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2 Factors affecting the efficiency of molecular species delimitation in a species-rich insect family

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4 *Running Title*

5 Molecular species delimitation efficiency

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19

20 ***Abstract***

21 In the contest of global biodiversity loss, molecular species delimitation approaches can be very
22 useful for accelerating species discovery through DNA taxonomy and inventory through DNA
23 metabarcoding. In this study, the effect of some intrinsic factors on the efficiency of various
24 single-marker species delimitation methods (fixed and variable nucleotide distance thresholds,

25 ABGD, ASAP, GMYC, mPTP) was tested on more than 90 empirical datasets, derived from a
26 set of 7,237 COI sequences attributed to 542 leaf beetles species (Coleoptera: Chrysomelidae).
27 The considered factors were: i) the number of haplotypes per species (as a proxy for genetic
28 diversity); ii) the geographic distance among conspecific collection localities (as a proxy of
29 sampling width); iii) the difficulty related to morphological identification of species; iv) the
30 taxonomic rank. Distance-based methods, with on average more than 70% of match with
31 morphological identification, outperformed those relying on phylogenetic trees, with less than
32 59%. A high number of haplotypes per species was found to have a negative effect on
33 delimitation efficiency, whereas large geographic distances within species had a positive effect.
34 All methods delimitations (except for GMYC) were significantly affected by the presence of
35 species that are difficult to be identified, decreasing their efficiency. Finally, the only method
36 influenced by the taxonomic rank of the dataset was GMYC, showing lower efficiency in
37 datasets at the genus than at higher levels. The observed biases we highlighted affecting
38 efficiency could be accounted for when developing input datasets for species delimitation
39 analyses to obtain a more reliable representation of biological diversity.

40

41 ***Keywords***

42 Chrysomelidae; ABGD; ASAP; GMYC; mPTP; Cytochrome c oxidase subunit I COI;

43

44 ***Introduction***

45 In the context of the rapid impoverishment of global biodiversity (Hallmann et al., 2017;
46 Sánchez-Bayo & Wyckhuys, 2019; Van Klink et al., 2020), it is easy to understand the reason
47 of the success of molecular taxonomy that, exploiting molecular information, has the potential
48 to accelerate the identification of organisms and the discovery process of new taxa (Hebert &

49 Gregory, 2005; Swartz, Mwale, & Hanner, 2008; Monaghan et al., 2009; Mutanen, Kaila, &
50 Tabell, 2013; Gaytán et al., 2020). The DNA-based identification methods represented by DNA
51 taxonomy, applicable also in DNA metabarcoding (Taberlet, Coissac, Pompanon, Brochmann,
52 & Willerslev, 2012), stand out as widely used approaches for biodiversity surveys (e.g., Telfer
53 et al., 2015; Pavan-Kumar, Gireesh-Babu, & Lakra, 2015; deWaard et al., 2019). These
54 methods, using standardized gene regions, allow the speeding up of organism identification and
55 help overcoming issues in morphological taxonomy, for example by making it possible to detect
56 the presence of organisms merely from their DNA, released in the environment (e.g., Ficetola,
57 Miaud, Pompanon, & Taberlet, 2008; Montagna et al., 2018; Ruppert, Kline, & Rahman, 2019).
58 Beside molecular identification of already known species, species delimitation tools in DNA
59 taxonomy allow inferring hypothetic species and/or evolutionary significant units from
60 molecular data (Puillandre, Lambert, Brouillet, & Achaz, 2012; Tang, Humphreys, Fontaneto,
61 & Barraclough, 2014). Species delimitation molecular methods are used for both biodiversity
62 investigation without a priori hypothesis on the possible species (Dincă et al., 2015; Gómez-
63 Zurita, 2016) or, more frequently, as support to resolve taxonomic issue when other delimitation
64 approaches give uncertain results (e.g., Montagna et al., 2016; Kajtoch, Montagna, & Wanat,
65 2018; Plewa et al., 2018; García-Melo et al., 2019), in the framework of the so-called
66 “integrative taxonomy” approach (Dayrat, 2005). Several molecular species delimitation
67 methods have been proposed: early methods analysed alloenzymes variability, allele frequency
68 or nuclear genes co-dominance (Highton, 1989; Porter, 1990; Davis & Nixon, 1992; Good &
69 Wake, 1992; Highton, 2000), but, currently, the most widely adopted molecular delimitation
70 methods rely on DNA sequences as markers. The result of species delimitation analyses using
71 nucleotide sequences is more accurate when taking information from more than one marker
72 (Rubinoff & Holland, 2005; Dupuis, Roe, & Sperling, 2012); nevertheless, single-locus species

73 delimitation approach is widely applied on data derived from DNA barcoding studies (DeSalle
74 & Goldstein, 2019; Puillandre, Brouillet, & Achaz, 2020) and is still the only reliable approach
75 for metabarcoding. Even when multiple markers are used for metabarcoding, the origin of the
76 different markers cannot be considered from the same individual and any delimitation analysis
77 has to be performed independently for each marker. Molecular taxonomy approaches are
78 particularly useful for investigating diversity of those organisms for which morphological
79 identification is challenging such as bacteria, fungi but also invertebrates (Göker, García-
80 Blázquez, Voglmayr, Tellería, & Martín, 2009; Seifert, 2009; Ceccarelli, Sharkey, & Zaldívar-
81 Riverón, 2011). For most invertebrates, the standard marker for DNA barcoding is a fragment
82 at the 5' region of Cytochrome c oxidase I (COI), which is considered robust and reliable due
83 to a nucleotide variability that allows the discrimination of organisms at species level and, in
84 particularly, due to the highly consolidate primers for its amplification, suitable for the majority
85 of the groups (Hebert, Cywinska, Ball, & De Waard, 2003; Mioduchowska, Czyż, Goodyen,
86 Kur, & Sell, 2018). One of the first methods proposed for delimiting species starting from COI
87 barcodes is a phenetic approach that relies on sequences pairwise nucleotide distances (Sneath
88 & Sokal, 1973). The approach is based on the idea that a “barcoding gap” in the frequency
89 distribution of intra- and inter-specific distances of different taxa occurs at the same value of
90 distance (Hebert et al., 2003), allowing to define a fixed threshold to discriminate between intra-
91 and inter-specific nucleotide variability (Hebert et al., 2003). This approach has been largely
92 criticized since a fixed threshold value may not be suitable for delimiting species when taxa
93 with different evolutionary histories are analysed together (e.g., Meyer & Paulay, 2005). In
94 light of this, the application of group-specific *ad hoc* thresholds is supposed to result in more
95 accurate delimitations. Without any a priori species hypothesis, the *ad hoc* threshold for
96 delimiting a group of taxa can be estimated looking for the barcoding gap in the frequency

97 distribution of nucleotide distances, but since a real gap infrequently occurs, a minimum in the
98 distribution can be identified as a point of transition between intra- and inter-specific level.
99 According to this idea, the Automatic Barcode Gap Discovery method (ABGD; Puillandre et
100 al., 2012) was developed. ABGD stand out as one of the most used species delimitation methods
101 (more than 1,800 citation in Google Scholar up to January 2021) and, very recently, from the
102 same authors a new method named ASAP was released (Puillandre et al., 2020). ASAP was
103 developed to be even more user-friendly than ABGD, improving the choice of priors and of the
104 ultimate delimitation, which in ABGD are left to the user. Alternative methods to distance-
105 based one, explicitly accounting for evolutionary processes (Hickerson, Meyer, & Moritz,
106 2006), are represented by the phylogenetic and coalescent-based methods; among them, the
107 most used ones are the Generalized Mixed Yule Coalescent model (GMYC, Pons et al., 2006;
108 Fujisawa & Barraclough, 2013) with >1700 citations in Google Scholar up to January 2021, its
109 Bayesian implementation bGMYC (Reid & Carstens, 2012), the Poisson Tree Process (PTP,
110 Zhang, Kapli, Pavlidis, & Stamatakis, 2013) with >1200 citations, and its multi-rate extension
111 (mPTP, Kapli et al., 2017), in addition to other approaches that have not been so frequently
112 used (e.g. K/theta by Birky, Adams, Gemmel, & Perry, 2010; haploweb by Flot, Couloux, &
113 Tillier, 2010; both with <120 citations). GMYC and PTP require as input a phylogeny of taxa
114 estimated from DNA sequences: GMYC uses the topology of a tree to identify the maximum
115 likelihood solution separating the branches between species modelled by a Yule process and
116 the branches within species modelled by neutral coalescent (Fujisawa & Barraclough, 2013);
117 PTP finds the transition point between intra- and inter-specific processes assuming a two
118 parameter model that accounts for speciation and for the coalescent process based on the
119 Poisson distribution of branch lengths (Zhang et al., 2013; Kapli et al., 2017).
120 Despite the fact that DNA-based species delimitation methods are widely used and

121 consolidated, factors intrinsic to the analysed data that could lead to improper delimitation
122 results still need to be empirically investigated. Some studies were carried out for this purpose,
123 but the majority of them were performed on simulated data where dataset characteristics were
124 a priori defined in order to test for specific hypotheses. Among these tests, some evaluated the
125 effect of sampling scale, variation of the effective population size, of speciation rate and
126 mutation rate, marker length and sampling size, on GMYC, ABGD and PTP performances
127 (Lohse, 2009; Reid & Carstens, 2012; Esselstyn, Evans, Sedlock, Anwarali Khan, & Heaney,
128 2012, Fujisawa & Barraclough, 2013; Dellicour & Flot, 2018). However, when analysing real
129 data, almost none of the previous factors can be realistically accounted for. Testing the factors
130 affecting species delimitation on real data can help to define practical guidelines to establish
131 the most reliable experimental design and/or to account for biases when reading delimitation
132 results. Some studies in this direction are available (Talavera, Dinca, & Vila, 2013, Ahrens et
133 al., 2016, Pentinsaari, Vos, & Mutanen, 2017), but further factors and methods should be
134 considered to obtain a complete overview on the potential biases of the methods.

135 The aim of the study is to analyse the influence of data-related factors on the efficiency of six
136 molecular species delimitation methods (i.e., fixed nucleotide distance threshold, *ad hoc*
137 developed nucleotide distance threshold, ABGD, ASAP, GMYC, and mPTP). A collection of
138 ~7,200 COI sequences belonging to 542 Euro-Mediterranean leaf beetles species (Coleoptera:
139 Chrysomelidae) was investigated. The sequences are representative of almost ¼ of Euro-
140 Mediterranean Chrysomelidae fauna, a family including almost entirely phytophagous species,
141 but with highly different ecological traits regarding habitat (e.g., Brunetti, Magoga, Iannella,
142 Biondi, & Montagna, 2019, Mende, Biström, Meichssner, & Kölsch, 2010), trophic
143 specialization (Mardulyn, Milinkovitch, & Pasteels, 1997; Fuss, Geiser, & Patzner, 2005),
144 range of distribution (Schmitt & Rönn, 2011; Biondi, Urbani, & D'Alessandro, 2013), size and

145 dispersal abilities (Strauss, 1988; Mende et al., 2010; Piper & Compton, 2010). Chrysomelidae
146 COI sequences were assembled in datasets with different characteristics to test the impact on
147 molecular delimitation efficiency of the following data-related factors: i) the number of
148 haplotypes per species, as a proxy for species *genetic diversity* (and potentially also for sample
149 size). A higher number of haplotypes per species could have differential impacts on species
150 delimitation methods: on the one hand genetic variability will be better represented (Goodall-
151 Copestake, Tarling, & Murphy, 2012), but on the other hand it could blur the differences
152 between species. ii) The *sampling structure* evaluated considering geographic distances
153 between collection localities of conspecifics. Sampling structure could impact species
154 delimitation efficiency, especially when genetic and geographic distances are related (Mason,
155 Fletcher, Gill, Funk, & Zamudio, 2020); a discontinuous sampling could cause oversplit
156 (Lohse, 2009; Talavera et al., 2013; Mason et al., 2020), a limited sampling could favour
157 delimitation efficiency (Pentinsaari et al., 2017), whereas wider sampling could result in both
158 split of species or better delimitation possibly depending on species population structure. iii)
159 The *taxonomic rank* of the dataset, in this work considered at the genus, subfamily and family
160 level, could be viewed as a proxy of the overall phylogenetic distance among the analysed
161 species. Having only closely related sequences or also sequences from rather distant species in
162 a single dataset may affect the estimation of several parameters in the methods from DNA
163 taxonomy, including also reliability of the phylogenetic reconstructions. iv) The *morphological*
164 *distinctiveness* of the species, evaluated as the difficulty related to their morphological
165 identification. Morphological identification could be easy in Chrysomelidae, with clear and
166 unambiguous features for some groups of species but in other cases only subtle differences of
167 the genital apparatuses are present as diagnostic characters (e.g., Bezdek & Baselga, 2015;
168 Montagna et al., 2016; Magoga, Coral Sahin, Fontaneto, & Montagna, 2018). Closely related

169 species, with only subtle differences in the copulatory system, may have recently diverged, with
170 no time for a clear differentiation between species in the molecular marker chosen for molecular
171 identification.

172 We explore the variability in the efficiency of the different methods for molecular identification
173 of species comparing the match with morphological identification, here considered as the most
174 appropriate descriptor of diversity in this well-studied group of insects (Doguet, 1994;
175 Warchałowski, 2003; Nadein, 2013; Petitpierre, 2016; Nie et al., 2019; Petitpierre, 2019),
176 disentangling the effects of the potential confounding factors we described including genetic
177 diversity, sampling structure, taxonomic rank, and morphological distinctiveness.

178

179 ***Material and Methods***

180 *Datasets*

181 In the present work the collection of COI sequences of Chrysomelidae (Insecta: Coleoptera)
182 developed and analysed in Magoga et al. (2018) was used as reference for the performed
183 analyses. This COI collection consists of 7,237 sequences (average 652 bp; range: 460 bp to
184 658 bp) that are taxonomically assigned to 542 species. The Magoga et al. (2018) COI sequence
185 collection was divided into multiple datasets according to taxa membership to the taxonomic
186 level of family, subfamily and genus (Fig. 1A); then the datasets were aligned at codon level
187 using MUSCLE in MEGA X (Kumar, Stecher, Li, Knyaz, & Tamura, 2018). The obtained
188 alignments were used as input for the species delimitation analyses, *per se*, after outgroups
189 inclusion, or after the inference of a phylogenetic tree (see below). The datasets used to infer
190 phylogenetic trees, input of tree-based species delimitation methods, were collapsed with R
191 software 3.6.2 (R Core Team, 2019) in order to retain for each species only unique haplotypes.

192

193 *Molecular species delimitation analyses*

194 In the present study, the following nucleotide distance-based and tree-based molecular species
195 delimitation methods were adopted (Fig. 1A): *i.* the 3% nucleotide distance threshold proposed
196 by Herbert et al. (2003); *ii.* nucleotide distance thresholds *ad hoc* estimated on each dataset; *iii.*
197 ABGD (Puillandre et al., 2012); *iv.* ASAP (Puillandre et al., 2020) *v.* GMYC (Pons et al., 2006;
198 Fujisawa & Barraclough, 2013); *vi.* the Multi-rate Poisson tree processes (mPTP; Kapli et al.,
199 2017). The 3% nucleotide distance threshold and *ad hoc* threshold analyses were carried out on
200 96 datasets (one family, 12 subfamily, and 83 genus datasets). ABGD and ASAP analyses were
201 performed on 94 datasets (one family, 12 subfamily, and 81 genus datasets) since datasets
202 composed by less than three sequences were excluded from the analyses. GMYC and mPTP
203 analyses were performed on phylogenetic trees inferred on 92 datasets (one family, 12
204 subfamily, and 79 genus datasets) by excluding two additional genus datasets with only one
205 sequence per species after haplotype collapsing.

206 The K2P pairwise nucleotide distance matrices, required for the species delimitation through
207 3% nucleotide distance threshold and *ad hoc* nucleotide distance threshold methods, were
208 estimated for each dataset using the R software library APE 5.3 (Paradis, Claude, & Strimmer,
209 2004). The *ad hoc* nucleotide distance threshold for each dataset was estimated using R function
210 *localMinima* of package SPIDER 1.5.0.9000 (Brown et al., 2012). The function, based on the
211 barcoding gap concept, identifies the minima in the density of nucleotide distances as possible
212 thresholds for delimiting the species present in the analysed dataset. The R function *tclust* of
213 package SPIDER was subsequently used to cluster nucleotide sequences at the distance threshold
214 of 3% and at the *ad hoc* threshold value previously identified.

215 ABGD analyses were performed using the command-line version downloaded from
216 <https://bioinfo.mnhn.fr/abi/public/abgd/> with the following settings: K2P nucleotide

217 substitution model (Kimura, 1980) to infer the nucleotide distance; relative gap width of 1.5,
218 when gap was not found using this value, a width of 1 and 0.5 were set; prior P ranging from
219 0.001 to 0.1 and the remaining parameters were left as default. ASAP analyses were run using
220 the program web-interface (<https://bioinfo.mnhn.fr/abi/public/asap>); K2P (Kimura, 1980) was
221 selected as nucleotide substitution model and other parameters were left as default. ASAP
222 delimitation was defined evaluating both the partitions with first and the second best asap-score
223 according to Puillandre et al., 2020.

224 In order to perform GMYC and mPTP species delimitation analyses, an ultrametric tree was
225 inferred using the software BEAST v 1.8 (Drummond, Suchard, Xie, & Rambaut, 2012) from
226 each haplotype-reduced dataset (Tang et al., 2014). One or more (in the case of the family
227 dataset) orthologous sequences from the appropriate outgroup was included to properly root the
228 tree. For each dataset two to five independent MCMC runs were performed using the following
229 parameters: Markov chain length from 30×10^6 to 300×10^6 generations depending on the runs
230 convergence assessed examining the estimated sample size of each parameter of the model and
231 a visual inspection of the likelihood with TRACER (Drummond et al., 2012); sampling of trees
232 and parameters every 1,000 to 5,000 generations depending on the total number of generations;
233 models of nucleotide evolution as selected according to the Bayesian information criterion after
234 the analysis performed by jModelTest 2 (Darriba, Taboada, Doallo, & Posada, 2012); Yule
235 process as speciation model (Yule, 1925); all other priors set to their default values. The runs
236 were then pooled or resampled according to the number of performed generations, after removal
237 of the proper tree burn-in fraction, using LogCombiner (Drummond et al., 2012) and the
238 majority-rule consensus tree obtained by TreeAnnotator (Drummond et al., 2012).

239 Single-threshold GMYC species delimitation analyses were performed using the default
240 settings of the R function *gmyc* of the library SPLITS 1.0.19 (Ezard, Fujisawa, & Barraclough,

241 2009).

242 mPTP analyses were performed through the binary version 0.2.4 available on
243 <https://github.com/Pas-Kapli/mptp>. For each dataset, we performed ten different runs with the
244 following settings: mcmc run of 100×10^6 generations (steps), sample frequency every 5,000
245 generations and a burnin of 20,000 generations; the convergence of the independent runs was
246 assessed through the average standard deviation of delimitation support values (ASDDSV) and
247 the overall support for the ML estimate calculated computing the mean of the average support
248 values (ASV) over the ten runs.

249 The delimitation results (called units) obtained for each analysis were compared to the identity
250 of the morphological species and classified in the following categories adopting an *ad-hoc*
251 developed R script: i) *match* = all the sequences of the same morphological species were
252 delimited as belonging to the same unit; ii) *split* = the sequences of a species are delimited as
253 belonging to two or more units; iii) *merge* = the sequences of two or more species are included
254 in the same unit; and iv) *mixture* = some sequences of a species are split while others are merged
255 (Fig. 1A).

256

257 *Factors affecting species delimitation*

258 In this work the number of observed matches in each analysis was used as a proxy of the
259 achieved efficiency. Efficiency of the different species delimitation methods was defined as the
260 number of matches (successes) against the number of failures, counted as the sum of merges,
261 splits, and mixtures.

262 The factors potentially affecting species delimitation were represented by a series of data-
263 related variables (Fig. 1B). The factors took in account: *i.* the mean number of haplotypes per
264 species in a dataset (genetic diversity and sample size); *ii.* the median value of the geographic

265 distance between conspecific collection localities (sampling structure and width); *iii.* the
266 difficulty level of the morphological identification of the species (morphological distinctiveness
267 of species); *iv.* the taxonomic rank of the dataset.

268 The number of haplotypes was calculated for each species using R and then for each dataset the
269 average number estimated. The geographic distances between collection localities were
270 calculated with the R package GEOSPHERE 1.5.10 (Hijmans, 2017) starting from the geographic
271 coordinates of the sampling points in WGS84 system (Table S1). The median value for each
272 species was then extracted and for each dataset, the median of these values was calculated. As
273 already mentioned, the analysed datasets belong to different taxonomic ranks (i.e., family,
274 subfamily, and genus); these ranks were accounted to evaluate their influence on delimitation
275 efficiency. In order to assess whatever there is an effect of morphological distinctiveness, each
276 species present in this study was assigned to one of the following three categories of
277 morphological identification difficulty: *i.* species that are easy to be morphologically identified
278 due to the presence of clear and easily detectable diagnostic characters (level I); *ii.* species that
279 can be identified through subtle differences in morphological characters (excluding genitalia)
280 or on the basis of strong differences in the genitalia shape (spermatheca or the median lobe of
281 the aedeagus) (level II); and *iii.* species that are identifiable only on the basis of subtle
282 differences in the shape of genitalia (level III). The level of difficulty was assigned to species
283 using information from specialized literature reporting species descriptions and dichotomous
284 keys (e.g., Müller, 1953; Burlini, 1955; Doguet, 1994; Warchałowski, 2003, Petitpierre, 2016;
285 Petitpierre, 2019). For each dataset, the proportion of species assigned to each difficulty level
286 was calculated and registered in two variables, namely proportion of difficult species (as level
287 II + level III) and proportion of extremely difficult species (only level III), in order to assess
288 the effect of the increasing difficulty.

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Statistical analyses

Before performing any explicit test, the presence of correlation between the explanatory variables was assessed through the R library PSYCH 1.9.12.31 (Revelle, 2018).

A generalized linear mixed-effects model (GLMEM) was then run for testing the differential influence of the explanatory variables that passed the preliminary test of multiple correlation (namely number of haplotypes, geographic distance, taxonomic level, taxonomic difficulty and taxonomic extreme difficulty) and the effect of species delimitation method on efficiency. The model was fitted using the R package LMERTEST 3.1.2 (Kuznetsova, Brockhoff, & Christensen, 2017) using a negative binomial distribution given that the response variable was proportional data, and adding the identity of dataset as a random effect in the error structure of the model to remove the effect of pseudoreplication, given that the same datasets were analysed with different methods. In addition, the model included the interaction terms of each explanatory variable with the species delimitation method.

In order to understand the differences between methods, additional statistical models were run separately for each delimitation method using GLMs with the same structure of explanatory variables of the previous GLMEM (obviously, excluding method).

The datasets of individuals at the genus, subfamily and family level are nested, but the data used in the models are not nested, because the analyses were performed separately for each taxonomic rank; thus, the nested effect of the taxonomic ranks was not included as a random effect in the error structure of the models.

The outcomes of GLMEM and GLMs are reported in the results section as Type II analysis-of-variance tables fitted with the R package CAR 3.0.6 (Fox & Weisberg, 2019).

313 **Results**

314 *Datasets composition and species delimitation efficiency*

315 The set of 7,237 sequences was organized into 99 datasets according to the taxonomic levels,
316 resulting in one family, 12 subfamily, and 86 genus datasets. Three genus datasets were
317 excluded from the delimitation analyses since they were composed by only one COI sequence.

318 The datasets were highly different in term of species and COI sequences number ranging from
319 the family dataset composed by 542 species and 7,237 sequences (4,066 after haplotype
320 reduction) to a mean of 6.5 species [range 1:93] and 87 sequences [2:1,014] for the genus
321 datasets (50.1 sequences after haplotypes reduction in mean) (Table S2).

322 Overall, 64.4% of the species were delimited in the same category by all the four delimitation
323 methods when the family dataset is considered, 63.8% in the case of subfamily datasets, and
324 only 50.9% at the genus level.

325 The highest percentage of matches between molecular delimited units and morphospecies
326 (80.5% on average) was obtained delimiting species through the *ad hoc* nucleotide distance
327 thresholds (Table S3, Table S4). Molecular delimitation using ABGD and ASAP resulted in a
328 similar percentage of matches (77.6% and 77.9% on average, respectively), moderately higher
329 than using the 3% nucleotide distance threshold (72.8% on average) (Fig. 2A). Ad hoc
330 nucleotide distance threshold, ABGD and ASAP showed also a nearly equal efficiency at any
331 taxonomic level (Fig. 2B). The likelihood-ratio test of GMYC analyses rejected the null model
332 (p -value < 0.05) for 51 out of 94 datasets, thus being unable to delimit molecular taxonomic
333 units in almost half of the datasets (43 datasets, Table S4); the average percentage of matches
334 resulted of 33.2% for this method (Fig. 2A). In mPTP analyses, ASDDSV values resulted $<$
335 0.01 for 90 over 94 datasets, indicating the convergence of the ten independent MCMC runs.
336 For the majority of the datasets (83 over 94) ASV values resulted high (median: 92.4%),

337 suggesting that the ML solution was supported by the data. For the 11 datasets with ASV values
338 lower than 50%, the results were considered not significant (Table S4). On average this method
339 produced 58.7% of match (Fig. 2A).

340 The highest number of merged morphological species was observed for GMYC, while split
341 species were more common from mPTP and 3% distance threshold delimitation (Fig. 2A). The
342 highest number of mismatches was associated to the genus datasets delimited by tree-based
343 methods, GMYC and mPTP (Fig. 2B). Delimitation results obtained on each dataset with each
344 method are reported in Table S4.

345

346 *Explanatory variables*

347 *Number of haplotypes.* Within the analysed datasets on average ~6 haplotypes per species were
348 present (range 1:40.5; Table S2).

349 *Geographic distance between conspecific collection localities.* A median distance among
350 collection localities of conspecifics of ~219 km was obtained with values ranging from 0
351 (species collected in one locality only) to 1,685 km (Table S2). The highest value was observed
352 for *Tituboea biguttata* for which two specimens were collected one in Italy and one in Morocco.

353 *Difficulty in species morphological identification.* About 35% of the species included in the set
354 of COI sequences analysed were easily identifiable morphologically (level I), 53.5% were
355 considered of intermediate difficulty (level II), 11.5% were extremely difficult (level III).

356 Among the species categorized in level I there are for example those of the genera *Zeugophora*
357 and *Syneta* and of the subfamily Hispinae; all the species included in level II belong to 11
358 genera (e.g., *Neocrepidodera*, *Orestia*, *Plagiosterna*, *Plateumaris*); most of the species of
359 *Altica* and *Oulema* were assigned to level III (Table S2).

360

361 *Factors affecting efficiency in species delimitation*

362 The five explanatory variables (mean number of haplotypes, median geographical distance,
363 proportion of difficult species, proportion of extremely difficult species, taxonomic rank) were
364 not correlated among each other (Fig. S1) and were kept in the analyses.

365 The efficiency of species delimitation was significantly different between the six molecular
366 species delimitation methods (GLMEM: $\chi^2 = 74.71$, $p < 0.0001$), with a complex scenario of
367 different factors differentially affecting the efficiency of the methods (Table 1).

368 Analysing each method separately, all factors were found to significantly affect species
369 delimitation efficiency (Table 2). The number of haplotypes per dataset resulted to affect the
370 efficiency of 3% nucleotide distance threshold, *ad hoc* nucleotide distance threshold, ABGD,
371 ASAP, and mPTP (Table 2, Table S5). A higher number of haplotypes correlated with lower
372 efficiency (Table 2, Table S5).

373 The efficiency of four delimitation methods (*ad hoc* nucleotide distance threshold, ABGD,
374 GMYC, and mPTP) was found to be positively related with the geographic distance between
375 conspecific collection localities (Table 2, Table S5). For these methods a larger geographic
376 range improved the proportion of matches.

377 GMYC resulted the only method to be affected by the taxonomic rank of the dataset (Table 2),
378 with efficiency decreasing when genus rank datasets were analysed (Fig. 2B, Table S4, Table
379 S5).

380 The presence in the dataset of species that are difficult to be identified (level II + level III) was
381 found to decrease ABGD efficiency (Table 2, Table S5), with a higher proportion of merges
382 related to difficult species. 3% threshold, *ad hoc* nucleotide distance threshold, ASAP, and
383 mPTP were negatively affected by the presence of extremely difficult species (level III) (Table
384 2, Table S5), with a higher proportion of mixtures in extremely difficult species.

385

386 ***Discussion***

387 *Species delimitation*

388 Our results suggested that the six molecular species delimitation methods considered in this
389 study could be differentially affected by the dataset features, as expected considering their
390 different assumptions and principles. Among the tested species delimitation methods, the
391 highest efficiency resulted from distance-based method analyses (Fig. 2A), with *ad hoc*
392 nucleotide distance threshold slightly outperforming ABGD and ASAP, and 3% threshold
393 coming next. Our results confirm that ASAP achieves its aim of overcoming the two main
394 limitations of ABGD, namely the need of defining a priori the maximum divergence of
395 intraspecific diversity (P), and especially, the lack of scores associated to partitions, forcing
396 ABGD users to choose independently the “best” partition among those resulting from the
397 analysis (Puillandre et al., 2020). Interestingly, only in the 69% of the analysed datasets one of
398 the two best ASAP-score partitions was identical to the partition selected as the “best” one from
399 the ABGD output. The almost identical overall efficiency of the two methods notwithstanding
400 the low proportion (69%) of shared selected partitions between them reveals that, regardless of
401 the details of the methods, the result is comparable and ASAP may soon replace ABGD in the
402 DNA taxonomy literature because of its advantages for the user.

403 In general, outperformance of distance-based methods over tree-based ones is consistent with
404 what has been previously observed (Pentinsaari et al., 2017; Hoffman et al., 2019) when
405 different species delimitation method performances were compared. In any case, it should be
406 considered that the delimitation obtained in this study using distance-based methods (except
407 for 3% nucleotide distance threshold) are not completely objective; in fact, the final choice of
408 the best partition, among many potential choices for *ad hoc* nucleotide distance threshold and

409 ABGD, or only between two choices for ASAP, needs to be made by the user. Despite GMYC
410 is known to be prone to oversplit species (Paz & Crawford, 2012; Hamilton, Hendrixson,
411 Brewer, & Bond, 2014; Hoffman et al., 2019), in our analysis it was not associated with a high
412 number of splits when compared to mPTP and the 3% threshold (Fig. 2); on the contrary, it
413 underestimated species diversity, having the highest number of merged species. Despite COI
414 intraspecific nucleotide distances of Chrysomelidae are known to be on average lower than 3%
415 (Kubisz, Kajtoch, Mazur, & Rizun, 2011; Germain et al., 2013; Montagna et al., 2016), the
416 species delimitation approach with such value as a fixed threshold led to a consistent number
417 of splits. This result is due to the presence of some taxa with a high intraspecific variability,
418 such as some species of the genus *Cassida*, which are known even to increase the optimal
419 threshold for molecular identification of Cassidinae to 5.9% (Magoga et al., 2018). For this
420 reason, 40% of Cassidine morphospecies were split using 3% threshold delimitation method. A
421 similar situation was observed for Cryptocephalinae, where delimitation analysis using 3%
422 threshold led to a consistent number of splits (22.5%, corresponding to 33 species out of 147).
423 Despite the high proportion of closely related species groups characterized by low values of
424 interspecific distances within this subfamily (Montagna, Sassi, & Giorgi, 2013; Montagna et
425 al., 2016; Magoga et al., 2018) that strongly decrease the optimal threshold value for molecular
426 identification (~1%, Magoga et al., 2018), values of intraspecific variability higher than the 5%
427 were estimated for some Cryptocephalinae species, probably due to differences in term of
428 evolutionary histories and/or ecology among the species (Magoga et al., 2018). Despite 3% not
429 being in principle the most likely threshold value for separating intra- to inter-interspecific
430 variability of the analysed taxa, the molecular delimitation using this fixed distance threshold
431 had only a slightly lower efficiency than other distance-based methods (Fig. 2A).

432

433 *Factors affecting species delimitation methods efficiency*

434 One of the factors found to affect the majority of the species delimitation methods adopted in
435 this study was the mean number of haplotypes per dataset. Higher mean numbers of haplotypes
436 were significantly related to a decreasing efficiency of all the methods, except for GMYC
437 (Table 2). This result could support the hypothesis that a higher number of sampled haplotypes
438 correlates with a higher probability to find intermediate haplotypes among closely related
439 species, in particular when sampling scale is large (Meyer & Pauly, 2005; Bergsten et al., 2012;
440 Pentinsaari, Hebert, Mutanen, 2014; Phillips, Gillis, & Hanner, 2019). A wider intraspecific
441 sampling could thus hide the barcoding gap and cause indecisive delimitations by adopting
442 distance-based methods, consequently decreasing their efficiency (Fig. 3). Furthermore, several
443 cases of splits were related to the 3% thresholds, which appeared too low to delimit the
444 intraspecific level of the most haplotype-abundant species (Fig. 3). Contrary to what has been
445 observed in this study, a positive relation between the number of haplotypes and the species
446 delimitation efficiency was previously observed for coalescent-tree based methods (assuming
447 an identical migration rate), with tendency to oversplit species when delimitation was
448 implemented on datasets including a low number of haplotypes per species (Lohse, 2009). In
449 our analyses, a low number of haplotypes per species did not result in an increase of the splits,
450 neither for GMYC nor for mPTP; on the contrary, mPTP frequently split haplotypes-rich
451 species. In fact, within the analysed datasets intraspecific sampling was often unbalanced,
452 including a large number of equally and poorly sampled species and few oversampled species.
453 Oversampled species resulted significantly prone to be oversplit by mPTP (Table S6), contrary
454 to what was observed to occur using a previous version of the PTP method (Zhang et al., 2013).
455 The efficiency of four methods (i.e., *ad hoc* nucleotide distance threshold, ABGD, GMYC, and
456 mPTP) resulted to significantly increase when the median geographic distance among the

457 sampling localities of conspecifics increases. It is already known that, when intraspecific
458 sampling is incomplete, the main factor affecting the efficiency of coalescent tree-based
459 methods is the migration rate among species demes, given that high migration rate corresponds
460 to low delimitation efficiency and *vice versa* (Lohse, 2009). Our results are in agreement with
461 this evidence since the delimitation efficiency increases when the geographic distance among
462 conspecific collecting points increases, a condition that likely reduces or prevents the migration
463 of individuals among demes.

464 Despite that *ad hoc* nucleotide distance threshold and ABGD efficiency resulted good on the
465 majority of datasets, low values of median geographic distance among conspecifics sampling
466 localities were found to be related with more inaccurately delimited species. Different factors
467 could have affected the recovery of the barcoding gap in these cases. In some datasets where
468 geographic distance values were low and closely related species were included, the methods
469 tended to merge different species in the same partition (Fig. 4A). Possible explanations rely on
470 the recent divergence of the taxa and/or on the incomplete lineage sorting phenomenon, both
471 preventing the recovery of the barcoding gap when within-species geographic sampling is
472 limited. Higher efficiency on datasets with high median geographic distance among
473 conspecifics sampling localities values (e.g., *Agelastica* = 1,470 km; *Lilioceris* = 1,528 km;
474 *Plagiosterna* = 1,172 km) could be related to the species composition of those datasets. Since
475 the analysed data were generated by barcoding studies performed within four European areas,
476 the most common and widely distributed west-Palearctic species have likely been collected
477 from all the investigated areas. These species show low values of nucleotide intraspecific
478 distances indicating a panmictic-like condition that appears to promote delimitation efficiency
479 (Fig. 4B). Moreover, a sub-optimal efficiency has been observed for species-rich datasets,
480 where a clear barcoding gap occurs less frequently (Dellicour & Flot, 2018; Hoffman et al.,

481 2019), corresponding to intermediate values of median geographic distance among conspecifics
482 sampling localities.

483 Further factors negatively affecting delimitation methods efficiency (except for GMYC) were
484 the presence in the analysed datasets of species that are difficult and extremely difficult to be
485 morphologically identified (difficulty level II+III and difficulty level III). A consistent
486 behaviour between variability of the COI marker and species morphology is plausible; in the
487 cases of recently diverged species, a high morphological similarity is mirrored by the lack of
488 enough mutations on COI gene to make it effective for distinguishing between species (Nice &
489 Shapiro, 1999; van Velzen, Weitschek, Felici, & Bakker, 2012; Chapple & Ritchie, 2013). On
490 the other hand, despite the proven accuracy of morphological identification of specimens
491 included in this study (Gómez-Rodríguez, Crampton-Platt, Timmermans, Baselga, & Vogler,
492 2014; Pentinsaari, Hebert, & Mutanen, 2014; Hendrich et al., 2015; Magoga et al., 2018),
493 misidentifications may have implications in the observed pattern, especially in relation to
494 species of extremely difficult identification.

495 GMYC resulted also the only method whose efficiency could be significantly affected by the
496 taxonomic rank, in particular a decrease in the efficiency was observed with genus level data
497 (Fig. 2B). In accordance to what was already observed by Talavera et al. (2013), the lack of
498 significance of GMYC analyses in separating species within a genus could be related to the
499 very low number of species included in some datasets, rather than to the taxonomic rank of the
500 dataset itself. For our datasets, the majority of the not significant GMYC analyses (41 over 43)
501 were indeed related to datasets including ≤ 3 species (39 genus datasets, 2 subfamily datasets).
502 Even if the influence of the taxonomic rank on the other delimitation methods efficiency was
503 not significant, it was possible to observe that an appreciably better efficiency of distance-based
504 methods was present on genus datasets (Fig. 2B).

505

506 **Conclusions**

507 The results of this study demonstrated how 3% distance threshold, *ad hoc* nucleotide distance
508 threshold, ABGD, ASAP, GMYC, and mPTP delimitation efficiency are influenced by
509 different data-related factors including the number of haplotypes per species, the geographic
510 distance between sampling points of individuals of the same species, the difficulty related to
511 species morphological identification, and the taxonomic rank of the analysed dataset. Despite
512 the difficulty of planning a species delimitation study accounting for all these factors, their
513 impact on the delimitation should be considered and possibly avoided before performing any
514 delimitation analysis. If their avoidance is not possible, they should be carefully considered as
515 a caveat before drawing any conclusion from the results. Moreover, the quite common practice
516 of comparing the results of various molecular species delimitation methods in species
517 delimitation studies could also avoid to accept improper delimitations, considering that data-
518 related factors differentially affected the efficiency of the methods.

519

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523

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847

848 **Data accessibility**

849 All sequences analysed are publicly available in GenBank, their accession numbers are listed
850 in Table S1, as well as the coordinates of collection localities. Information on each analysed
851 dataset, from which GLMEM and GLMs independent variables are derived, is reported in Table
852 S2. Results of each performed species delimitation analysis are summarized in Table S4. R
853 scripts used to perform statistical analyses and to tabulate species delimitation results according
854 to categories (match, merge, mixture, split) are available at
855 <https://github.com/MontagnaLab/species-delimitation-methods-comparison>.

856

857 **Author contributions**

858 G.M., M.M. and D.F. designed the study. M.M. inferred species delimitation input trees. D.F.
859 and G.M. performed the statistical analyses. G.M. analysed the data and wrote the manuscript.
860 M.M. and D.F. contribute to the final version of the manuscript.

861

862 **Table legends**

863 Table 1. Effect of species delimitation method, number of haplotypes, geographic distance,
864 taxonomic difficulty and taxonomic level on efficiency. The results are reported as a Type II
865 analysis-of-variance table on a Generalised Linear Mixed Effects Model (GLMEM) object.

866

867 Table 2. Effect of number of haplotypes, geographic distance, taxonomic difficulty and
868 taxonomic level on efficiency separately for each species delimitation method. The results are
869 reported as a Type II analysis-of-variance table on Generalised Linear Model (GLM) objects.

870

871 **Figure legends**

872 Fig. 1. Rationale of the study. (a). datasets, molecular species delimitation analyses, and
873 categorization of results. (b). statistical analyses.

874

875 Fig. 2. Efficiency of the species delimitation methods. (a). Mean percentage of match (red),
876 merge (yellow), mixture (light blue), and split (purple) observed for each method. (b).
877 Efficiency of the species delimitation methods recovered in three taxonomic levels, i.e.,
878 family (blue), subfamily (light blue) and genus (teal).

879

880 Fig. 3. Impact of the number of haplotypes on the 3% threshold delimitation efficiency.
881 Distribution of pairwise nucleotide distances in five exemplar datasets (a-e), corresponding to
882 different efficiency values and mean number of haplotypes; efficiency in percentage and the
883 mean number of haplotypes per dataset are reported in the scatterplot as points coordinates.
884 Regression line is a mere example of the relation between the variables.

885

886 Fig. 4. Impact of the geographic distance between conspecific collection localities on the
887 ABGD and *ad hoc* threshold delimitation efficiency. The boxplot shows the distribution of
888 median values of geographic distance between conspecific collection localities of all analysed
889 datasets. Two datasets representatives for lower and higher geographic distance values (α and
890 γ), which correspond to higher and lower species delimitation efficiency, are reported. (a)
891 *Lochmaea* dataset consisting of four species; the map reports species collection localities. (b)
892 *Plagiosterna* dataset consisting of one species; the map reports the collection localities of the
893 different haplotypes recovered for *Plagiosterna aenea*. Minimum-spanning haplotypes
894 networks show haplotypes relations (connecting lines) and abundance (size of the circles).

895

896 **Supporting information**

897 Table S1. COI sequences analysed in the study. Specimens taxonomy, GenBank accession

898 numbers and collection localities.
899 Table S2. Datasets analysed and explanatory variables.
900 Table S3. Nucleotide distance threshold values for species delimitation *ad hoc* estimated for
901 each dataset.
902 Table S4. Molecular species delimitation analyses results.
903 Table S5. Output of GLMs related to each delimitation method.
904 Table S6. Relation between number of split resulting from mPTP delimitation and number of
905 sequences per species analysed.
906
907 Fig. S1. Result of the test of multiple correlations between the explanatory variables. The
908 diagonal reports the histograms of each variable; the scatterplots below the diagonal report the
909 visual correlation between pairs of variables; the numbers in the squares above the diagonal
910 report the Pearsons's correlation r values for each pair of variables.