

Zurich Open Repository and Archive University of Zurich Main Library Strickhofstrasse 39 CH-8057 Zurich www.zora.uzh.ch

Year: 2015

Structure-dynamic relationship of plant–insect networks along a primary succession gradient on a glacier foreland

Losapio, Gianalberto ; Jordán, Ferenc ; Caccianiga, Marco ; Gobbi, Mauro

Abstract: There is a growing interest in understanding the structure–dynamic relationship of ecological networks. Ecological network changes along primary successions are poorly known: to address such topic, gradient of primary succession on glacier forelands is an ideal model, as sites of different age since deglaciation stand for different ecosystem developmental stages. We aimed to investigate the assembly processes of plant–insect networks and to elucidate its functional implications for ecosystem stability along this time sequence succession. We collected data on the functional role of anthophilous insect groups and performed network analysis to evaluate their relative importance in the structure of plant–insect interaction networks with increasing time since deglaciation along the primary succession of a debriscovered glacier foreland. We sampled anthophilous insects visiting the flowers of two models plant species, Leucanthemopsis alpina and Saxifraga bryoides. Insects were identified and trophic roles were attributed to each taxon (detritivores, parasitoids, phytophagous, pollinators, predators, and opportunists) at five sites representing the primary succession gradient. Plant–insect interactions were visually represented by a bipartite network for each successional stage. For each plant species and insect group, centrality indices were computed quantifying their community importance. For the whole network, centralization and link density were calculated. Pollinators dominated pioneer communities on the debris-covered glacier and in recently deglaciated areas, while parasitoids, predators and opportunists characterised late-succession stages. Plant species centrality varied along the succession. Pollinators showed initially higher but then decreasing centrality, while the centrality of predators and parasitoids increased with time since deglaciation. Along the same gradient link density showed an increasing trend while network centralization tended to decrease. The present study provides new insight into the initial steps of plant– insect network assembly and sheds light on the relationship between structure and dynamic in ecological networks. In particular, during the succession process, more links are formed and plant–anthophilous insect interactions change from a network dominated by pollinators to a functionally more diversified food web. We conclude that applying network theory to the study of primary succession provides a useful framework to investigate the relationship between community structure and ecosystem stability.

DOI: https://doi.org/10.1016/j.ecolmodel.2015.07.014

Posted at the Zurich Open Repository and Archive, University of Zurich ZORA URL: https://doi.org/10.5167/uzh-123318 Journal Article Accepted Version

The following work is licensed under a Creative Commons: Attribution-NonCommercial-NoDerivatives 4.0 International (CC BY-NC-ND 4.0) License.

Originally published at: Losapio, Gianalberto; Jordán, Ferenc; Caccianiga, Marco; Gobbi, Mauro (2015). Structure-dynamic relationship of plant–insect networks along a primary succession gradient on a glacier foreland. Ecological Modelling, 314:73-79. DOI: https://doi.org/10.1016/j.ecolmodel.2015.07.014

investigate the assembly processes of plant-insect networks and to elucidate its functional implications for ecosystem stability along this time sequence succession. We collected data on the functional role of anthophilous insect groups and performed network analysis to evaluate their relative importance in the structure of plant-insect interaction networks with increasing time since 22 23 24 25

deglaciation along the primary succession of a debris-covered glacier foreland. We sampled anthophilous insects visiting the flowers of two models plant species, *Leucanthemopsis alpina* and *Saxifraga bryoides.* Insects were identified and trophic roles were attributed to each *taxon* (detritivores, parasitoids, phytophagous, pollinators, predators, and opportunists) at five sites representing the primary succession gradient. Plant-insect interactions were visually represented by a bipartite network for each successional stage. For each plant species and insect group, centrality indices were computed quantifying their community importance. For the whole network, centralization and link density were calculated. Pollinators dominated pioneer communities in recently deglaciated areas, while parasitoids, predators and opportunists characterised latesuccession stages. Plant species centrality varied along the succession. Pollinators showed initially higher but then decreasing centrality, while the centrality of predators and parasitoids increased with time since deglaciation. Along the same gradient link density showed an increasing trend while network centralization tended to decrease. The present study provides new insight into the initial steps of plant–insect network assembly and sheds light on the relationship between structure and dynamic in ecological networks. In particular, during the succession process, more links are formed and plant-anthophilous insect interactions change from a network dominated by pollinators to a functionally more diversified food web. We conclude that applying network theory to the study of primary succession provides a useful framework to investigate the relationship between community structure and ecosystem stability. 26 27 28 29 30 31 32 33 34 35 36 37 38 39 40 41 42 43 44

45

Keywords: bipartite network, chronosequence, community assembly, ecological network, ecosystem stability 46 47

1. Introduction 49

The study of ecological interaction networks is becoming a key approach for understanding ecological and evolutionary processes (Vázquez et al. 2009) as it provides useful depictions of biodiversity, species interactions, ecosystem structure and functioning (Dunne et al. 2002b). Despite the growing recognition of the importance in analysing the whole-community organization following an ecological network approach (Sridhar et al. 2013), there is still a lack of information on how ecological networks are assembled (Bascompte and Stouffer 2009) and the relationship between ecosystem dynamics and network structure is still poorly understood (Jordán 2009). 50 51 52 53 54 55 56

Recent researches on network ecology provided new insight into structural invariant patterns underlying species interactions. The organization in connected modules (Olesen et al 2007) with a heterogeneous distribution of the number of interactions per species (Dunne et al. 2002a) and asymmetric interaction strength among species (Bascompte 2009) has been related to ecological network robustness (Pocock et al. 2012) and stability (Thébault and Fontaine 2010). Little attention, however, is given to the spatial aspects and to the temporal dimension of ecological networks, despite their relevance for mechanisms of network formation (Bascompte and Stouffer 2009) and for network robustness to species extinction (Pascual and Dunne 2006). 57 58 59 60 61 62 63 64

Ecological succession (i.e. the change of species composition over time) provides temporal and spatial dimensions to analyse the change in the characteristics of populations, communities and ecosystems (Walker & del Moral 2003), and may therefore be suitable to look at the temporal dynamics of ecological networks. Glacier forelands represent such a gradient of primary succession, as sites of different age since deglaciation stand for different ecosystem developmental stages (Matthews 1992). 65 66 67 68 69 70

The use of the chronosequence as a space-for-time substitution (Foster and Tilman 2000) along glacier foreland has provided significant insights into the patterns and mechanisms of plant (Walker et al. 2010) and arthropod (Kaufmann 2001) community assembly. Vegetation cover, plant and 71 72 73

arthropod diversity increase throughout the succession (Hodkinson et al. 2001, Gobbi et al. 2010). Plant community structure changes due to different efficient resource-use among pioneer and latesuccessional species (Caccianiga et al. 2006). In parallel, the turnover of arthropods is influenced by the stabilization of environmental conditions and vegetation structure (Gobbi et al. 2006). However, previous studies have focused on a single trophic level and very little is known about ecological network assembly during primary succession (Albrecht et al. 2010). The only plant-pollinator network examined along such a gradient showed an increase in interaction diversity and indicated an increase in pollinator diet breadth (Albrecht et al. 2010). 74 75 76 77 78 79 80 81

While the majority of ecological network studies examine one static network at a time, we aimed to analyse a network gradient, one of the important perspectives in ecological network analysis. Thus, we applied the network analysis by bipartite network and local to global importance indices analysis to describe the structure of plant-anthophilous insect network and we compared it among different successional stage of a glacier foreland, focusing on insect trophic roles. Thus, we performed network analysis to evaluate the structure of the network and compared it among different successional stages of a glacier foreland. In particular we addressed the following questions: (i) Does the structural importance of plants and insects vary along the successional gradient? (ii) How does the network architecture change along the spatio-temporal gradient? Finally, by integrating structure with dynamic of ecological networks we provide new insight into network assembly and arise empirical models for species coexistence and ecosystem stability. 82 83 84 85 86 87 88 89 90 91 92

93

2. Material and methods 94

95

Data sampling 96

The study was performed along the glacier foreland of Vedretta d'Amola glacier (Central Italian Alps, 46°13'16"N, 10°40'41"E), which is a debris-covered glacier of about 82.1 ha, with two thirds 97 98

covered by stony debris with variable depth, from few centimetres to approximately one meter. The glacier foreland is 1.23km long, covers an altitudinal range from 2425m to 2560m a.s.l., and is characterized by a big moraine system dating back to the Little Ice Age (1500-1850 A.D.). Field observations and various sources including maps, reports, aerial photographs, iconography, and records of length change made over the last 100 years allowed the reconstruction of the glacier tongue position during the largest extent reached: at the end of the Little Ice Age (c. 1850 A.D.), in 1925 A.D., and in 1994 A.D (Fig.1). Following this deglaciation gradient five sampling sites were located to represent the main successional stages: glacier surface (stage 0), 1-20 years (stage I), 21- 90 years (stage II), 91-160 years (stage III), and more than 160 years (stage IV). 99 100 101 102 103 104 105 106 107

We selected the flowering plants *Leucanthemopsis alpina* (L.) Heyw. (Asteraceae) and *Saxifraga bryoides* L. (Saxifragaceae) as model species because they were the only two entomophilous plant species that occurred throughout the whole primary succession gradient. At each successional stage, two 25 m² plots were established and three *L. alpina* tufts and three *S. bryoides* cushions were selected for each species and marked for further use over the course of the study. The number of flowers of each tuft or cushion was recorded in July 2012. Plant-anthophilous insect interactions were observed during the flowering seasons (between the end of July and the end of August) of the summers of 2012 and 2013. All anthophilous insects visiting the flowers were sampled with an entomological aspirator by observing the three plant species units together during three periods of 40 minutes a day at 11am, 1pm and 3pm (90 samples in total per year). Anthophilous insects were identified at species level if possible, otherwise at genus or family level. Insects were classified into six ecological roles based on trophic habits (Fath and Killian 2007; Gobbi and Latella 2011) by literature survey (e.g. Mellini 1997; Gregor et al*.* 2002; Oosterbroek 2006; Jedlička et al. 2009): detritivores, parasitoids, phytophagous, pollinators, predators, and opportunists. 108 109 110 111 112 113 114 115 116 117 118 119 120 121

122

Data analysis 123

We quantified structural changes (Dunne et al. 2002a) of the plant-insect network at both local (node positions) and global (network architecture) levels along the primary succession gradient. The patterns and frequency of plant-insect interactions were represented and visualized by bipartite networks (Jordano 1987, Memmott 1999, Dormann et al. 2009). A bipartite network consists of two sets of nodes (i.e. plant species and insect functional groups) linked by a set of edges in such a way that each edge links two species belonging to different node sets. In the adjacency matrix plants were represented in the rows (= lower level in the network) and insect functional groups were represented in the columns (= upper level in the network). Link weights showed the number of insect individuals that visited the corresponding plant species. In the resulted bipartite network, the size of rectangles representing plants and insects was proportional to the relative number of visits received and made within each successional stage, respectively (Dormann et al. 2009). 124 125 126 127 128 129 130 131 132 133 134

In order to quantify the change in network structure along the successional gradient, we computed local indices describing nodes and global indicators of network architecture. Information about the changes in topological properties along the succession provides useful information to understand the relative importance of various functional groups and may shine a light on the dynamical consequence of network assembly. 135 136 137 138 139

140

Local (node level) indices 141

142

Weighted degree (wDi) 143

In a directed and weighted network, it is the sum of weights of the links connected directly to a node. This is the most local measure and often provides a fast and simple evaluation (Jordán 2009). 144 145 146

Weighted topological importance index (WI_iⁿ) 147

We assume a network with undirected links where trophic effects can spread in many directions 148

without bias. Indirect effects can spread in both bottom-up and top-down directions and, as a result, horizontally, too (i.e. from plant to plant and from insect to insect). We use WI_i^n as the topological importance of species *i* for plant-insect network with weighted links when effects "up to" *n* steps are considered as 149 150 151 152

153

154
$$
WI_i^n = \frac{\sum_{m=1}^n \sigma_{m,i}}{n} = \frac{\sum_{m=1}^n \sum_{j=1}^N a_{m,j}}{n}
$$

155

which is the sum of effects originated from species *i* up tp *n* steps averaged over by the maximum number of steps considered (*n*). By this index, it is possible to quantify the internal interactions structure of the network (Jordán 2009). 156 157 158

Both *wD* and WIⁿ were calculated according to Valentini and Jordán (2010). These structural importance indices assume that well-connected nodes are more important in the network in a structural and possibly also in a dynamical sense (Jordán et al. 2007, Jordán et al. 2008). As pollinators, predators and parasitoids were the consistently most abundant groups, the indices were calculated only for these. 159 160 161 162 163

164

Global (network level) indices 165

166

Density 167

The density of a network is the number of links divided by the maximal number of potential links. In the case of weighted networks, Wasserman and Faust (1994) suggests to use the sum of link weights instead of the number of links. Also, in the case of bipartite networks, the denominator should be *NM*, where *N* and *M* are the number of nodes in the first (plants) and in the second set (insects) of nodes, respectively. Thus, we used the following formula for our weighted and bipartite 168 169 170 171 172

networks:

$$
175 \quad d = \frac{\sum_{i=1}^{L} W_i}{N \times M}
$$

where *L* is the number of links, W_i is the weight on the ith link, *N* is the number of plant species and *M* is the number of insect functional groups.

Centralization (*NCI^D*)

We calculated the degree–based network centrality index (*NCI^D*), where degree (*D*) is the number of neighbours of a graph node (Wasserman and Faust 1994). Note that network centrality, expressed in percentages, is maximal (100%) if a central node is directly connected to all other nodes and there is no other link in the network (i.e. a perfect star-shape), and it is minimal (0%) if the positions of all nodes are topologically equal (i.e. a lattice).

3. Results

A total of 911 insects specimens belonging to 40 families and 6 functional groups were sampled. The *L. alpina* anthophilous insect community was dominated by pollinators (84%), predators (9%) and opportunists (5%); phytophagous and detritivores both accounted for 1%. The *S. bryoides* insect community was composed of pollinators (59%), followed by parasitoids (15%), opportunists (14%), predators (8%); phytophagous and detritivores made up 3% and 1%, respectively.

Plant-anthophilous insect bipartite network

Pollinators dominated the insect community of the debris-covered glacier (stage 0) as well as of

stages I and II (Fig. 2). In stage III, the insect community was more diverse and structured, as opportunists, parasitoids and predators increased particularly on *S. bryoides* flowers. On the oldest terrain in stage IV, the insect community on *L. alpina* was still mainly represented by pollinators, while *S. bryoides* showed a more functionally diverse insect community (Fig. 2). In this latesuccessional stage, *S. bryoides* interacted mainly with parasitoids, while opportunists and predators increased their frequency and pollinators became relatively less abundant (Fig. 2). 197 198 199 200 201 202

203

Network indices 204

L. alpina had higher *wD* in stage 0, and among all other stages no evident trend coherent with the successional gradient emerged (Fig. 3a). The *WI²* of *L. alpina* generally increased from stage 0 to IV, with a major change between stage II and stage III (Fig. 3b). The *wD* of *S. bryoides* increased from stage 0 to II, where it reached the maximum and later decreased to stage IV (Fig. 3c). The *WI²* of *S. bryoides* increased from stage 0 to III, then weakly decreased in stage IV (Fig. 3d). 205 206 207 208 209

The *wD* of pollinators tended to decrease along the primary succession gradient towards earlier deglaciation (Fig. 4a), whereas WI² of pollinators increased with later successional stages (Fig. 4b). The *wD* and *WI²* of predators increased from stage 0 to IV (Fig. 4c), with a major change between stage III and IV (Fig. 4d). The *wD* and *WI²* of parasitoids remained low during the early and mid successional stages, then increased mainly from stage III to IV (Fig. 4e-f). Pollinators had, on average, higher *wD* and *WI²* compared to predators and parasitoids. 210 211 212 213 214 215

Regarding the network-level properties, link density (*d*) increased (Fig. 5a) while network centralisation (*NCI^D*) decreased with increasing successional stage (Fig. 5b). This means that the various species and functional groups in the community developed more and more interactions among themselves and became more and more connected, while their relative importance became less heterogeneous with more equally distributed importance among them. 216 217 218 219 220

4. Discussion 222

223

Our study represents one of the first applications of bipartite network centrality analysis to the study of ecological succession. The key finding of our work is that, during succession, the importance of different insect ecological groups and the global network structure changed. Plant-anthophilous insect interactions changed from a network dominated by pollinators to a functionally more diversified food web, where the density of interactions increased while the network became decentralized and homogeneous. We hypothesize that such process may increase the robustness of the network against the local extinction of species, providing new insight into the relationship between structure and dynamic in ecological networks. 224 225 226 227 228 229 230 231

232

Plant-anthophilous insect bipartite network 233

L. alpina interacted mainly with pollinators, which were the dominant insect group along the entire successional gradient studied here. In contrast, *S. bryoides* interacted with a wider ecological spectrum where pollinators were not the dominant group along the whole successional gradient. Difference in the ecological role of anthophilous insects between plant species may be due to the presence of floral nectaries in *S. bryoides* but not in *L. alpina*, suggesting that more insect ecological roles may benefit from this sugar resource. In the *S. bryoides* community pollinators remained the dominant group during the early and mid successional stages, whereas in the late successional stage parasitoids, predators and opportunists strongly increased and pollinators decreased. 234 235 236 237 238 239 240 241 242

Albeit our study does not allow the inference of direct trophic relationship among insects, it may be possible that the relative decrease of pollinators is linked to the increase in predators (Raso et al. 2014) and parasitoids with the succession proceeding. Indeed, a mature ecosystem, found at the late successional stages showing high plant productivity and community diversity (Gobbi et al. 2010), 243 244 245 246

may boost more parasitoids and predators that will interact with pollinators. By meaning of bipartite networks dynamically linked we showed for the first time how different insect ecological groups are assembled and differentially interact with plants along an ecological succession gradient. 247 248 249

250

Local (node) structural changes 251

Weighted centrality indices of *L. alpina* showed contrasting trends. Although no clear variation emerged in wD , the increase in WI^2 indicates that, even if the number of direct links may not vary, indirect effects became more important with later successional stages. Highest values of the two topological indices at the successional stages II and III for *S. bryoides* indicated a more important structural position (rich interaction structure) at the intermediate successional stages. This corresponds with the abundance distribution of this species along the successional gradient, suggesting a relationship between the species environmental requirement (i.e. the species niche) and its importance in structuring the network. 252 253 254 255 256 257 258 259

With succession proceeding, pollinator *wD* decreased and that of predators and parasitoids increased: these opposite trends indicate a change in positional importance from a local to a mesoscale view (Jordán 2009). Due to different roles of species in the network (Jordán 2009), changes in their relative importance along the successional gradient causes structural changes in the network architecture. This local change may influence the global dynamic of plant-anthophilous insect interaction network, which is varying from a pollination-driven system to a more complex ecological network. In other words, the plant-anthophilous insect interactions change along the successional gradient from predominantly mutualistic interactions to a diverse set of interactions including mutualism, parasitism and predation. 260 261 262 263 264 265 266 267 268

Indirect effects play an important role in governing ecosystem dynamics (Jordán et al. 2008). The consistent increase in $WI²$ indices for all insect ecological roles makes them more functionally important along the primary succession gradient. The importance of nodes is therefore likely to be 269 270 271

determined by the interaction between insect functional role type and community features along the primary succession gradient, the later mainly represented by the increase in plant species diversity and ecosystem productivity (Gobbi et al. 2010). 272 273 274

Thus, the use of centrality indices highlights the link in the assembly process of plant and related anthophilous insect communities and the functioning of this interaction system. Our study showed that the local importance of insect ecological roles changed within as well as among the groups along the succession gradient, suggesting a role in the dynamic of global network properties. 275 276 277 278

279

Global (network) structural changes 280

Our results show that with succession proceeding more and more links among plants and insects are formed. This may be due to the effect of time: directly, by increasing the interaction probability among the species pool, and indirectly, as a consequence of a more structured and diversified network that lastly enhances trait matching and complementarity (Vázquez et al. 2009) between plants and anthophilous insects. Furthermore, link density is affected by network dimension (Dormann et al. 2009), the latter being a consequence, in our study system, of the increase in insect abundance and plant community maturity. The increase in link density with proceeding succession is in accordance with Albrecht et al. (2010) who found an increase in the unweighted number of links per species along the Morteratsch glacier foreland (Switzerland). As link density is an indicator of complexity (May 1973), the increase of network complexity along the primary succession gradient confers stability and robustness against potential local species losses (Montoya et al. 2006) because highly connected networks will tend to be more robust to stochastic removal of nodes (Dunne et al. 2002b). Conversely, a very densely connected network is more vulnerable to species invasion as a perturbation is more likely to spread rapidly (Scheffer et al. 2012). 281 282 283 284 285 286 287 288 289 290 291 292 293 294

Patchy pioneer communities in early successional stages are highly centralised, with few nodes playing a key role while most nodes occupy peripheral positions. Throughout the primary 295 296

succession the network became de-centralised, as all nodes are connected with a similar number of links. This highlights a gradient towards a more homogeneous structural organisation: the network architecture shifts from a star-like network, where few central nodes have many connections, to a lattice-like network with no central nodes but a balanced number of edges per node (Wassermann & Faust 1994). The network de-centralisation may emerge from the decrease in direct structural importance of pollinators and the parallel increase in importance of parasitoids and predators towards later successional stages. As reported in others studies on pollination and mutualistic networks (Olesen et al. 2007, Bascompte and Stouffer 2009, Pocock et al. 2012), low interaction richness makes a network less cohesive, thus making the network more vulnerable to the removal of nodes. Thus, we could hypothesise that the observed network de-centralisation may increase the local extinction risk of species, especially in a global warming context causing glacier retreat. 297 298 299 300 301 302 303 304 305 306 307

In summary, considering simultaneously all the global network properties, the network shows increasing link density and increasing de-centralisation during primary succession. These two assembly processes appear to confer two contrasting properties: the former may lead to network robustness against local species extinctions whereas the latter may confer the opposite feature. Indeed, as some simulation studies suggested (Bascompte & Stouffer 2009, Pocock et al. 2012), a cohesive organization make the network more robust and stable. In our study system, probably the outcome is a balance between these two functions (i.e. increase in link density and decrease in centralisation) and the importance of species getting extinct (i.e. the species ecological group and their structural role within the network). More studies are require to understand the dynamic consequence of network structural evolution along primary succession gradients; however, our application of local and global centrality indices has proved to give important insights into these processes and their consequences. 308 309 310 311 312 313 314 315 316 317 318 319

320

Conclusion 321

We shed light on dynamically linked ecological networks in a natural time-sequence succession. Although in our study we analysed the anthophilous insect network of two ubiquitous plant species, we believe that the emerged trends may also be extended to a wider community-level pattern. New insight was provided into the initial steps of plant–insect network assembly and new light was shed on the relationship between structure and dynamic in network ecology. In conclusion, our framework highlighted the relevance of combining ecological trait and network theory to increase the link between community structure, network assembly and ecosystem functioning. Indeed, by applying network theory we will move away from pairwise comparisons and start searching for the existence of network-wide patterns of species dependence (Ings et al. 2009). We emphasize that, in the present context of climate and environmental changes, network analysis of primary succession gradients by meaning of bipartite networks and centrality indices may provide a useful framework also for the management of endangered habitat and species. 322 323 324 325 326 327 328 329 330 331 332 333

334

Acknowledgements 335

We thank the Adamello-Brenta Natural Park for providing sampling permission and subsidized accommodation. We are grateful to Giuseppe Marano and Patrizia Boracchi for their help in R programming, to Roberto Seppi for providing the glacier foreland chronology, to Chiara Compostella for providing the GIS cartography of the study area, and to the following taxonomic specialists for help with the species identification: Daniele Avesani (Diptera) and Daniele Sommaggio (Diptera), Adriano Zanetti (Coleoptera), Maurizio Pavesi (Hymenoptera). We thank Christian Schöb, the editor (Brian D. Fath) and two anonymous reviewers for fruitful comments to the first draft of the manuscript. 336 337 338 339 340 341 342 343

344

References 345

Abrams, P. A., Menge, B. A., Mittelbach, G. G., Spiller. D. A., Yodzis, P. 1996. The role of indirect 346

- effects in food webs. In: Polis, G. A. and Winemiller, K. O. (eds.) Food webs: integration of 347
- patterns and dynamics. Chapman and Hall, London. pp. 371-395. 348
- Albrecht, M., Riesen, M., Schmid, B. 2010. Plant-pollinator network assembly along the chronosequence of a glacier foreland. *Oikos*, **119**, 1610-1624. 349 350
- Bascompte, J. 2009. Disentangling the web of life. *Science*, **325**, 416-419. 351
- Bascompte, J. & Stouffer, D.B., 2009. The assembly and disassembly of ecological networks. *Phil. Trans. R. Soc. B,* **364**, 1781-1787. 352 353
- Caccianiga, M., Luzzaro, A., Pierce, S., Ceriani, R.M., Cerabolini, B. 2006. The functional basis of a primary succession resolved by CSR classification. *Oikos*, **112**, 10-20. 354 355
- Fath, B.D., Scharler, U.M., Ulanowicz, R.E., Hannon, B. 2007. Ecological network analysis: 356
- network construction. *Ecological Modelling*, **208**, 49-55. 357
- Fath, B.D., Killian, M.C. 2007. The relevance of ecological pyramids in community assemblages. *Ecological Modelling*, **208**, 286-294. 358 359
- Dormann, C.F., Fründ, J., Blüthgen, N., Gruber B. 2009. Indices, Graphs, and Null Models: Analyzing Bipartite Ecological Networks. *The Open Ecology Journal*, 2009, **2**, 7-24. 360 361
- Dunne, J.A., Williams, R.J., Martinez, N.D. 2002a. Food-web structure and network theory: the role of connectance and size. *PNAS*, **99**, 12917-12922. 362 363
- Dunne, J.A., Williams, R.J. and Martinez, N.D. 2002b. Network structure and biodiversity loss in food webs: robustness increases with connectance. *Ecology Letters*, **5**, 558-567. 364 365
- Foster, B.L., Tilman, D. 2000. Dynamic and static view of succession: testing the descriptive power 366
- of the chronosequence approach. *Plant Ecology*, **146**, 1-10. 367
- Gobbi, M., De Bernardi, F., Pelfini, M., Rossaro, B., Brandmayr, P. 2006. Epigean arthropod 368
- succession along a 154 year glacier foreland chronosequence in the Forni Valley (Central Italian 369
- Alps). *Artic, Antarctic, and Alpine Research*, **38**, 357-362. 370
- Gobbi, M., Caccianiga, M., Cerabolini, B., Luzzaro, A., De Bernardi, F. 2010. Plant adaptive 371
- response during primary succession are associated with functional adaptations in ground beetles on deglaciated terrain. *Community Ecology*, **11**, 223-231. 372 373
- Gobbi, M., Latella, L. 2011. *La fauna dei prati -1: tassonomia, ecologia e metodi di studio dei principali gruppi di invertebrati terrestri italiani.* Quaderni del Museo delle Scienze, Trento. 374 375
- Gregor, F., Rozkošny, R., Barták, M., Vaňhara, J. 2002. The Muscidae (Diptera) of Central Europe. 376
- *Folia Facultatis Scientiarum Naturalium Universitatis Masarykianae Brunensis, Biologia*.**107**,1- 280. 377 378
- Hodkinson, I.D., Coulson, S.J., Harrison, J., Webb, N.R. 2001. What a wonderful web they weave spiders, nutrient capture and early ecosystem development in the high Arctic – some counterintuitive ideas on community assembly. *Oikos*, **95**, 349-352. 379 380 381
- Ings, T.C., Montoya, J.M., Bascompte, J., Blüthgen, N., Brown, L., Dormann, C.F., Edwards, F., 382
- Figueroa, D., Jacob, U., Jones, J.I., Lauridsen, R.B., Ledger, M.E., Lewis, H.M., Olesen, J.M., 383
- van Veen, F.J.F., Warren, P.H., Woodward, G. 2009. Ecological networks beyond food webs. *Journal of Animal Ecology*, **78**, 253-269. 384 385
- Jedlička, L., Kúdela, M., Stloukalová, V. (2009) *Cheklist of Diptera of the Czech Republic and Slovakia.* Electronic version 2. http://zoology.fns.uniba.sk/diptera2009 386 387
- Jordán, F., Benedek, Z., Podani, J. 2007. Quantifying positional importance in food webs: A comparison of centrality indices. *Ecological Modelling*, **205**, 270-275. 388 389
- Jordán, F., Okey, T.A., Bauer, B., Libralato, S. 2008. Identifying important species: Linking structure and function in ecological networks. *Ecological Modelling*, **216**, 75-80. 390 391
- Jordán, F. 2009. Keystone species and food web. *Phil. Trans. R. Soc. B*, **364**, 1733-1741. 392
- Jordano, P. 1987. Patterns of mutualistic interactions in pollination and seed dispersal: connectance, 393
- dependences, asymmetries and coevolution. *American Naturalist,* **129**, 657-677. 394
- Kaufmann, R. 2001. Invertebrate succession on an Alpine glacier foreland. *Ecology*, **82**, 2261-2278. 395
- Matthews, J.A. 1992. *The Ecology of Recently-deglaciated Terrain: A Geoecological Approach to* 396
- *Glacier Forelands and Primary Succession.* Cambridge University Press, Cambridge. 397
- May, R.M. 1973. *Stability and Complexity in Model Ecosystems*. Princeton University Press, Princeton. 398 399
- Mellini, E. 1997. From predation to parassitoidism in the Diptera order. *Bollettino dell'Istituto di Entomologia "Guido Grandi" della Università degli Studi di Bologna*, **51**, 91-159. 400 401
- Memmott, J. 1999. The structure of a plant-pollinator food web. *Ecology Letters*, **2**, 276-280. 402
- Menge, B. A. 1995. Indirect effects in marine rocky intertidal interaction webs: patterns and importance. *Ecological Monographs*, **65**, 21-74. 403 404
- Montoya, J.M., Pimm, S.L., Solé, R.V. 2006. Ecological networks and their fragility. *Nature*, **442**, 259-264. 405 406
- Pascual, M., Dunne, J.A. 2006. *Ecological Networks: Linking Structure to Dynamics in Food Webs.* 407
- Santa Fe Institute, Studies in the Sciences of Complexity. Oxford University Press, Oxford. 408
- Olesen, J., Bascompte, J., Dupont, Y., Jordano, P. 2007. The modularity of pollination network. *PNAS*, **104**, 19891-19896. 409 410
- Oosterbroek, P. 2006. *The European families of Diptera: Identification, diagnosis, biology.* KNNV Publishing, Utrecht. 411 412
- Pocock, M.J., Evans, D.M., Memmott, J. 2012. The robustness and Restoration of a Network of Ecological Networks. *Science*, **335**, 973-977. 413 414
- Raso, L., Sint, D., Mayer, R., Plangg, S., Recheis, R., Kaufmann, R., Traugott M. 2014. Intraguild predation in pioneer predator communities of Alpine glacier forelands. *Molecular Ecology*, **23**, 3744-3754. 415 416 417
- Scheffer M., Carpenter, S., Lenton, T.M., Bascompte, J., Brock, W., Dakos, V., van de Koppel, J., 418
- van de Leemput, I., Levin, S.A., van Nes, E.H., Pascual, M., Vandermeer, J. 2012. Anticipating 419
- Critical Transitions. *Science*, **338**, 344-348. 420
- Sridhar, H., Jordán, F., Shanker, K. 2013. Species importance in a heterospecific foraging 421
- 17 17
- association network. *Oikos*, **122**, 1325-1334. 422
- Thébault, E., Fontaine, C. 2010. Stability of ecological communities and the architecture of mutualistic and trophic networks. *Science*, **329**, 853–856. 423 424
- Valentini, R., Jordán, F. 2010. CoSBiLab Graph: the network analysis module of CoSBiLab. *Environmental Modelling and Software*, **25**, 886-888. 425 426
- Vázquez, D., Blüthgen, N., Cagnolo, L., Chacoff, N. 2009. Uniting pattern and process in plant-427
- animal mutualistic networks: a review. *Annals of Botany*, **103**, 1445-1457. 428
- Walker, L.R., del Moral, R. 2003. *Primary succession and Ecosystem Rehabilitation.* Cambridge, Cambridge University Press. 429 430
- Walker, L.R., Wardle, D.A., Bardgett, R.D., Clarkson, B.D. 2010. The use of chronosequence in 431
- studies of ecological succession and soil development. *Journal of Ecology*, **98**, 725-736. 432
- Wasserman, S., Faust, K. 1994. *Social Network Analysis: Methods and Applications.* Cambridge: Cambridge University Press. 433 434
- Wootton, J. T. 1994. The nature and consequences of indirect effects in ecological communities. 435
- *Ann. Rev. Ecol. Syst.*, **25**, 443-466. 436