Insoluble Polymers, Nanoparticles,  
505 Nanocomposites and Hybrids With Ability to Remove Hazardous Inorganic Pollutants in  
506 Water. Front. Chem., 6, 320.

507 Rivetti, C., Campos, B., Carata, C., 2016. Low environmental levels of neuro-active pharmaceuticals  
508 alter phototactic behaviour and reproduction in *Daphnia magna*. Aquat. Toxicol., 170, 289-  
509 296.

510 Rolsky, C., Kelkar, V., 2021. Degradation of Polyvinyl Alcohol in US Wastewater Treatment Plants  
511 and Subsequent Nationwide Emission Estimate. Int. J. Environ. Res. Public Health, 18, 6027.

512 Rozman, U., Kalčíkova, G., 2021. Seeking for a perfect (non-spherical) microplastic particle – the  
513 most comprehensive review on microplastic laboratory research. J. Hazard. Mater.,  
514 424, 127529.

515 Saunders, K.J., 2012. Organic Polymer Chemistry: An Introduction to the Organic Chemistry of  
516 Adhesives, Fibres, Paints, Plastics and Rubbers. Springer Science & Business Media.

517 Simão, C.P.F., Jerónimo, F.M., Blasco, V., Moreno, F., Porta, J.M., Pestana, J.L.T., Soares,  
518 A.M.V.M., Raldúa, D., Barata, C., 2019. Using a new high-throughput video-tracking  
519 platform to assess behavioural changes in *Daphnia magna* exposed to neuro-active drugs. Sci.  
520 Total Environ., 662, 160-167.

521 Spulber, S., Kilian, P., Ibrahim, W.N.W., Onishchenko, N., Ulhaq, M., Norrgen, L., Negri, S., Di  
522 Tuccio, M., Ceccatelli, S., 2014. PFOS Induces behavioural alterations, including spontaneous  
523 hyperactivity that is corrected by dexamfetamine in Zebrafish larvae. Plos ONE, 9, e94227.

524 Suaria, G., Avio, C.G, Mineo, A., Lattin, G.L., Magaldi, M.G., Belmonte, G., Moore, C.J., Regoli,  
525 F., Aliani, S., 2016. The Mediterranean Plastic Soup: synthetic polymers in Mediterranean  
526 surface waters. Sci. Rep., 6, 37551.

527 Tałanda, J., Maszczyk, P., Babkiewicz, E., 2018. The reaction distance of a planktivorous fish  
528 (*Scardinius erythrophthalmus*) and the evasiveness of its prey (*Daphnia pulex* × *pulicaria*)  
529 under different artificial light spectra. Limnology 19, 311-319.

- 530 Talbot, R., Chang, H., 2022. Microplastics in freshwater: A global review of factors affecting spatial  
531 and temporal variations. *Environ. Pollut.*, 292, 118393.
- 532 Thirumalai, V., Cline, H.T., 2008. Endogenous dopamine suppresses initiation of swimming in  
533 prefeeding zebrafish larvae. *J. Neurophysiol.*, 100, 1635-1648.
- 534 Tilzer, M.M., Stambler, N., Lovengreen, C., 1995. The role of phytoplankton in determining the  
535 underwater light climate in Lake Constance. *Hydrobiologia*, 316, 161-172.
- 536 UNI EN ISO 6341:2013
- 537 Xu, S., Akbar, Malik, A.M., Qi, Z., Huang, B.T., Li, Q., Sarkar, M., 2018. Influence of the PVA  
538 fibers and SiO<sub>2</sub> NPs on the structural properties of fly ash based sustainable geopolymer.  
539 *Constr. Build Mater.*, 9-39.
- 540 Yin, D., Xiang, A., Li, Y., Qi, H., Tian, H., Fan, G., 2019. Effect of plasticizer on the morphology  
541 and foaming properties of poly(vinyl alcohol) foams by supercritical CO<sub>2</sub> foaming agents. *J.*  
542 *Polym. Environ.*, 27, 2878-2885.
- 543 Zheng, S., Huang, W., Liu, C., Xiao, J., Wu, R., Wang, X., Cai, Z., Wu, K., 2021. Behavioral change  
544 and transcriptomics reveal the effects of 2, 2', 4, 4'-tetrabromodiphenyl ether exposure on  
545 neurodevelopmental toxicity to zebrafish (*Danio rerio*) in early life stage. *Sci. total Environ.*,  
546 752

547

548 Captions

549

550 Figure 1: (A) <sup>1</sup>H-NMR spectrum of Mowiol<sup>®</sup> and (B) of PVA-based bag. The peaks centred at 1.96  
551 (methyl group of the acetyl group of the *non*-hydrolysed repeating units), 4.57 and 3.43 ppm  
552 (additives in the PVA-based bag) were highlighted by arrows in the spectra. (C) Superimposed FT-  
553 IR spectra for Mowiol<sup>®</sup> (blue infrared spectrum) and PVA-based bag (black infrared spectrum). The  
554 strong band at 1,736 cm<sup>-1</sup> was attributed to the carbonyl stretching of the acetyl moieties on the  
555 partially hydrolysed PVA-based bag.

556 Figure 2: GC-MS chromatograms relative to the extracts of Mowiol<sup>®</sup> (A) and PVA-based bag (B).  
557 The peaks in the chromatogram of PVA-based bag extract were referred to additives triethylen glycol  
558 (TEG; 10.22 min), tetraethylen glycol (TetraEG; 13.78 min) and pentaethylen glycol (17.00 min).

559 Figure 3: Horizontal movement (mean value; the standard deviations (SDs) were removed from the  
560 graphs to increase the readability of presented results, see the Supplementary materials for SD) of *D.*  
561 *magna* (A, B) and zebrafish embryos (C, D) at the end of exposure (14 days and 120 hpf, respectively)  
562 to PVA and PVA-based bag (exposure in triplicate; n = 18 specimens *per* treatment; two-way  
563 ANOVA). The measurements (every 30 s) were conducted as the distance moved across consecutive  
564 2 cycles of 5 min of dark and 5 min of low intensity light at 300 lx, 2 cycles of 5 min of dark and 5  
565 min of light at 2200 lx and 2 cycles of 5 min of dark and 5 min of highest intensity light at 4400 lx.  
566 We used the same control groups for both PVA powder and PVA-based bag treatments.

567 Figure 4: Positive/negative phototaxis ratio (mean  $\pm$  standard deviation; SD) of *D. magna* at the end  
568 of exposure (14 days) to PVA powder (A,B,C and D, the letters are referred to different light  
569 intensities) and PVA-based bag (E,F,G and H; exposure in triplicate; n = 18 specimens *per* treatment;  
570 one-way ANOVA). The measurements were conducted as the distance moved across consecutive 5  
571 min of dark (80 lx), low intensity light at 300 lx, light at 2200 lx and highest intensity light at 4400  
572 lx.

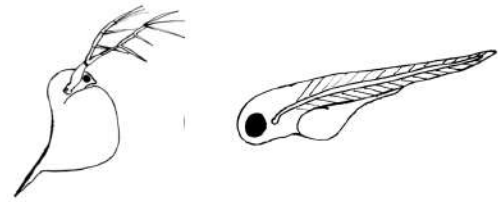
573 Figure 5: Vertical migration (mean value; the standard deviations (SDs) were removed from the  
574 graphs to increase the readability of presented results, see the Supplementary materials for SD) of *D.*  
575 *magna* at the end of exposure (14 days) to PVA powder (A) and PVA-based bag (B; exposure in  
576 triplicate; n = 18 specimens *per* treatment; two-way ANOVA). The measurements (every 30 s) were  
577 conducted as the distance moved across consecutive 5 min of dark (80 lx), low intensity light at 300  
578 lx, light at 2200 lx and highest intensity light at 4400 lx. We used the same control groups for both  
579 PVA powder and PVA-based bag treatments.

580 Figure 6: MAO activity (mean  $\pm$  standard deviation; SD) in *D. magna* (A, B) and zebrafish embryos  
581 (C, D) at the end of exposure (14 days and 120 hpf, respectively) to PVA and PVA-based bag (for *D.*  
582 *magna*: exposure in triplicate with 4 beakers for each treatment - 5 specimens *per* beakers, n = 3 pools  
583 of 20 specimens *per* treatment; for zebrafish: exposure in triplicate with 1 Petri dish for each treatment  
584 with - 20 specimens *per* Petri dish; n = 3 pools of 20 specimens *per* treatment; one-way ANOVA).  
585 We used the same control groups for both PVA powder and PVA-based bag treatments.

- Water-soluble polymers represent an overlooked global issue
- The effects of polyvinyl alcohol were investigated on freshwater organisms
- Chronic toxicity was evaluated through behavioural and neurotoxicity biomarkers
- Polyvinyl alcohol did not induce significant effects on exposed organisms



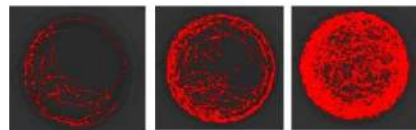
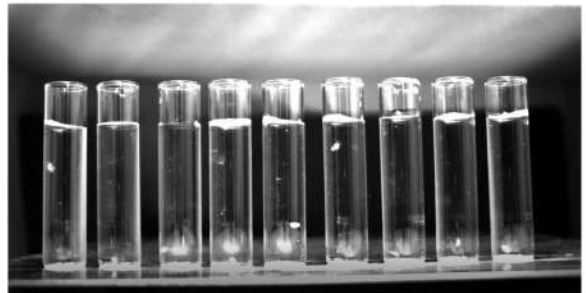
PVA powder



*D. magna* and zebrafish specimens



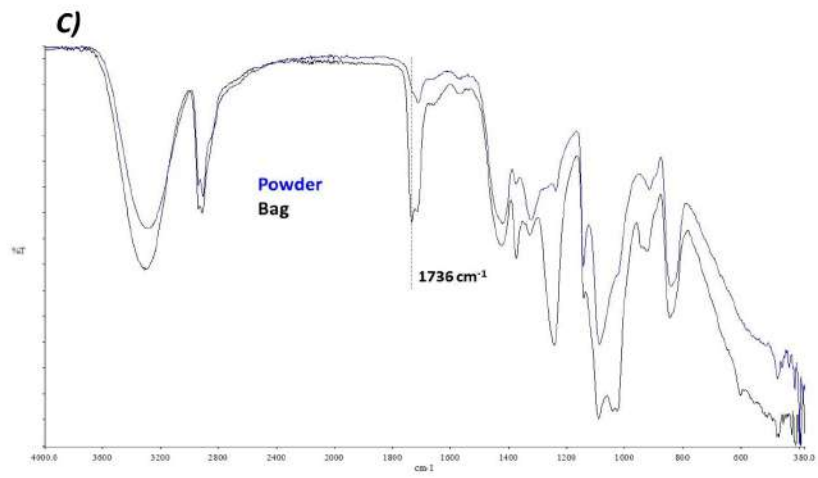
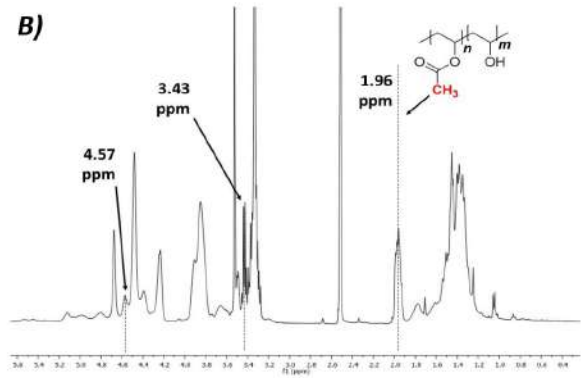
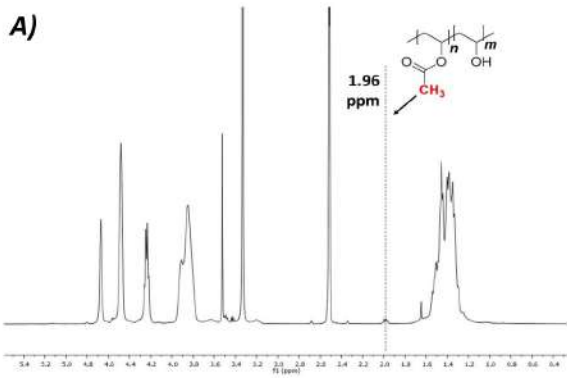
Behavioural alteration and MAO activity evaluation

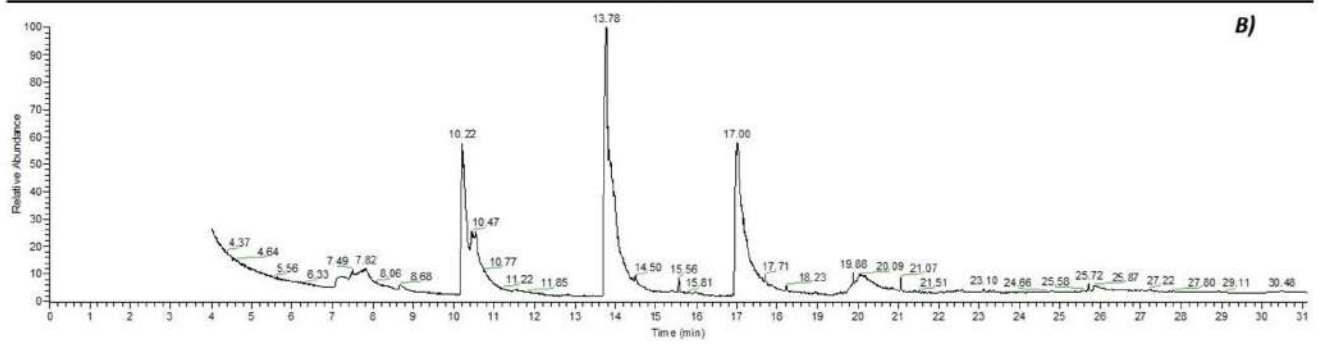
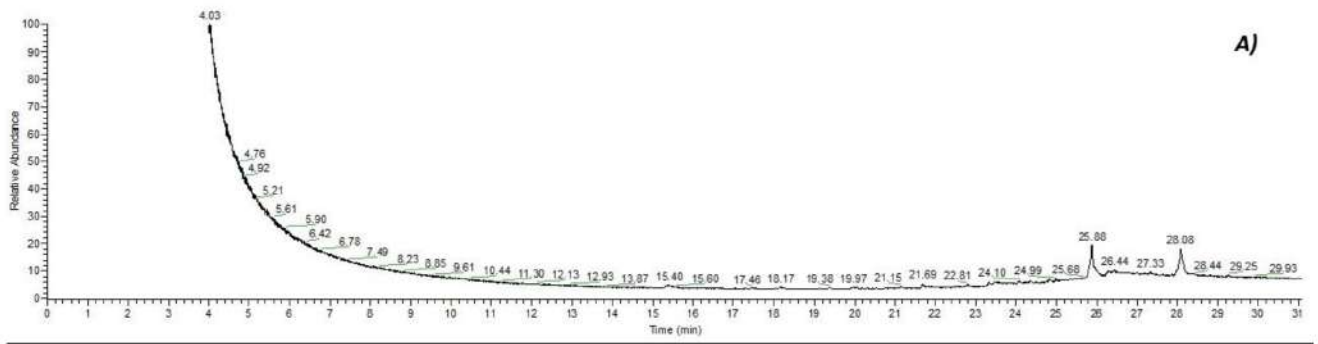


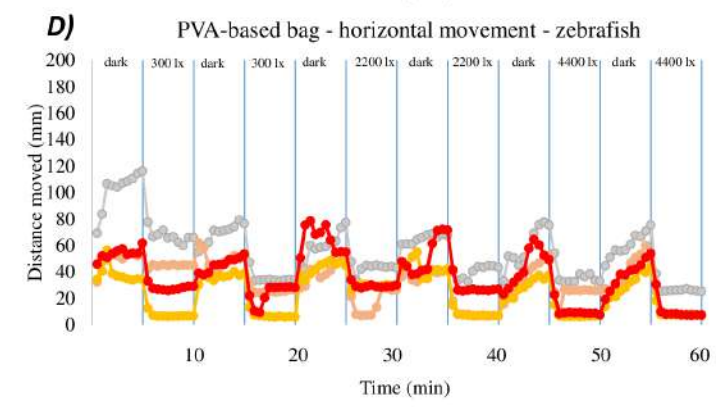
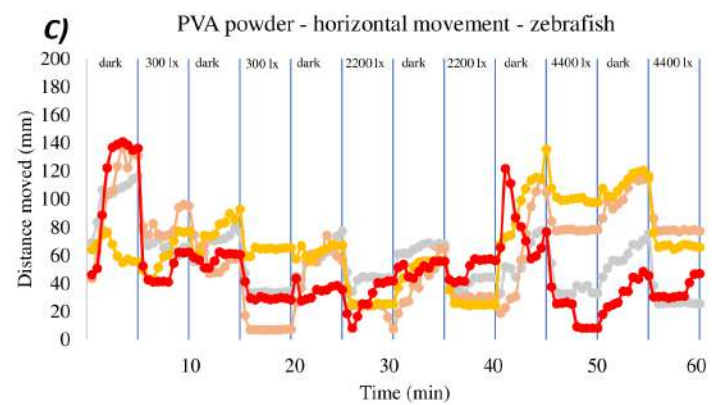
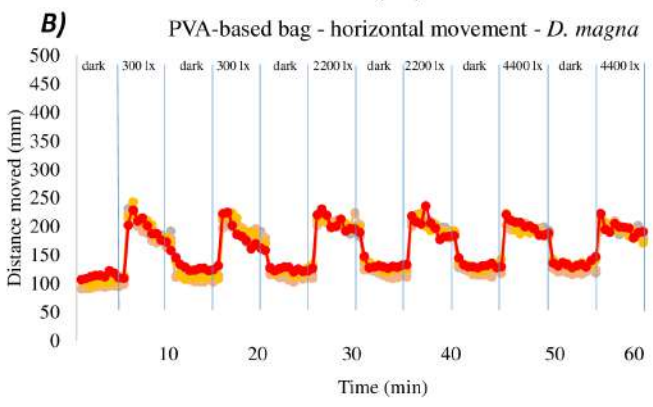
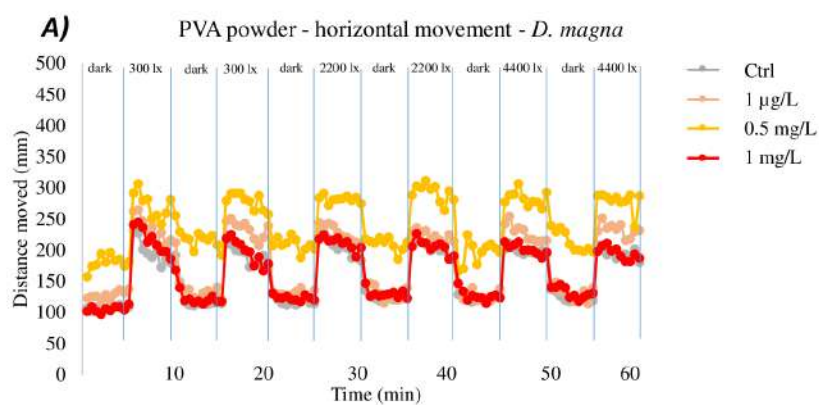
video-tracking system

NO EFFECTS

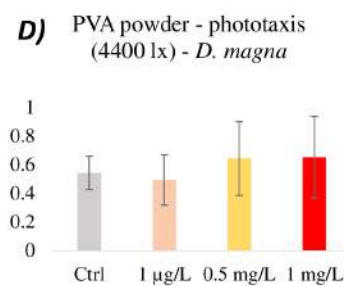
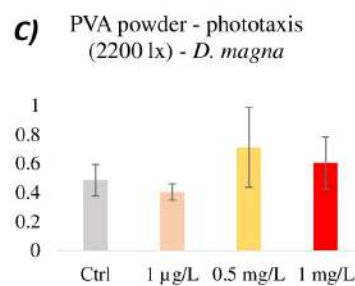
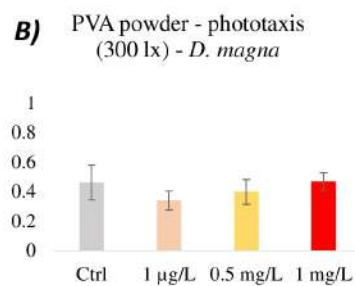
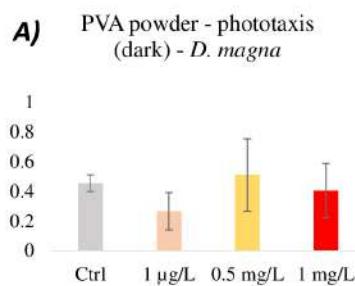




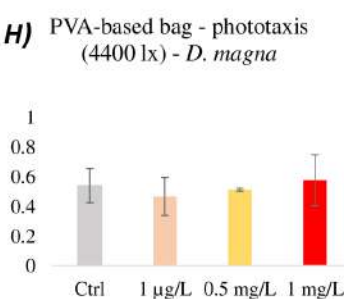
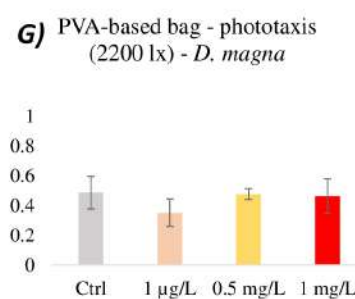
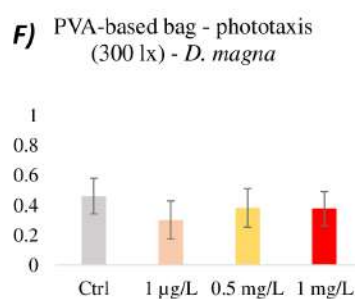
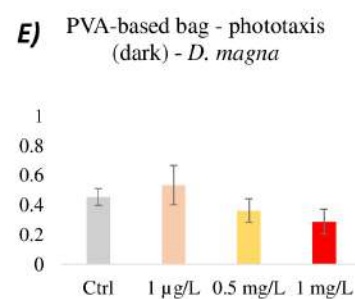


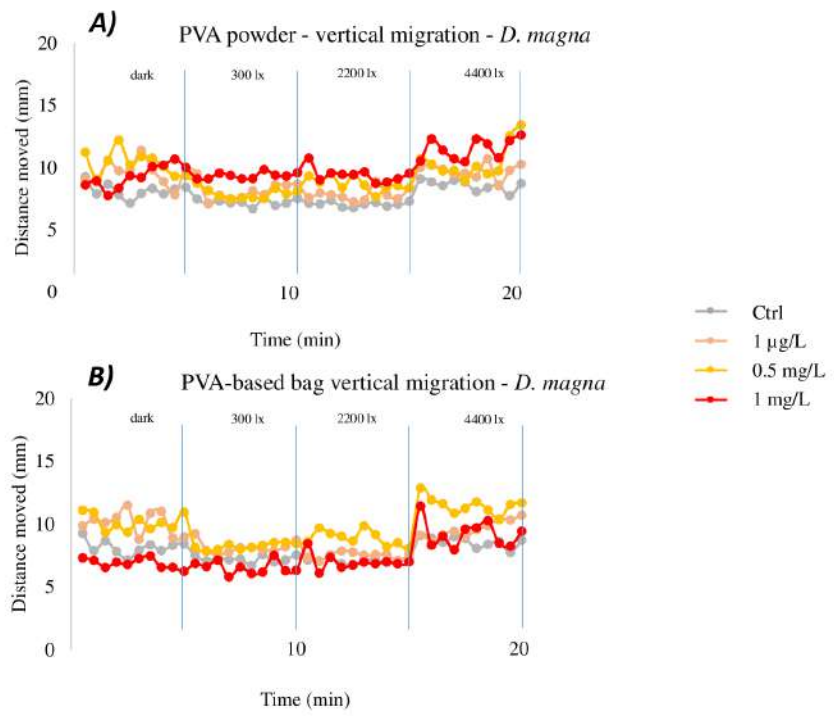


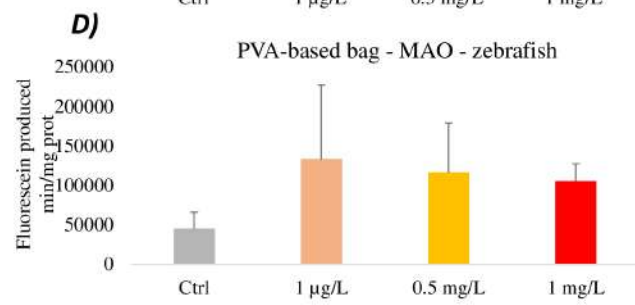
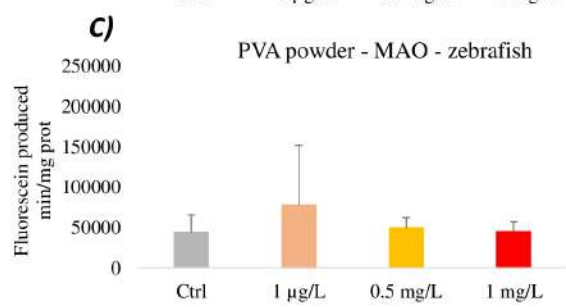
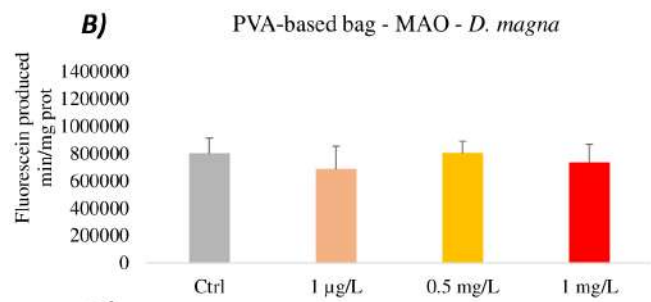
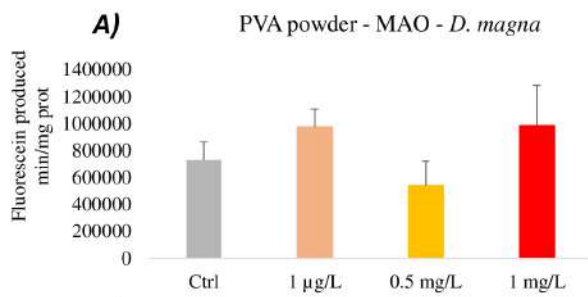
Distance moved ratio mm (zone 1/zone 2)



Distance moved ratio mm (zone 1/zone 2)







End-point	Species	Sequence of dark/light conditions	Time (min)
Horizontal movement	<i>D. magna</i> ✓ zebrafish ✓	dark	5
		300 lx	5
		dark	5
		300 lx	5
		dark	5
		2200 lx	5
		dark	5
		2200 lx	5
		dark	5
		4400 lx	5
		dark	5
4400 lx	5		
Vertical migration	<i>D. magna</i> ✓	dark	5
		300 lx	5
		2200 lx	5
		4400 lx	5
Phototaxis	<i>D. magna</i> ✓	dark	5
		300 lx	5
		2000 lx	5
		4400 lx	5

Table 1: Sequence of dark/light conditions used in the behavioural parameter evaluation (horizontal movement, vertical migration and phototaxis).